

GOLETA SANITARY DISTRICT NPDES MONITORING PROGRAM 2014 ANNUAL REPORT

Submitted: April 1, 2015

GOLETA SANITARY DISTRICT

NPDES Monitoring and Reporting Program

2014 Annual Report

Quarterly and Annual Receiving Water Monitoring
Conducted by
Aquatic Bioassay and Consulting Laboratories, Inc.
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Submitted April, 2015

I certify under penalty of law that this document and all attachments were prepared under my direction or supervision in accordance with a system designed to assure that qualified personnel properly gather and evaluate the information submitted. Based on my inquiry of the person or persons who manage the system, or those persons directly responsible for gathering the information, the information submitted is, to the best of my knowledge and belief, true, accurate, and complete. I am aware that there are significant penalties for submitting false information, including the possibility of fine and imprisonment for knowing violations.

Robert Hidalgo

Plant Superintendent Goleta Sanitary District

Date: <u>Opril 1st</u> 2015

AQUATIC BIOASSAY OCEANOGRAPHIC TEAM

MARINE BIOLOGISTS

S. Johnson, Senior Scientist
Jim Mann, Senior Biologist
Karin Patrick, Senior Biologist
Wendy Willis, Biologist
Brianna Jones, Biologist

MARINE CHEMISTS

R. Gossett (PHYSIS Laboratories)

TAXONOMISTS

A. Phillips

BOAT CREW

J. Gelsinger, Captain

CHAPTER 1

INTRODUCTION

The Goleta Sanitary District (GSD) treatment plant operates under Clean Water Act Section 301(h) which waives secondary treatment requirements. On November 19, 2004 the California Regional Water Quality Control Board, Central Coast Region (RWQCB), adopted Waste Discharge Requirements (WDR) Order R3-2004-0129 and the United States Environmental Protection Agency (EPA), Region IX issued NPDES permit CA 0048150 to the Goleta Sanitary District (GSD). A settlement agreement was made a part of the NPDES 301(h) waiver permit issued in 2004. The settlement agreement required GSD to upgrade its wasetwater treatment plant to full secondary treatment by November of 2014.

As required by waste discharge requirements GSD submitted an NPDES permit renewal application to the RWQCB and the EPA in May 2009. At the time of the application submittal, the District was five years into the ten year conversion schedule described in the settlement agreement of 2004. Both the RWQCB and the EPA agreed to renew the 301(h) waiver permit for another five years while GSD continued to make progress to upgrade its treatment facility. The treatment plant is operating under WDR Order No. R3-2010-0012 and NPDES Permit No. CA0048160 which became effective September 2010.

Although GSD continues to operate the wastewater treatment facility under the 301(h) waiver provision of the Clean Water Act the final full secondary tie-in of the newly built structures to the existing plant was completed on May 15 to 16, 2013.

This annual report will discuss the upgade treatment processes, uder the following section: WASTEWATER TREATMENT PROCESS. January through December 2014 the plant was operating utilizing the full secondary process.

As a condition of the NPDES permit, GSD is required to conduct an extensive monitoring and reporting program to assess compliance with limitations established by the California Ocean Plan and the federal Clean Water Act. Under conditions set forth in the permit, GSD must monitor the influent, effluent, biosolids (sludge), the outfall and diffuser, receiving water, bottom sediment, and biology to demonstrate that the discharge of wastewater is not causing adverse impacts on the ocean environment.

The Goleta wastewater treatment plant (WWTP) is located in an unincorporated coastal area of Santa Barbara County, California. Treated wastewater is discharged to the Pacific Ocean approximately one mile offshore of Goleta Beach County Park via a south-trending ocean outfall. The outfall lies within and extends outside of a small embayment formed by Goleta Point directly to the west.

The Goleta WWTP treats wastewater from the service areas of the Goleta Sanitary District (GSD), the Goleta West Sanitary District, the University of California at Santa Barbara, the Santa Barbara Municipal Airport, and certain Santa Barbara County facilities. Existing

agreements among the agencies establish GSD as the owner of the joint wastewater treatment facilities and assign the responsibility of operation and maintenance of the facilities to GSD. However, each agency "owns" an "indeterminate, perpetual and exclusive capacity right" in the facilities and an "easement right of flow through" the facilities.

WASTEWATER TREATMENT PROCESS

The following discussion focuses on the principal features of GSD's full secondary process of wastewater and sludge treatment. The performance capacities and characteristics of the treatment plant are detailed in Chapter 2.

Treatment Plant Facilities

The Goleta Sanitary District Wastewater Treatment Plant is located at One William Moffett Place, in an unincorporated area of Santa Barbara County, CA. The plant site is approximately 10 miles west of the City of Santa Barbara, near the Pacific Coast. A regional view of the study area is shown in Figure 1-1.

On average, over the past 10 years, 2005 to 2014, the plant has discharged about 3.9 million gallons per day (MGD) of treated effluent to the open coastal waters of the Santa Barbara Channel via an ocean outfall. The treatment plant is currently discharging municipal wastewater in accordance with NPDES permit CA 0048160. The treatment plant's discharge meets the state water quality standards as set forth in the Water Quality Control Plan for Ocean Waters of California (California Ocean plan) and the federal Clean Water Act.

Facilities Description

The Goleta wastewater treatment plant underwent it's first substantial upgrade completed in June 1988. The upgraded plant was designed to assure compliance with monthly 30-day average discharge limitations of 63 mg/L for suspended solids and 98 mg/L for BOD under an average dry weather flow 9.0 MGD. The facility has utilized a split-stream process of physical and biological treatment until December of 2013. The current biological treatment is provided by two trickling filters and an aeration basin to achieve full secondary treatment. The following sections describe the treatment process.

Collection System

Over 190 miles of pipelines collect wastewater that flows almost entirely by gravity to pump stations located in each agency's service area. These stations pump the flow to the treatment facility.

Pump Station and Headworks

Influent from the collection system of each agency is pumped to the treatment plant headworks where raw wastewater flows through a bar screen which removes large debris. Influent is then routed to aerated grit tanks where sand and grit are allowed to settle out and pumped to two cyclone separators, which dumps grit into a hopper for dewatering. This debris and grit is then transported via truck to a local landfill. Air collected from the

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influent pump stations and headworks is scrubbed in two odor reduction towers equipped with activated carbon.

The upgrade of 2013 also included upgrading structures that had reached the end of their useful life. The headworks upgrade included the installation of two new bar screens with smaller screen spacing, ¼ inch in order to better remove more inorganic materials and the installation of two new screening washer/compactor units. The odor reduction tower at the headworks was removed and replaced with a biological odor reduction tower.

Primary Sedimentation

Wastewater then flows into one of three circular primary sedimentation basins (primary clarifiers) where solids settling to the bottom and floatable materials rising to the surface are mechanically collected and pumped to digesters.

No new structures were added to the primary treatment stage as part of the upgrade. However as part of the renovations performed under the treatment plant upgrade all three primary clarifiers were drained and inspected. Renovations included replacement of the boom sweeps, removal and replacement of the sweep motors, the catwalks were sand blasted and both the catwalks and troughs were painted. Additionally, the concrete around the effluent trough was deteriorating and this area was patched.

Secondary Treatment

Secondary treatment involves three treatment elements: the biofilter, an aeration basin, and secondary sedimentation tanks. In the biofilter, primary effluent trickles over plastic media where bacteria feed on organic wastes, thus removing these wastes from the water. Effluent from the trickling filter flows to an aeration basin where air is injected and the effluent is mixed with recirculated sludge from the secondary sedimentation basins. The resulting biological action coagulates these fine particles and the organic solids settle out as sludge in two secondary sedimentation tanks. The waste activated sludge (WAS) is pumped to two new mechanical thickeners and then is pumped to the three anaerobic digesters. A portion of the secondary process flow can be diverted to the reclamation facilities for tertiary treatment with gravity filters.

The upgrade included the construction of a new biofilter identical to the existing, demolition of the solids contact channel and construction of a three train aeration basin with structures stubbed out for the construction of a fourth train in the future, if needed. New construction also included a new blower building, two new secondary clarifiers and construction of various supporting structures, such as pumping stations, interstage pump, return activated sludge (RAS) station, etc.

Chlorine Contact Channel

The secondary effluent flows to the head of the chlorine contact channel where sodium hypochlorite is injected to kill bacteria in the effluent. Prior to discharge into the ocean, sodium bisulfite is added for dechlorination, thus completing the disinfection process.

Sludge Treatment and Biosolids Disposal

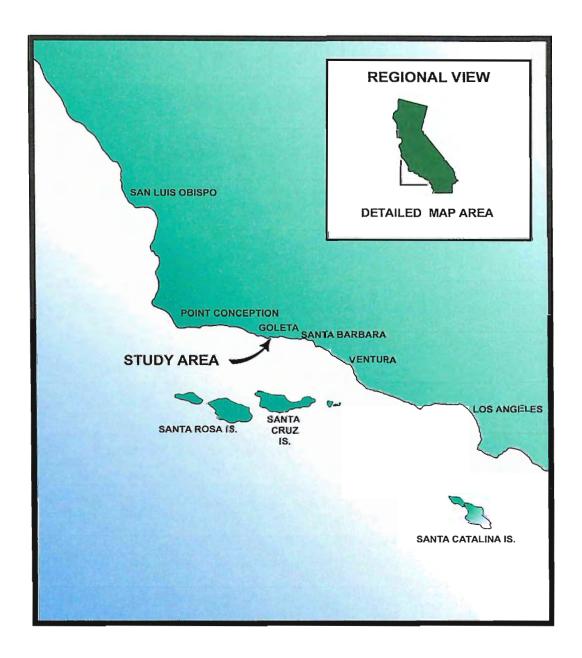
Settleable solids and floatable materials from the primary clarifiers are treated in three heated anaerobic sludge digesters for at least 15 days. Anaerobic digestion decomposes organic material and produces digester gas composed primarily of methane. This digester gas fuels boilers used to heat sludge in the digesters. Sludge from the digesters then flows to one of two stabilization basins where it settles and bacteria can continue the organic decomposition. Stabilized sludge is dredged from the bottom of these basins and is dewatered by new two screw presses. The digested supernatant from the three anaerobic digesters can also be diverted from the stabilization basins directly to the two new screw presses for dewatering.

Because of the extensive construction taking place to upgrade the secondary portion of the treatment facilities, the sludge that would be air dried in the sludge drying beds and converted into Class A biosolids, for use by the local community has been temporarily discontinued. The belt pressed and then the screw pressed biosolids, identified as Class B, were transported under a new three year contract signed in 2012 by Western Express, Inc. to the Holloway Solid Waste Facility in Lost Hills, CA. The biosolids are used to reclaim the depleted gypsum mine pits. Goleta Sanitary District was notified through Western Express on February 21, 2014, that Holloway Solid Waste Facility would not continue to accept biosolids from GSD due to the high moisture content. As a result, GSD entered into an agreement with Liberty Composting Inc. on March 28, 2014, to accept the biosolids generated by GSD. The agreement with Western Express was amended on April 4, 2014, to transport GSD's biosolids to Liberty Composting Inc. located at 12421 Holloway Road, Lost Hills, CA 93249. Copies of the agreement with Liberty Composting and the agreement amendment with Western Express are available upon request. The administrative office for Western Express Inc., is located at 1533 E. Shields Ave., Suite F, Fresno, CA 93607.

A complete biosolids report describing the treatment and disposal process is prepared each year and submitted to the EPA. The deadline for submittal of this report is Februay 19th of each year.

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Figure 1-1. Regional View of the Goleta Valley



Reclamation Facilities

On September 13, 1991, the California Regional Water Quality Control Board, Central Coast Region approved Order No. 91-03 that permits the Goleta Sanitary District to produce up to 3.0 MGD of reclaimed water. The reclaimed water produced at the Goleta Sanitary District is distributed by the Goleta Water District for use within their service area. Reclaimed water is used for landscape irrigation and for incidental uses including construction dust control and compaction, and to flush toilets within several buildings within Goleta. The Goleta Water District is regulated by separate water reclamation requirements.

Secondary effluent enters the reclamation facilities where a flash mixer disperses aluminum sulfate (alum) and polymer into the water. The flocculated suspension is then filtered through a bed of anthracite coal where the floc is removed. The filtered water then flows to a chlorine contact tank where sodium hypochlorite is added for disinfection. The highly chlorinated treated water then flows to a 3 million-gallon underground storage tank where it is stored until needed. Reclaimed water is distributed throughout the Goleta Valley by a distribution system operated and maintained by the Goleta Water District.

An annual report describing the reclamation treatment process, operational parameters, water quality, and production rates is prepared and sumbitted to the RWQCB by January 31st.

No changes were made to the reclamation facilities as part of the 2013 upgrade.

Ocean Outfall

The treated secondary effluent is discharged to the ocean through an outfall pipe that extends 5800 feet offshore and terminates at a depth of approximately 92 feet below Mean Lower Low Water (MLLW) level. At the pipe terminus, a multi-port diffuser with 36, four inch diameter ports mixes one part of effluent with approximately 122 parts of seawater (Tetra Tech, Inc. 1993) to achieve a high initial wastewater dilution.

No changes were made to the outfall as part of the 2013 upgrade.

Staff

Mr. Kamil Azoury, P.E., serves as GSD's General Manager and District Engineer. The General Manager is responsible for overall operation and performance of the treatment plant. Mr. Steve Wagner, P.E., joined the GSD team in March 2014 as the Assistant General Manager / Assistant District Engineer.

Ten state certified treatment plant operators operate the wastewater treatment plant under the direction of Mr. Robert Hidalgo, the District Plant Superintendent. Mr. Hidalgo also supervises the treatment plant's industrial waste staff. Mr. Chuck Smolnikar, supervises the maintenance staff and the laboratory is under the direction of Ms. Lena Cox, the Laboratory Supervisor. The grade and certification number of operations, maintenance, and laboratory personnel employed during the 2014 operational year are shown in Table 1-1.

Table 1-1. Goleta Sanitary District Operation Staff, 2014

Staff	Grade	California Certification No.
Oper	ators	
Robert Hidalgo	IV	6905
Todd Frederick	IV	27633
Stephen Conklin	III	7065
Ricardo Lopez	Ш	10756
Francisco M. Lemus	III	10893
John Crisman	III	28857
Pete Regis	Hi	28277
Jes Hulbert	I	28266
Morgan Lee	ı	28400
River Ferrara	ı	28488
Lab Tech	nologis	st
Lena Cox	IV	90334003
Ray Giordano	111	70733003
Robert Hidalgo	I	741
Teresa Kistner	I	99076111
Todd Frederick	i	60731013
Maintenance	Techno	logist
Carl Easter	H	110662004
John Corral	· ·	770
Robert Hidalgo	i	1087
Mark Bumgarner	I	1308210330
Electrical / Instrumentation		
Charles Smolnikar	11	60172004

Monitoring and Reporting Program

The Goleta Sanitary District monitoring and reporting program was conducted in accordance with the requirements of the NPDES permit CA0048160. The objectives of the monitoring program and this report are to:

- Document short- and long-term effects of discharge on receiving waters, sediment, biota, and beneficial uses of the receiving waters.
- Determine compliance with NPDES permit terms and conditions.
- Document training and certification of wastewater treatment facility operators.
- Assess treatment plant performance and the effectiveness of industrial pretreatment and toxics control programs.
- Evaluate the monitoring and reporting program and make recommendations for improving the program.

The receiving water monitoring program consists of assessing water quality and ocean sediment chemistry, evaluating community structures of benthic biota, bottom fish, and epibenthic macroinvertebrates, and determining the bioaccumulation of pollutants in various marine organisms. Table 1-2 summarizes the sampling schedule for various elements of the monitoring and reporting program conducted during 2014.

Table 1-2. Schedule for NPDES Monitoring, Goleta Sanitary District, 2014

Monitoring Program Component	Frequency	Schedule
Standard Wastewater Parameters	Daily - Weekly	As Specified
Influent and Effluent Metals	Monthly	Every Month
Acute Toxicity	Quarterly	Jan, April, July, and Oct
Chronic Toxicity	Quarterly	Jan, April, July, and Oct
Influent and Effluent Priority		
Pollutants	Annually	October
Surf-Zone Bacteria	Weekly	Every Month
Receiving Water Bacteria	Quarterly	Jan, April, July, and Oct
Ocean Water Quality	Quarterly	Jan, April, July, and Oct
Benthic Sediments	Annually	October
Benthic Biota	Annually	October
Fish Trawls	Annually	October
Outfall Inspection	Annually	October
Bioaccumulation	Annually	October

Influent, effluent, and receiving water monitoring is conducted in accordance with U.S. Environmental Protection Agency approved test procedures as stipulated under Title 40 of the Code of Federal Regulations, Section 136 (40 CFR 136): *Guidelines establishing test procedures for the analysis of pollutants*. Water quality analyses for compliance monitoring are performed by analytical laboratories certified by the California Department of Health Services. Bioassay testing is conducted in accordance with guidelines approved by the State Water Resources Control Board and the EPA.

In order to comply with a request from the Central Coast RWQCB in a letter dated June 27, 2008 the District is no longer submitting hard copies of NPDES reports to the RWQCB.

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All documents are converted into a searchable PDF format and are submitted electronically.

WASTEWATER TREATMENT UPGRADING PROJECT BACKGROUND

A condition of the November 2004 301(h) permit renewal included a provision to upgrade the current blended effluent treatment process to full secondary. Under a negotiated settlement agreement between the RWQCB and GSD the District agreed to follow a detailed conversion schedule to ensure that the treatment plant was discharging full secondary treated effluent by November 2014. The conversion schedule is shown below. The District completed the project in December 2013, the ninth year of the 10 year conversion timeline, almost a full year before the November 2014 required date.

The District awarded the facility planning contract to HDR Engineering and the environmental review contract was awarded to Tetra Tech. A preliminary draft of the facilities plan was sent to the Regional Water Quality Control Board in December 2006 and was circulated for review to all treatment plant contract users and other interested parties. The final facilities planning document was completed in June 2008.

A separate contract was then awarded to HDR Engineering to initiate and complete the design of the new treatment plant. The treatment plant design process addressed concerns regarding cultural resources and construction costs and allowed GSD to retain the use of current structures while planning for future regulatory changes. The new secondary treatment structures include the construction of a second biofilter, identical to the existing, an aeration basin and two new secondary sedimentation tanks. Primary and secondary solids will be co-thickened in mechanical thickeners located in a newly constructed solids handling building. The solids treatment will continue with anaerobic digestion, lagoon stabilization and finally, conversion to biosolids with two new screw presses.

Other features of the proposed upgraded plant include:

- □ Conversion of stabilization basin #1 into a flow equalization basin
- Construction of a second biofilter with a total media depth of 6 feet to match the existing biofilter
- Construction of a three train aeration basin with stub outs to add a fourth train at a future date if needed
- Construction of two new secondary clarifiers, and
- Construction of a solids handling building that will house the mechanical thickeners, polymer tanks and screw presses.

The District met all timelines specified in the conversion schedule except for C. 2. Environmental Review & Permitting, Certification of Final CEQA Document. The January 31, 2009 deadline to complete this milestone proved to be unattainable due to a flaw in the original negotiated agreement. The District requested and received, from the RWQCB, an extension for this task. The extension request was based on force majeure reasons caused by unforeseen cultural resource issues at the treatment plant site. The District

conducted an extended phase 1 archaeology study in the areas of the new biofilter, proposed DAFTs and new pipeline corridors. Two inch geoprobes were drilled approximately every 10 meters and the contents of the geoprobe were examined by a geomorphologist, an archaeologist and a Native American representative. The results of this extensive archaeological investigation indicated that the location proposed for the DAFTs had the possibility of containing some Native American artifacts and the District was advised to relocate these structures. Eventually the design was modified and the DAFT structures were removed and replaced with mechanical thickeners that were located in the southern portion of the plant.

No indication of artifacts were found in the location proposed for the new biofilter and corresponding pipeline corridors. These structures did not need to be relocated, however archaeologists and native American monitors were on site during the excavation of these areas.

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Tasks		Date of Co	ompletion*
A. <u>Prelimina</u>	ry Activities:		
 Submittal 	of Detailed Conversion Plan and Time	line 01/0	01/05
to Owners	of Capacity in District's Plant		
	ion of Conversion concepts w/Owners	of 06/3	30/05
capacity ir	District's Plant (Education regarding		
participation	on in conversion)		
Send Req	uests for Environmental & Consulting	12/3	31/05
Engineerir	ng Contracts		
Award of I	Environmental & Consulting Engineerir	ng 06/3	30/06
Contracts			
	Planning:		
	e Draft Facilities Plan		31/06
•	Final Facilities Plan	06/3	30/08
	tal Review & Permitting:		
	& Circulate Draft CEQA Document	06/3	30/08
,			30/10
	pplications for all Necessary Permits		31/09
	Necessary Permits	01/3	31/11
D. <u>Financing</u>			
 Complete 	Draft Plan for Project Design &	01/3	30/07
	ion Financing		
-	Final Plan for Project Design &	03/3	31/08
	ion Financing		
	oof that all Necessary Construction		31/10
	has been Secured, Including Complia	nce	
•	osition 218		
	Construction:		
 Initiate De 	•		30/08
2.30% Desi	•		31/08
3.60% Desi	gn	11/3	30/09
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04/30/14

 4. 90% Design
 03/31/10

 5. 100% Design
 09/30/10

 6. Issue Notice to Proceed to Contractor
 04/30/11

7. Construction Progress Reports Quarterly (w/self monitoring reports)

8. Complete Construction & Commence Debugging and Startup

9. Full Compliance w/Secondary Requirements 11/01/14

By the end of December 2010, the District was successful in meeting all regulatory conditions and received all permits necessary to complete the project. For reference purposes, the following permits have been approved:

Permitting Agency	Type of Permit	Permit Number
Santa Barbara County	Government Code Consistency	09GOV-00000-00001
Santa Barbara County	Revised Development Plan	09RVP-00000-00001
Santa Barbara County	Grading Permit	09GRD-00000-00073
Santa Barbara County	Coastal Development Permit	09CDP-00000-00099
California Coastal Commission	Coastal Development Permit	4-09-011
Santa Barbara County	Land Use Permit	10LUP-00000-00235
Santa Barbara County	Land Use Permit	10LUP-00000-00360
	Authority to Construct	13378
Pollution Control District		
Santa Barbara County	Grading Permit	10GRD-00000-00075
Santa Barbara County	Building Permit	10BDP-00000-00553
Goleta Water District	Can & Will Serve Letter	

PCL Construction company was the low bidder and was awarded the construction contract. Their bid submittal was for \$28.6 M. The final cost of construction is still to be determined as change orders and costs of those change orders are still under discussion. To date the cost of the project has reached \$31 M. Mobilization took place in April 2011 and construction started in May 2011. A total of ten quarterly construction progress reports were prepared and submitted to the state and regional water quality control boards and several other interested parties. The last quarterly construction report was submitted on January 27, 2014 and covered the last quarter of construction work from July 1, 2013 to September 30, 2013. By the end of December 2012 all new structures had been built. The new biofilter, the aeration basin and one of the new secondary clarifiers had been put on line and were operational.

The plant began producing full secondary treated wastewater on May 16, 2013 when the final tie in was completed. PCL construction demobilized September 2013 and the project was deemed complete by December 2013.

^{*}Any completion date falling on a Saturday, Sunday or State Holiday shall be extended until the next business day. The district shall submit proof of completion of each task within 30 days after the due date for completion.

REPORT ORGANIZATION

This report summarizes data collected during the 2014 monitoring and reporting program, and analyzes this data to determine compliance with the discharge permit terms and conditions. Chapters in this report have been organized to parallel sections of the monitoring and reporting program. The chapter sequence also follows the flow of wastewater as it undergoes treatment in the plant, as it is discharged to the marine receiving waters, and as it encounters nearby sediments and resident biota. Chapter 9 presents a summary of the lift station and collection system overflows, the causes of the overflows, the corrective actions taken, and any corrective actions planned. Chapter presentation is as follows:

Chapter 4 Physical Characteristics of Benthic Sediments Chapter 5 Chemical Characteristics of Benthic Sediments Chapter 6 Biological Characteristics of Benthic Sediments Chapter 7 Fish Populations	Chapter 5 Chapter 6 Chapter 7 Chapter 8 Chapter 9	Chemical Characteristics of Benthic Sediments Biological Characteristics of Benthic Sediments Fish Populations Chemical Characteristics of Fish and Mussel Tissue Outfall Dive Survey Collection System Summary
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CHAPTER 2

TREATMENT PLANT PERFORMANCE

The performance of a wastewater treatment plant is measured by its ability to reduce influent contaminants to levels acceptable for discharge to the environment. Federal and state authorities mandate these levels of treatment in order to protect the marine environment. Proper operation of the Goleta Sanitary District's wastewater treatment plant is assured through the monitoring of several effluent parameters such as flow, total suspended solids, biochemical oxygen demand, residual chlorine, hydrogen-ion concentration (pH), turbidity, ammonia, settleable solids, oil and grease, and toxicity concentration. Metals, pesticides, and other priority pollutants are also analyzed to aid in determining the impact the wastewater discharge has on receiving waters, evaluating compliance with discharge permit limitations, and monitoring the effectiveness of the industrial pretreatment and toxic control program.

WASTEWATER CHARACTERIZATION

Goleta Sanitary District's NPDES monitoring program requires measurement of many parameters at frequencies ranging from continuous to once per year. During 2014, influent, effluent, biosolids (sludge), and surf zone samples were collected by treatment plant personnel, and analyzed by the Goleta Sanitary District wastewater treatment plant laboratory and various contract laboratories such as: Aquatic Bioassay Laboratories for ocean monitoring, Aquatic Testing Laboratories (ATL) for acute and chronic toxicity, FGL Environmental Laboratories and Vista Analytical Laboratory, Weck Laboratories as subcontractors to FGL. Treatment plant personnel monitored and analyzed wastewater for performance-evaluating parameters including wastewater flow, suspended solids, biochemical oxygen demand (BOD), pH, turbidity, settleable solids, ammonia, oil and grease, temperature, residual chlorine, coliform and enterococcus bacteria. analyses for influent and effluent metals were performed by FGL Environmental Laboratories of Santa Paula, CA. FGL Environmental Laboratories, and their certified subcontract laboratories performed annual analysis of priority pollutants and other parameters in influent, effluent, and biosolids samples. Influent and effluent samples were also analyzed for radioactivity. Bioassay tests for acute and chronic toxicity concentration were performed quarterly by Aquatic Testing Laboratory.

Analytical methodologies used by Goleta Sanitary District Laboratory and other contract laboratories used by GSD are based on approved U.S. Environmental Protection Agency (EPA) methods (EPA 1983; Federal Register 1984) and other methods in *Standard Methods for the Examination of Water and Wastewater, 21st ed.* (Standard Methods 2005). All methodologies employed during 2014 were approved for NPDES monitoring programs. Quality assurance and quality control procedures followed those presented in *Standard Methods for the Examination of Water and Wastewater, 21st edition.*

Results of the wastewater chemical analyses used to monitor proper operation of the treatment plant during 2014, and the respective discharge permit limitations, are presented

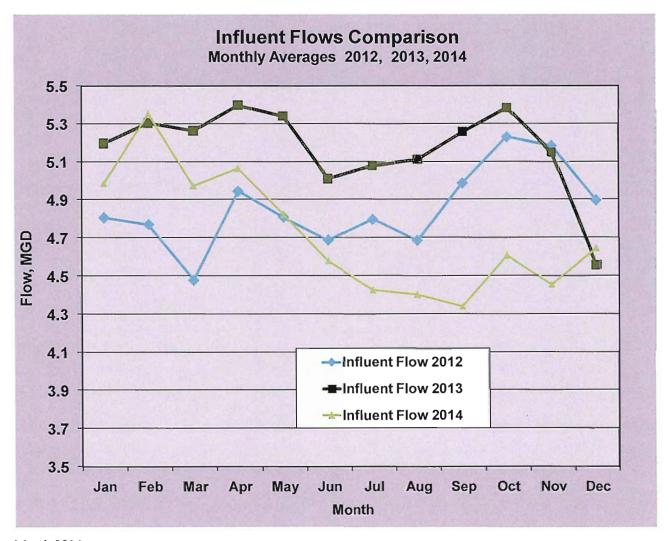
in Tables 2-1 and Table 2-2. All monthly averaged data presented in these tables are calculated from daily values at the treatment plant, with the exception of removal efficiencies, which are calculated from the monthly averages of the respective influent and effluent parameters.

Influent Flow

The daily influent flow into the treatment plant was monitored continuously throughout 2014. Influent flow without the internal plant recirculated flow, averaged 4.698 million gallons per day (MGD) a 9% decrease compared to the average of 5.174 MGD that was treated in 2013.

Overall, the average monthly influent flows for 2014 varied throughout the year, fluctuating from a low of 4.35 MGD in September to a high of 5.07 MGD in April. The decrease in average influent flow observed at the plant is likely due to water conservation implemented by residents in response to the drought conditions. See Figure 2-1.

Figure 2-1. Influent Flows Monthly Average Comparison for 2012, 2013 and 2014



The highest flows into the plant during 2014 occurred in April, and may be associated with heavy rains that occurred in March.

Since 2001 the Goleta West Sanitary District and Goleta Sanitary District have maintained an aggressive collection system rehabilitation program. Numerous sections of the collection system in both Districts have been relined or replaced to correct structural deficiencies while significantly reducing the inflow and infiltration (I&I) problems. However, even with the reduction of I&I the amount of rainfall during the year can affect the total amount of influent flow measured. The District's storm water pollution prevention plan requires all storm water collected from process areas to be treated before disposal. After several dry years the low ground water table and dry creeks can reduce the potential for ground water intrusion into the collection systems.

Effluent Flow

The effluent flow from the treatment plant was monitored continuously during 2014 and averaged 3.6 MGD for the year. The difference between the influent and effluent flow is due to the production of reclaimed water, which is not discharged into the ocean but is distributed throughout the community for landscape irrigation and other uses.

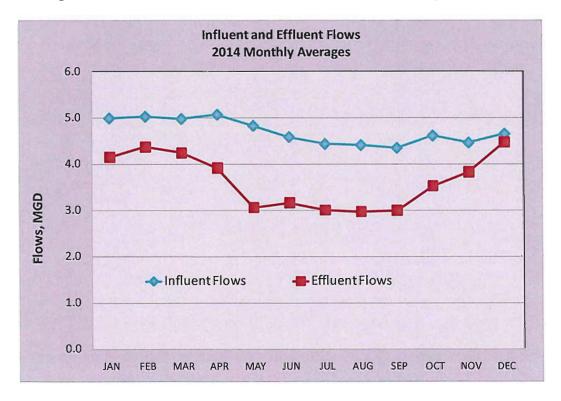


Figure 2-2. Influent and Effluent Flows 2014 Monthly Averages

Figure 2-2 shows the monthly average influent and effluent flows for 2014. Higher wastewater effluent flow generally occurs during the winter months when influent flow is also the highest and recycling is minimal. The most important factor contributing to

fluctuations in the effluent flow is the amount of wastewater that is processed into reclaimed water and used for irrigation. The lowest effluent flow occurred during August when the amount of flow discharged to the Pacific Ocean dropped to 2.96 MGD as depicted in Figure 2-2. The temporal variations in the monthly average effluent flow seen in 2014 fluctuated from a low of 2.96 MGD in August, when the daily production of reclaimed water was the second highest production month of the year and averaged 1.44 MGD for the month to a high of 4.48 MGD during December when the reclaimed facility was on line for 5 days out of the month and a total of 5.4 million gallons were filtered. There was also significant rainfall during December with approximately 2.97 inches of rain. Figure 2-2 is a time history of the influent and effluent flows and Table 2-1 shows the actual monthly flow average values.

Table 2-1. Monthly Averages Flow, Suspended Solids and BOD, Goleta Sanitary District, 2014.

	Flo	w	Total Suspended Solids			Biochemical Oxygen Demand				
Month	Influent MGD	Effluen MGD	Influent mg/L	Effluent mg/L	Removal (%)	Mass Emission (lbs/day)	Influent mg/L	Effluent mg/L	Removal (%)	Mass Emission (lbs/day)
Jan	4.99	4.15	345	7.7	98	233	326	4.1	99	124
Feb	5.03	4.37	358	6.5	98	204	390	4.1	99	131
Mar	4.98	4.24	469	6.8	98	200	431	3.8	99	114
Apr	5.07	3.91	351	8.3	98	268	396	3.9	99	127
May	4.83	3.06	370	9.6	97	247	374	4.2	99	108
Jun	4.58	3.16	337	5.6	98	150	323	6.7	98	169
Jul	4.43	4.43	329	5.9	98	220	363	5.5	98	204
Aug	4.41	2.96	303	4.2	99	102	293	2.9	99	71
Sep	4.35	2.99	333	4.2	99	102	326	4.5	99	110
Oct	4.61	3.52	340	5.9	98	174	355	5.2	98	153
Nov	4.46	3.83	364	6.0	98	191	392	2.9	99	92
Dec	4.65	4.48	331	5.7	98	204	310	2.8	99	103
Average	4.70	3.76	352	6.4	98	191	357	4.2	99	125
Limit	NL	7.64	NL	63	75	4010	NL	98	30	6240

^{**}ND = Non-Detected

NL = No Limit

Suspended Solids

Influent and effluent suspended solids were measured five days per week on 24-hour composite samples. The effectiveness of the treatment plant in removing suspended solids is demonstrated by the variation of influent solids versus the low-level and consistent output of effluent solids (see Figure 2-3). Influent suspended solids concentrations averaged 352 mg/L for the year an increase of about 6% from the 2013 annual average of 331 mg/L which was a 9% increase from the 300 mg/L annual average of 2012. For the past three years the concentration of suspended solids entering the plant has been on a steady increase. Figure 2-3 below shows a spike in concentration of suspended solids that occurred during March in the influent TSS.

The treatment process reduced the concentration of total suspended solids in the effluent to an annual average of 6.4 mg/L a 73% annual decrease of the 24 mg/L average of 2013.

All 30-day monthly averages were well below the 63-mg/L monthly average limitation. Overall removal efficiency for the year was an average of 98 percent, see Table 2-1.

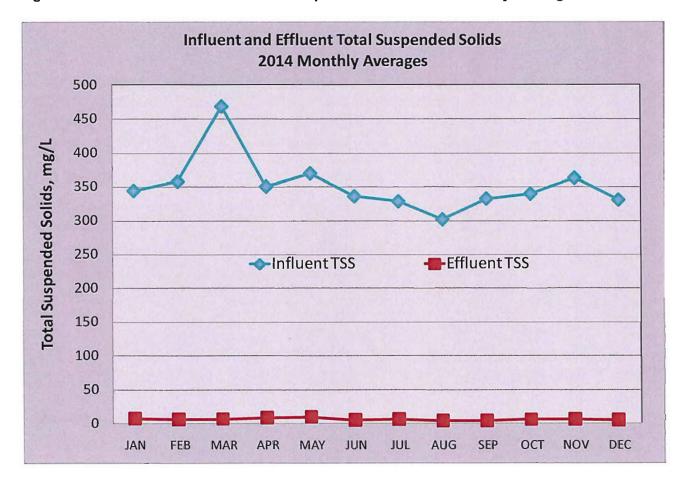


Figure 2-3. Influent and Effluent Total Suspended Solids 2014 Monthly Averages

Average monthly suspended solids mass loading rates for 2014 are represented graphically in Figure 2-4. The mass emission limit is based on average dry weather flow (ADWF) and is a limit applied to dry weather flows (DWF). There is no limit for mass emissions on wet weather flows.

The maximum average monthly mass emission loading for 2014 occurred in April at a high of 268 lbs/day, which is approximately seven percent of the permitted monthly 30-day average limit of 4,010 lbs/day. Loading rates were well below the discharge limit throughout the year.

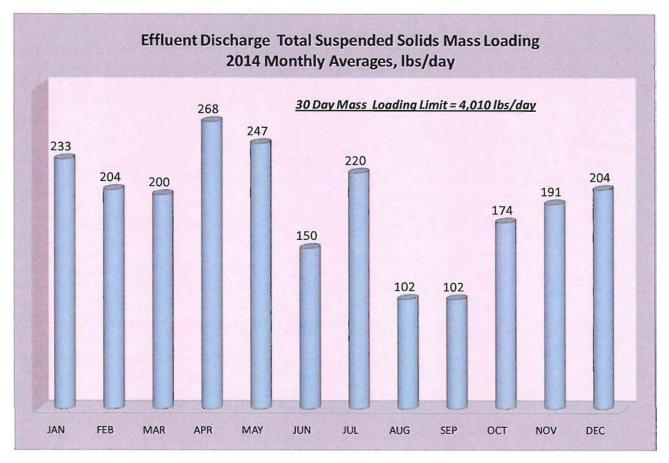


Figure 2-4. Effluent Discharge Total Suspended Solids Mass Loading, 2014 Monthly Averages, Ibs/day

Biochemical Oxygen Demand

Biochemical oxygen demand (BOD) levels were measured on 24 hour composite samples of the influent and effluent, at least three and five days per week, respectively.

During 2014 influent BOD averaged 357 mg/L showing a small increase from the annual influent average of 305 for 2013, and 282 for 2012. The increase in the influent BOD concentration mirrored that of the influent total suspended solids concentrations and may be reflective of a decrease in water use by the community in response to the beginning of drought conditions. The influent BOD varied throughout the year, ranging from a monthly average low of 293 mg/L in August to a high of 431 mg/L in March.

A small variation in the monthly average final effluent BOD concentration was observed throughout the year with the annual average of 4.2 mg/l and the range extending from a low of 2.8 in December to a high of 6.7 in June, (Table 2-1). The difference between influent and effluent BOD represents an overall removal rate of 99 percent.

The NPDES effluent BOD monthly average limitation and the maximum at any time limitation are 98 mg/L and 150 mg/L, respectively. All BOD NPDES limitations were achieved throughout the year.

Figure 2-5. Influent and Effluent Biochemical Oxygen Demand 2014 Monthly Averages

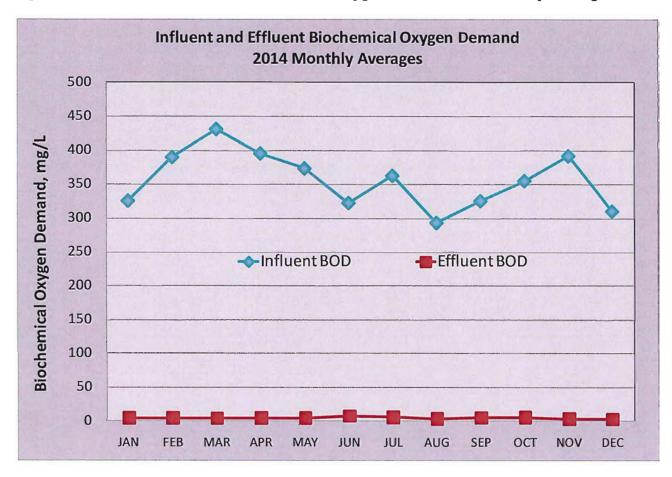


Table 2-2. Monthly Averages of Influent and Effluent Parameters, Goleta Sanitary District, 2014

				Settleable					Тох	icity
	р	Н	Turbidity	Solids	Ammonia	0	il and Grea	ise_	Acute	Chronic
						,		Mass		
			Effluent	Effluent	Effluent			Emission	Effluent	Effluent
Month	Influent	Effluent	(NTU)	(mL/L/hr)	(mg/L)	Influent	Effluent	(lbs/day)	(TUa)	(TUc)
Jan	7.8	6.4	3	0.14	ND	29.8	ND	9	0.0	5.6
Feb	7.7	6.5	3	0.17	ND	33.0	8.3	50		
Mar	7.7	6.6	3	0.14	ND	46.3	4.8	27		
Apr	7.6	6.4	3	0.16	ND	42.3	ND	11	0.0	3.1
May	7.6	6.5	4	0.17	ND	48.7	4.0	102		
Jun	7.7	6.4	2	0.14	ND	57.6	ND	26		
Jul	7.5	6.6	2	0.17	ND	44.2	ND	54	0.0	3.1
Aug	7.5	6.6	2	0.15	ND	30.9	ND	68		
Sep	7.5	6.8	2	0.12	ND	32.8	ND	55		
Oct	7.5	6.6	2	0.14	ND	34.5	ND	24	0.41	3.1
Nov	7.5	6.6	2	0.13	ND	34.7	ND	25		
Dec	7.6	6.7	2	0.12	ND	39.5	3.0	114		
Average	7.6	6.6	2	0.15	<1	39.5	2.1	47	0.10	3.7
Limit	NL	6 to 9	75	1.0	74	NL	25	1590	4.0	123

^{**}ND = Non-Detected

NL = No Limit

In 2014, all effluent BOD mass emission values were below all limitations. The maximum monthly average mass emission was 204 lbs/day for July. The mass emission limit is based on average dry weather flow (ADWF) and is a limit, which is only applied to dry weather flows (DWF). There is no limit for mass emissions on wet weather flows. The mass emissions monthly average limitation of 6,240 lbs/day and the maximum at any time limitation of 9,560 lbs/day were never exceeded during 2014.

Hydrogen-Ion Concentration (pH)

Influent and effluent pH levels were monitored five days per week to ensure that the effluent remained within an acceptable range when discharged into the ocean. Influent pH averaged 7.6 units for the year; effluent pH averaged 6.6 units. The NPDES effluent pH limitations are established as a minimum of 6.0 and a maximum of 9.0 pH units, all pH values were well within these limitations for 2014.

Ammonia

The effluent was monitored monthly to determine the concentration of ammonia. The permit specifies six-month median, daily maximum, and instantaneous maximum limitations of 74 mg/L, 300 mg/L, and 740 mg/L, respectively. The monthly measured ammonia concentration was below the lowest calibration standard at 1.0 mg/L throughout

the year (Table 2-2). The monthly average for the year was <1 mg/L. The values for ammonia were well below all their respective permit limitations.

Turbidity

Effluent turbidity was monitored five days per week. The permit limitations for effluent turbidity consists of a monthly average of 75 Nephelometric Turbidity Units (NTU), a weekly average of 100 NTU, and a maximum at any time limitation of 225 NTU. Effluent turbidity data are shown graphically in Figure 2-6. The maximum value at any time, 8.4 NTU, occurred on May 21 which was still well below the effluent limits. Monthly averages ranged from a low of 2 NTU to a high of 4 NTU. All values were significantly below their respective permit limitations.

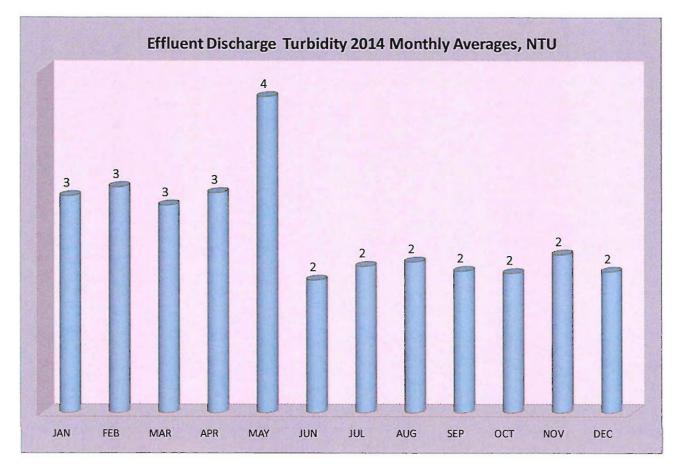


Figure 2-6. Effluent Discharge Turbidity 2014 Monthly Averages, NTU

Acute Toxicity Concentration

All quarterly acute toxicity tests were performed on 24-hour composite effluent samples. The acute toxicity has a daily maximum limit of 4.0 acute toxicity units (TUa). All four quarterly acute toxicity samples for 2014 were collected under the conditions of the new NPDES WDR Order No. R3-2010-0012 which requires the District to use Topsmelt as the acute toxicity test species, replacing fathead minnow larvae. The annual average acute toxicity value was 0.10 TUa. (See Table 2-2). All values were below the permit limitation of 4 TUa.

Chronic Toxicity Concentration

The effluent was analyzed for chronic toxicity (TU_C) on a quarterly basis in January, April, July, and October. The special testing conducted during 2011 to identify the most sensitive chronic toxicity organism showed that the abalone development test was the most sensitive. All results were well below the daily maximum limitation of 123 TU_C .

Settleable Solids

The effluent was monitored for settleable solids concentrations 5 days per week. The permit specifies that the monthly average, weekly average, and maximum at any time may not exceed 1.0 milliliters/liter/hour (ml/L/hr), 1.5 ml/L/hr, and 3.0 ml/L/hr, respectively. Monthly averages ranged from 0.12 ml/L/hr to 0.17 mL/L/hr. The maximum value at any time was 0.4 mL/L/hr which occurred twice during February. All values were well below their respective permit limitations.

Oil and Grease

Influent and effluent oil and grease were monitored bi-weekly (once every two weeks) and weekly, respectively. Monthly average results are shown graphically in Figure 2-7. Prior to August 2007 Freon was the solvent used in the standard method to extract oil and greases from water samples. According to EPA regulations, in August 2007 the GSD laboratory ceased using Freon as the extraction solvent and began using hexane as the required solvent. The District continued to use the liquid-liquid extraction method, the only change at this time was the solvent. In December 2010, the GSD laboratory began analyzing for oil and grease using the approved standard solid phase extraction (SPE) method.

Influent grease and oil results were varied throughout the year. Average monthly concentrations spiked in June due to one high sample on the 22nd with a result of 63.8 mg/L which caused the increase in the monthly average. The influent annual average value of 39.5 mg/L was reduced to an annual average of 3 mg/L in the final effluent resulting in an 92 percent annual average removal rate.

Effluent grease and oil concentrations were very consistent during 2014. All monthly, weekly, and maximum permit limits were met. Mass emissions values ranged from a monthly average low of 9 lbs/day in January to a high of 114 lbs/day in December. Both are well below the permit limitation of 1,590 lbs/day. Monthly average oil and grease concentrations in the effluent ranged from 0.4 mg/L which is below the method detection limit to 8.3 mg/L in February. (Table 2-2). All permit limitations for effluent oil and grease were met during 2014.

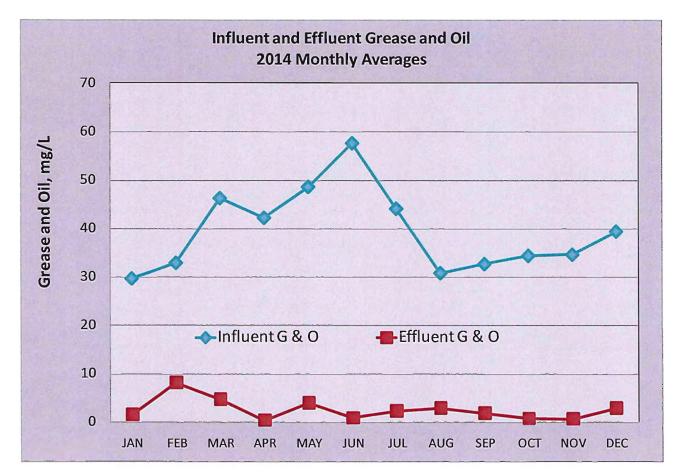


Figure 2-7. Influent and Effluent Grease and Oil 2014 Monthly Averages

Temperature

Effluent temperature was sampled five days per week throughout 2014. The data reflect a typical response to seasonal changes (Figure 2-8). The coolest temperatures occurred during January and February with average monthly temperatures of 20.4 $^{\circ}$ C and 20.6 $^{\circ}$ C. A warming trend continued throughout the spring and summer months to reach a monthly averaged high in July and August of 25.8 $^{\circ}$ C and 25.9 $^{\circ}$ C. As expected, the year ended with a cooling trend during November and December.

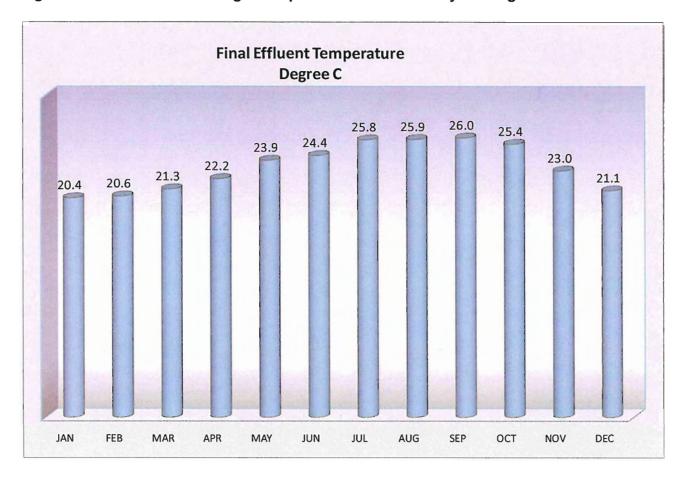


Figure 2-8. Effluent Discharge Temperature 2014 Monthly Averages

Wastewater Disinfection

Sodium hypochlorite is used to disinfect the treated wastewater at the Goleta Sanitary District. The sodium hypochlorite is flash mixed into the wastewater at the beginning of the chlorine contact channel. At an average effluent flow rate of 4 MGD, the chlorine is in contact with the wastewater for approximately 2½ hours (145 minutes). The NPDES permit specifies that the District must maintain a total chlorine residual of at least 5 mg/L at the end of the chlorine contact channel under total suspended solids peak loading conditions. The Goleta Sanitary District maintains its chlorine contact tank to provide maximum chlorination effectiveness at all times. The chlorine residual at the end of the chlorine contact channel averaged 6.6 mg/L during 2014. The average monthly values are reported in Table 2-3.

After the disinfection process is completed, the sodium hypochlorite is neutralized (dechlorinated) by adding sodium bisulfite to the wastewater stream. This process lowers residual chlorine to levels that are environmentally safe, before discharge to the ocean such that the chlorine poses no risk to the receiving water environment. Treatment plant

personnel continuously monitor the residual chlorine levels as required by the NPDES permit.

The permit limitations for residual chlorine in the effluent immediately prior to discharge and after dechlorination are as follows: 6-month median of 0.25 mg/L, daily maximum of 0.98 mg/L, and instantaneous maximum of 7.4 mg/L. After dechlorination, the monthly average residual chlorine levels were very consistent throughout the year; at or below the detection limit of 0.1 mg/L for all months. The monthly average values are shown in Table 2-3. No chlorine residual exceedences occurred during 2014.

Effluent Coliform Bacteria

The effluent was analyzed five days a week for coliform bacteria. The monthly average values for total coliform, fecal coliform, and enterococcus bacteria detected in the effluent are presented in Table 2-3. Monthly average values ranged from 8 to 64 MPN/100 mL for total coliform and from 2 to 5 MPN/100 mL for fecal coliform. The permit prohibits more than 10 percent of the final effluent samples, in any thirty-day period, to exceed a total coliform density of 2,400 MPN/100mL with no sample exceeding a total coliform concentration of 16,000 MPN/100mL. The maximum total coliform concentration was measured on November 6th at 540 MPN/100mL.

Effluent Enterococcus Bacteria

The effluent was also analyzed five days a week for enterococcus bacteria. The monthly mean values are presented in Table 2-3 and the values were consistently low throughout the entire year, thereby demonstrating the effectiveness of the chlorination process.

Table 2-3. Chlorine and Bacteria Monthly Averages, 2014

Month	Chlorine at the end of the CCC	Chlorine after Dechlorination	Total Coliform	Fecal Coliform	Enterococcus
	mg/L	mg/L	THE RESERVE OF THE PERSON OF T	MPN/100	mL
January	7.3	< 0.1	25	5	1.8
February	7.0	< 0.1	12	3	1.8
March	7.1	< 0.1	8	3	1.8
April	6.6	< 0.1	26	3	2.0
Мау	6.2	< 0.1	35	5	1.8
June	6.5	< 0.1	34	3	1.8
July	6.4	< 0.1	52	5	1.9
August	5.7	< 0.1	35	3	1.8
September	6.4	< 0.1	12	2	1.8
October	6.5	< 0.1	20	2	1.8
November	6.8	< 0.1	46	5	1.8
December	6.7	< 0.1	64	4	1.8

SURF ZONE BACTERIA

The Goleta Sanitary District has an extensive bacteria monitoring program that measures the concentrations of enterococcus, total coliform, and fecal coliform groups of bacteria at the end of the treatment process immediately before discharge to the ocean, at the end of the pipeline in the zone of initial dilution, at far shore and near shore ocean sampling locations and in the surf zone at stations extending west from Goleta Point to 1,000 meters east of the outfall line. Table 2-4 summarizes the locations and frequency of all bacteria monitoring conducted at the Goleta Sanitary District.

Table 2-4. Bacteria Monitoring Program

Location	Frequency of Total Coliform, Fecal Coliform and Enterococcus Bacteria Testing
Final Effluent prior to ocean	
discharge	5 days/week
Zone of Initial Dilution in	Quarterly: 3 samples at each location;
the discharge plume at 25	1m below surface, mid-depth and 1 m above bottom
m and 100 m from outfall	
pipe	
Far Shore (ocean) Stations;	Quarterly: 3 samples at each location;
B1, B2, B3, B4, B5 and B6	1m below surface, mid-depth and 1 m above bottom
Near Shore (ocean)	Quarterly: 3 samples at each location;
Stations; K1, K2, K3, K4	1m below surface, mid-depth and 1 m above bottom
and K5	
Surf Zone Stations; A, A1,	Weekly
A2, B, C, D, E	

Final effluent samples and weekly receiving water surf zone samples are collected and analyzed in-house by GSD personnel the results of which are discussed in this chapter. Zone of initial dilution, far shore and near shore bacteria samples are collected and analyzed by ABC Laboratories of Ventura. Results of this testing is presented in chapter 3.

Approximately 358 samples are collected each year from the surf zone and each sample is analyzed for total coliform, fecal coliform and enterococcus for a total of approximately 1,074 bacteria tests conducted every year. These samples are collected and indicator organism concentrations are monitored in order to ensure that the beneficial uses of the Goleta Beach coastal area are protected. The following section discusses the 2014 bacterial trends found in the surf zone environment.

Surf-zone Stations.

Consistent with historical trends, bacteria monitoring at surf-zone stations usually yield more frequent and higher amounts of coliform bacteria than at the near shore and farshore (ocean) stations and even from the final effluent that is discharged to the ocean. The occurrence of bacteria in the shoreline monitoring area is often in response to the

drainage, tidal flushing, and dredging of Goleta Slough. Over the years it has been determined that coastal bird populations, organic beach debris (including dog waste), and most importantly, the urban flushing effects of storm water runoff can be contributors to high surf zone bacteria concentrations. There has never been any indication that the treatment plant discharge has contributed to bacteria concentrations along the shoreline.

Goleta Slough, which is the confluence of the San Jose, Atascadero, and San Pedro creeks, is a slow-flowing, estuarine water body, which discharges directly into the Pacific Ocean between two of the Goleta Sanitary District's monitoring stations (stations D and E). Because the slough receives little flushing (except during storm runoff episodes) and is a rich waterfowl habitat, slough waters are relatively high in organics and coliform bacteria with respect to surf-zone waters.

Concentrations of bacteria at surf-zone stations in 2014 in general, were higher than that observed in the effluent, offshore and near shore ocean stations. This is consistent with the results of earlier years. Throughout the year, annual average levels of bacteria at surf-zone monitoring stations ranged from 10 to 130 MPN/100mL for total coliform, 7 to 117 MPN/100mL for fecal coliform bacteria and <1.8 to 11 MPN/100mL for enterococcus bacteria. Several maximum one time exceedences occurred throughout the year and were reported in the corresponding monthly report. Table 2-5 is a summary of the 2014 surf zone exceedences.

Table 2-5. Surf Zone Exceedences 2014

<u>Date</u>	Station	Exceedence Limit (Result)	Possible Cause	Final Effluent Result
4/13/2014	С	One time fecal coliform >= 400 MPN/100mL (540 MPN/100mL)	No clear reason	< 1.8 MPN/100mL
5/18/2014	A1	One time fecal coliform >= 400 MPN/100mL (540 MPN/100mL)	No clear reason	
5/18/2014	D	One time fecal coliform >=400 MPN/100mL (920 MPN/100mL)	No clear reason	
7/15/2014	С	One time fecal coliform >=400 MPN/100mL (540 MPN/100mL)	No clear reason; however, over 100 kids observed by pier.	13 MPN/100mL
7/30/2014	D	One time fecal coliform >=400 MPN/100mL (540 MPN/100mL)	No clear reason	< 1.8 MPN/100mL
9/3/2014	D	One time fecal coliform >=400 MPN/100mL (540 MPN/100mL)	No clear reason	< 1.8 MPN/100mL
9/23/2014	A2	One time fecal coliform >=400 MPN/100mL (540 MPN/100mL)	No clear reason	< 1.8 MPN/100mL
11/20/2014	A2	One time enterococcus >= 104 MPN/100mL (170 MPN/100mL)	No clear reason	< 1.8 MPN/100mL
11/20/2014	В	One time enterococcus >= 104 MPN/100mL (350 MPN/100mL)	No clear reason	< 1.8 MPN/100mL
11/20/2014	С	One time enterococcus >= 104 MPN/100mL (140 MPN/100mL)	No clear reason	< 1.8 MPN/100mL
11/20/2014	D	One time enterococcus >= 104 MPN/100mL (170 MPN/100mL)	No clear reason	< 1.8 MPN/100mL

Throughout the year the final effluent samples analyzed previous to and on the surf zone collection days indicated no or very low concentrations of coliform and/or enterococcus bacteria, see Table 2-5.

For safety reasons due to high surf conditions and potential rip tides samples were not collected at one or more stations on the following dates. No sample was collected from Station E on March 20th. No samples were collected from Stations A2 and E on December 4th and December 10th. No sample was collected from Station E on December 24th. Station A2 was inaccessible due to high tide causing the water level to submerge the staircase. Station E was inaccessible due to high flows in the Goleta Slough and the Slough could not be crossed safely by District staff.

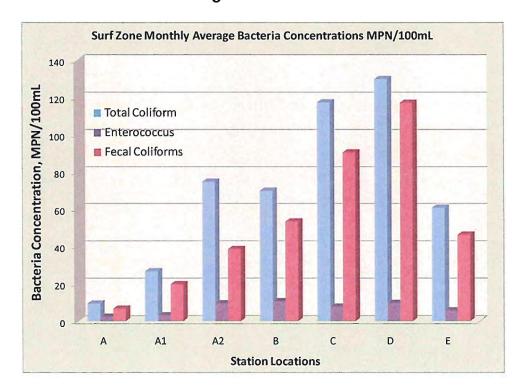


Figure 2-9. Surf Zone Annual Average Bacteria Concentrations 2014

Figure 2-9 shows the impact of the Goleta Slough discharge on the surf zone samples. Goleta Slough empties between station location D and E which show some of the highest overall annual average bacteria concentrations for all three indicator organisms measured weekly. Station A, located at Campus Point, the furthest point west with the "cleanest" samples.

Effluent bacteria samples collected at the end of the treatment and disinfection process, during these same time periods showed low or undetected concentrations of bacteria discharged from the treatment plant demonstrating that the effluent was not a source for the high surf zone bacteria concentrations.

The impact of Goleta Slough on bacteria water quality in the surf zone of the study area has been documented for the past 22 years. This historical data has shown, year after year that the highest concentration of indicator organisms are found in and adjacent to the Goleta Slough mouth and are associated with storm water run off.

Metals

Twenty four-hour composite samples of influent and effluent were collected monthly and analyzed for metals (Table 2-6). In all instances, the concentrations of metals in the effluent for 2014 (Table 2-6) were low or undetected and were well below all permit limitations. Although the wastewater treatment process is not particularly efficient at removing metals, hence the need for the pretreatment program. Metal concentrations in the influent were consistent throughout the year.

Table 2-6. Influent and Effluent Metals (ug/L), Goleta Sanitary District, 2014.

	Arsenic	Cadmium	Chromium	Copper	Lead	Mercury	Nickel	Silver	Zinc		
	Influent (ug/L)										
January	1.37	0.245	5.56	119	7.39	0.11	8.43	1.1	148		
February	1.32	0.252	4.61	119	1.78	0.06	6.5	0.012	169		
March	4.41	0.796	36.1	387	8.12	0.435	28.6	6.9	503		
April	1.92	0.292	6.92	126	3.45	0.313	9.34	0.698	188		
May	1.66	0.351	7.85	134	2.83	0.115	7.79	0.722	195		
June	1.62	0.245	5.79	145	2.18	0.091	7.3	1.42	164		
July	1.71	0.275	6.58	124	3.42	0.114	7.54	1.15	155		
August	1.31	0.201	4.81	45.9	2.3	0.282	6.34	0.143	143		
September	1.54	0.259	8.33	112	3.09	0.135	10.9	0.819	144		
October	1.42	0.427	6.87	112	3.74	0.156	10.0	1.52	177		
November	1.44	0.259	9.43	119	2.46	0.159	10.6	0.779	160		
December	2.01	0.363	13.1	116	3.29	0.224	12.6	1.15	172		
			Efflue	nt (ug/L)							
January	0.724	0.041	0.373	7.75	13.4	0.030	4.93	0.012	58.4		
February	0.83	0.038	3.23	11.2	0.513	0.0081	4.01	0.012	54.4		
March	1.21	0.049	1.88	7.76	0.566	0.0106	5.12	0.012	81.6		
April	0.92	0.029	0.355	7.15	0.193	0.0085	3.73	0.012	30.8		
May	0.88	0.029	0.368	7.0	0.084	0.0071	2.64	0.012	30.0		
June	0.98	0.034	0.839	6.18	0.248	0.0035	3.19	0.012	57.7		
July	0.989	0.20	0.362	5.32	0.144	0.00999	3.0	0.012	20.2		
August	0.952	0.036	0.316	4.49	0.038	0.07990	3.72	0.012	25.2		
September	0.894	0.044	0.548	5.74	0.083	0.0215	4.58	0.012	27.2		
October	0.949	0.0580	0.735	6.43	3.3	0.0110	5.12	0.012	60.8		
November	1.010	0.0460	0.659	6.7	0.251	0.00903	5.43	0.038	58.5		
December	1.090	0.0320	0.837	8.0	0.223	0.0139	5.4	0.025	73.6		
	Effluent Limits (ug/L)										
6-month											
median	620	120	250	120	250	4.9	620	67	1,500		

Priority Pollutants

The NPDES permit requires priority pollutant analyses to be performed on influent and effluent composite samples annually. Compounds detected in the influent and/or effluent samples are presented in Table 2-7; complete copies of all the laboratory reports listing all the chemical compounds and analytical methods are available for review at Goleta Sanitary laboratory. Seventeen compounds were detected in the influent and eight in the effluent. Concentrations of detected chemicals are all reported as parts per billion except for TCDD and radioactivity which the units are noted next to the parameter in the table.

Results of influent and effluent radioactivity determinations for 2014 are also presented in Table 2-7.

Table 2-7. Detected Priority Pollutants, Goleta Sanitary District, 2014

Parameter, units	Influent,	Effluent,
	ug/L	ug/L
Acetone	1730	ND
Antimony	1.42	1.31
Bis(2-Ethylhexyl)phthalate	3.02	ND
Bromodichloromethane	2.91	51.2
Bromoform	0.788	2.10
Carbon Disulfide	25.8	ND
Chloroform	10.7	71.3
Chloromethane	0.281	ND
Dibromochloromethane	2.30	30.1
Diethylphthalate	2.34	ND
Di-n-octylphthalate	0.537	ND
TCDD, equivalents, pg/L	185	ND
Methylene Chloride	1.52	ND
Styrene	0.299	ND
Toluene	0.313	1.28
Radioactivity, gross Alpha	1.91	0.658
pCi/L	+/- 2.16	+/-2.05
Radioactivity, gross Beta	5.02	10.6
pCi/L	+/-2.13	+/-2.56
ND = Not Detected		

Limits for radioactivity are defined in Title 17 of the California Code of Regulations section 30269, which state limitations of $3x10^{-8} \mu \text{Ci/mL}$ (or 30 pCi/L) for alpha emission and $3x10^{-6} \mu \text{Ci/mL}$ (or 3000 pCi/L) for beta emission. Samples collected during 2014 were below these limitations.

DISCHARGE COMPLIANCE

Throughout 2014 the wastewater discharge from Goleta Sanitary District complied with all applicable permit effluent limitations. All monitored parameters were below their respective limitations as required by the permit. All metals, priority pollutants, and pesticides were low or undetected throughout the year.

OCEAN OUTFALL CONDITIONS

The outfall pipeline, diffuser section, and armor rock protection were inspected by divers from Aquatic Bioassay and Consulting Laboratories, Inc. in October 2014. A report was prepared and videotape was made of the diffuser section and along the outfall pipeline and armor rock.

During the diffuser dive survey, 36 diffuser ports were carefully inspected for flow and general efficiency. The remainder of the outfall pipe was inspected for damage, leaks or evidence of leaks and general stability of the pipe and armor rock. Inspection of the outfall yielded no evidence of damage, holes, cracks, or erosion. The pipe and associated armor rock appeared stable with little or no displacement.

The complete report of the outfall dive survey is included as Chapter 9 of this report. Copies of the outfall dive on DVDs are available at the District for review.

CHAPTER 3

Receiving Water Environment

3.1. Scope and Period of Performance

This report covers the period of field and laboratory studies conducted from January 1, 2014 through December 31, 2014. The Aquatic Bioassay consulting team conducted water quality surveys in the vicinity of the of the Goleta Sanitary Districts outfall on January 30th, April 8th, July 16th, and October 22nd, 2014. The team evaluated the local effect of the discharge within the immediate vicinity of the outfall terminus, and compared conditions there with those at control sites up-coast and down-coast of the outfall. During each field survey, the team recorded general observations of weather, etc., sampled for bacteria and water column variables (temperature, salinity, pH, transmissance and dissolved oxygen). On July 16th, the team deployed a series of caged mussel arrays for bioaccumulation analysis and on October 22nd, the team retrieved the mussels. On October 23rd, the team collected epibenthic fish and macroinvertebrates by otter trawl, and collected benthic sediments for physical, chemical, and infaunal analysis using a Van Veen Grab.

3.2. Station Locations and Descriptions

Water-column monitoring was conducted at ocean stations that are located at fixed distances from the midpoint of the diffuser (Figure 3-1). Stations B4 and B5 are located at the boundary of the zone of initial dilution (ZID), 25 meters (m) west and east of the diffuser, respectively. Station B2 and B3 are near-field stations located 500 and 250 m west of the diffuser, respectively. Station B1 is a far-field station located 1500 m west of the diffuser offshore Goleta Point. Station B6 is a reference station located 3000 m east of the diffuser. Plume stations WCZID and WC100 are respectively located 25 and 100 m away from the discharge in the direction of current flow. Nearshore Stations K1 through K5 are also at fixed distances west and east of the outfall in 20 m of water. Historically, the location of the 20 m depth contour represents the offshore limit of kelp beds in the study area.

Mussel arrays were deployed at Stations B3, B4, and B6. Trawl sampling was initiated at Stations B3 moving west for ten minutes and at Station B6 moving east for ten minutes (trawl stations TB3 and TB6, respectively).



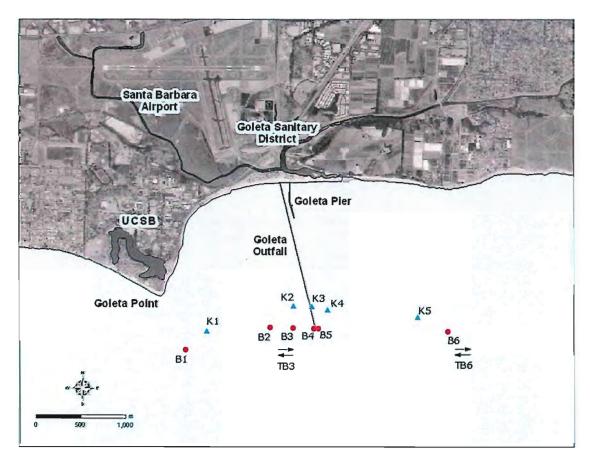


Figure 3-1. Goleta Sanitary District receiving water monitoring stations. Trawl stations are represented by arrows (--->).

3.3. Navigation and Positioning

The outfall diffuser and all sampling stations were located using a *Lowrance Global Map 2000* differential global positioning system (DGPS). DGPS positions were checked visually and by bottom-finder. Once the outfall terminus location was verified, a water quality analyzer cast was taken directly over the diffuser and water quality profiles were simultaneously downloaded to an onboard computer. Aquatic Bioassay biologists inspected the water quality traces for excursions from ambient such as higher temperature or lower salinity, dissolved oxygen, light transmissance, or pH. Any of these would reflect the presence of the wastewater plume. Once the plume was identified, a sail-drogue was deployed over the diffuser at the same depth as the discharge plume signature. The drogue was allowed to move with the current until an obvious direction and velocity could be determined. Stations WCZID (25 m from terminus) and WC100 (100 m from terminus) were then positioned along the drogue's line of travel.

3.4. Statistical Analysis

For this report, two types of statistical tests were performed; trend analysis using correlation coefficient analysis, and comparative analysis using t-tests and analysis of variance (ANOVA). For this report, statistical significance is highlighted at two



levels. For most ecologists, a pattern that is strong enough so that there is only a one chance or less in 20 that it is random is said to be statistically significant. In other words, the probability (p) is that there is only a 5% chance (0.05) or less that the pattern is random (p < 0.05). A pattern that has only one chance in ten or less (but more than one chance in 20) is said to be "marginally significant". That is, the probability is less than 10% but greater than 5% of being random (0.05 < p < 0.10).

3.5.1. <u>Correlation Coefficients.</u> Correlation analysis compares two variables to determine if they tend to increase or decrease in the same way. If two measurements tend to vary in opposite ways, their correlation coefficient (r-value) will tend to have a negative sign. If two measurements tend to vary in the same way, their r-value will tend to have a positive sign.

In addition to its sign, the size of an r-value is important. r-values range from – 1.000 to +1.000. An r-value of –1.000 means that the two measurements being compared vary exactly opposite from each other, an r-value of +1.000 means that the two measurements vary exactly in the same way, and an r-value of 0.000 means that the two measurements have no relationship to each other at all. Most r-values, however, fall somewhere among these three values. Depending upon the number of samples that are used to represent the true population, we have more confidence in our r-values when they are high. If an r-value is large enough so that the chance that the relationship could be random is only one in 20 or less (p \leq 0.05), we can have confidence that the relationship is probably real. We would have less confidence in a relationship between two variables if the probability was only one in ten (0.05 \leq 0.10) and no confidence if it was greater than ten (0.10 < p).

Based upon experience from past studies, we know that wastewater discharges can negatively impact the marine environment in very specific ways. If the outfall discharge is causing chemicals to accumulate in sediments and/or tissues, it follows that their concentrations would be higher nearer the diffuser than farther away. In this report, the distances of the stations from the diffuser were correlated against the concentration of the individual chemical components that were measured from these stations. Thus, the sign of the correlation coefficient between distance from outfall and chemical concentration would be *expected* if that chemical correlation was *negative*. That is, as the distance from the outfall becomes *larger*, the concentration of the compound becomes *smaller*. Another r-value that is expected to be negative is temperature. The effluent is always warmer than the ocean water, so temperatures, like chemicals, would be expected to become smaller with larger distances.

If the discharge were disrupting biological communities; abundance, diversity, etc., it would be expected to be lower near the outfall than farther away. Thus, population variables would be *expected* to correlate *positively* with distance from outfall, i.e. as distance becomes *larger* these variables would become *larger*. However, it is well documented that infauna populations can thrive near the nutrient enriching effects of ocean outfall where nutrients have enriched the area (Pearson and Rosenberg 1978). A positive and significant correlation between distance from the outfall abundance, numbers of species and diversity could signal that this is the case. Other r-values that are expected to be positive with distance are salinity, pH, dissolved oxygen, surface transparency, and light transmissance. This is because effluents are usually less saline, less clear, and lower in dissolved oxygen and pH than ocean water. If



the discharge were affecting the receiving waters, an increasing pattern of these variables with distance from outfall would be expected.

In conclusion, variables that vary in patterns that are both expected and significant should be those which bear further scrutiny.

3.5.2. <u>T-tests.</u> This statistic is used to compare variables when there are only two. Unlike correlation coefficients, the trend with distance is not evaluated. For most variables, the mean of values near the outfall and the mean of values farther away will be different. The t-test determines whether or not that difference is statistically significant. Note that trend with distance or sign of the statistic is not of importance for this test. The question asked is only if they are different beyond what might be expected of random chance.

T-tests are used in this report for trawled fish and invertebrate population metrics and chemical compounds in fish tissue, since these variables were replicated and collected at two locations (i.e. TB3 and TB6). If the average difference in concentration of a chemical compound between these two stations is large enough that the probability is less than or equal to 5% (p \leq 0.05), the difference is said to be statistically significant. If the difference is large enough so that the probability is less than or equal to 10% but greater than 5% (0.05 \leq 0.10), the difference is said to be marginally significant. If the concentration of the compound is larger at the near-outfall station, and the t-test is significant, the pattern should be further evaluated.

3.5.3. Analysis of Variance (ANOVA). ANOVA is similar to the t-test, except it can be used test for significant differences among more than two stations. ANOVAs were used for population variables and tissue analysis of bivalves. ANOVA analysis requires two steps. In the first step, differences in a variable among stations are evaluated to determine if they are sufficiently large to be statistically significant (p \leq 0.05). If they are, then a second test must be performed to determine which stations' variables are significantly larger than which other station or stations. In this report, this second step is called the comparison of means. For example, a comparison of means stating: B1 > B2, B3 > B4, indicates that, for that particular variable, Station B1 is significantly larger than Stations B2, B3, and B4, and Stations B2 and B3 are also significantly larger than Station B4. For chemical contaminants, if stations near the outfall are significantly higher than stations farther away, that compound should be evaluated further. For population variables, the opposite is true.

3.6. General Oceanographic Conditions

With the exception of somewhat sporadic freshwater runoff from non-point sources, the aquatic conditions in Goleta offshore area are controlled by the oceanographic conditions in the Southern California Bight. The mean circulation in the Southern California Bight is dominated by the northward-flowing Southern California Countercurrent, which may be considered as an eddy of the offshore, southward-flowing California Current (Daily, et. al. 1993). Nutrient rich, upwelled waters from the California Current can enter the western end of the Santa Barbara Channel promoting primary productivity (Dugdale and Wilkerson, 1989). The California Countercurrent transports nutrient poor, warmer water northward into the eastern Santa Barbara Channel (Hickey 1998). The California Countercurrent is seasonal in



nature and is usually well developed in the summer and fall and weak (or absent) in winter and spring (SCCWRP 1973). This causes relatively nutrient-poor waters to predominate in the warmer water months and nutrient rich waters to predominate in the colder water months (Soule, et. al. 1997).

Superimposed upon annual trends are the sporadic occurrences of the El Nino Southern Oscillation (ENSO) that can be described as an oceanographic anomaly whereby particularly warm, nutrient-poor water moves northward from the tropics and overwhelms the typical upwelling of colder nutrient-rich water. The El Nino Watch (http://coastwatch.pfel.noaa.gov/erddap/index.html) program continuously monitors global sea surface temperatures. These temperature data are compared to the long-term sea surface temperatures generated from data collected from 1950 to 2014. Comparison of the monthly sea temperature with this long term average creates a temperature anomaly so that the average monthly temperature falls either above or below the average. This anomaly allows us to determine how a given month or time period deviates from the long term ocean temperature trend.

Water temperatures offshore Goleta were at or above the long term average during each month in 2014 (Figure 3-2). The greatest excursions above the average temperature occurred in March, July, and October thru December. Water temperatures were closer to the long term average in each of the other months.

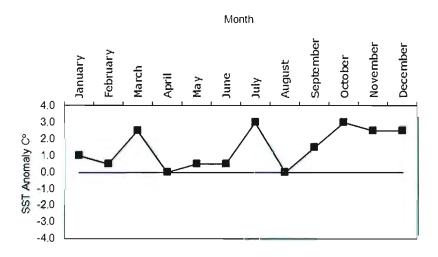


Figure 3-2. Sea surface anomaly temperatures for 2014 compared with long term trends.



3.7. Anthropogenic Inputs

In addition to the Goleta discharge, several other natural and anthropogenic sources could potentially impact the coastal area. Three marshes (Devereux Lagoon, Campus Lagoon, and Goleta Slough) and several creeks discharge into the local area. All are a potential source of contaminated water and sediments, coliform and enterococcus bacteria, and nutrients; particularly during the rainy season. Several sources of crude oil are also present. Natural seeps occur west of the diffuser in the vicinity of Coal Oil Point and Goleta Point, and offshore production activity occurs throughout the Santa Barbara Channel.

3.8. Rainfall

Total rainfall is not as important in terms of impacting an area as the timing of the rainfall, the amount in a given storm, and the duration of a storm (or consecutive storms). Relative to timing, the first major storm of the season will wash off the majority of the pollutants and nutrients accumulated on the land over the preceding dry period. An early, large, long duration storm would have the greatest impact on the waters. In addition, determining the impact of the rainfall and runoff is also a function of the timing of the sampling surveys. With a greater lag between runoff and survey sampling, mixing with oceanic waters would reduce observable impacts (Soule, et. al. 1996).

The rainfall reported in this document is for Santa Barbara Airport obtained from the Western Regional Climate Center in Reno, Nevada. Data is summarized in Table 3-2 and Figure 3-3, where periods of precipitation and water column survey days are highlighted. The rainfall for this period (11.51 inches) was 5.39 inches below the average yearly rainfall since 1981 (18.96 inches). The wettest month was December (4.95 in), followed by February (3.83 in). No rain fell in May and June. Rain in all other months ranged from 0.01 to 1.16 inches. The January and October sampling events were preceded by at least three days of dry weather, while both the April and August events had at least trace rain on or immediately prior to sampling.



Table 3-2. Daily 2014 Santa Barbara Airport rainfall (inches) with dates of water column surveys bordered and rain days in gray.

0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	Day/Month	January	February	March	April	May	June	AIN	August	September	October	November	December
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0		0.00	0.00	0.76	0.25	0.00	0.00	0.00	0.00	0.00	0.00	0.18	۲
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0		0.0	0.20	0.07	0.01	0.00	0.00	0.00	00.00	0.00	0.00	0.00	2.14
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0		0.00	_	0.00	0.00	0.00	0.00	0.00	F	۲	0.00	0.00	0.22
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0		0.00	0.00	0.00	۲	0.00	0.00	0.00	00.00	0.00	0.00	0.00	0.02
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0		0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	00.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0		0.00	0.22	0.00	0.00	0.00	0.00	0.00	00.00	0.00	0.00	0.00	00.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0		0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	m	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	00.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	•	0.0	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	. 0	0.00	0.00	0.00	0.00	0.00	0.00	0.00	00.00	0.00	0.00	0.00	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	· -	0.00	0.00	0.00	0.00	0.00	0.00	0.00	00.00	0.00	0.00	0.00	0.81
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	2	0.0	0.00	0.00	0.00	0.00	0.00	0.00	00.00	0.00	0.00	0.00	1.32
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	ဗ	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	4	0.00	0.00	0.00	0.00	0.00	0.00	F	0.00	0.00	0.00	0.00	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	2	0.00	0.00	0.00	0.00	0.00	0.00	0.00	00.00	0.00	0.00	0.00	0.24
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	91	0.00	0.00	0.00	0.00	0.00	0.00	0.01	00:00	0.00	0.00	0.00	0.05
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	7	0.00	0.00	0.00	0.00	0.00	0.00	0.00	00.00	0.00	0.00	0.00	0.13
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	8	0.00	0.00	0.00	0.00	0.00	0.00	0.00	00.00	0.00	0.00	0.00	0.02
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	19	0.00	0.00	0.00	0.00	0.00	0.00	F	00.00	0.00	0.00	۲	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	20	0.00	0.00	0.00	0.00	0.00	0.00	F	0.00	0.00	0.00	0.00	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	21	0.00	0.00	0.00	0.00	0.00	0.00	0.00	00.00	0.00	0.00	0.00	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	22	0.00	0.00	0.00	0.00	۲	0.00	0.00	0.00	0.00	0.00	0.00	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	23	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.01	0.00	0.00	0.00	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	54	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
0.00	55	0.00	00.00	00.00	0.01	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
0.00 0.81 0.00 0.00 0.00 0.00 0.00 0.00	97	0.01	1.00	۲	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	7.	0.00	0.81	0.00	0.00	0.00	0.00	0.00	00.00	0.00	0.00	0.00	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	83	0.00	1.60	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	59	0.00		0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
0.00 0.26 0.00 0.00 0.01 3.83 1.09 0.27 0.00 0.00	8	0.00		۲	0.00	0.00	۲	0.00	0.00	0.00	0.00	F	۲
0.01 3.83 1.09 0.27 0.00 0.00 0.01	31	0.00		0.26		0.00		0.00	0.00		1.16		0.00
	othly	0.01	3.83	1.09	0.27	0.00	0.00	0.01	0.01	0.00	1.16	0.18	4.95
)tal												
	I Total	11.51											

T = Trace, some precipitation fell but not enough to measure.



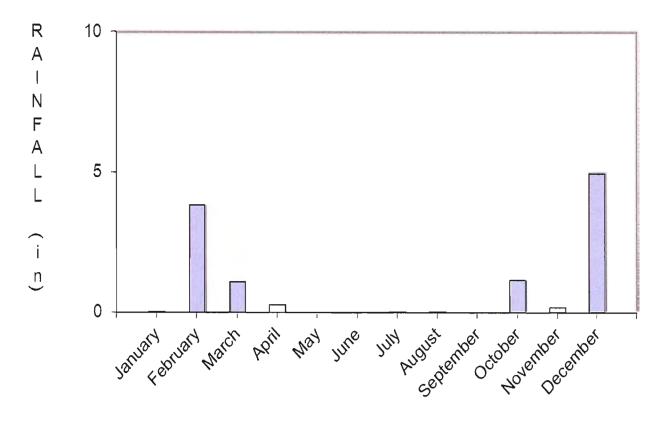


Figure 3-3. Santa Barbara rainfall for 2014.



3.9. Water Quality Materials and Methods

Sampling and data collection for water quality assessment was conducted quarterly at the 13 stations described above. Temperature, conductivity (later converted to salinity), dissolved oxygen, pH, and light transmissance were measured continuously through the water column using a SeaBird 25plus CTD Water Quality Analyzer with associated WetLabs 25-cm Transmissometer. All probes were calibrated immediately prior to each field excursion and, if any data were questionable, they were calibrated again immediately after the instruments were returned to the laboratory. Measurements of light penetration were measured using a Secchi disk. At all stations, water samples were collected at the surface, at mid-depth, and above the bottom with a Nauman sampler. In January, 2014 the pH sensor failed making measurements presented herein questionable.

Water was distributed into sterile 125 mL polypropylene bottles for bacterial analysis. At all stations, temperature and pH were measured directly at the surface using an NBS traceable standard mercury thermometer and hand-held, buffer-calibrated pH meter (respectively). Extra water samples were also collected and set for dissolved oxygen and chloride titration in the field. These extra samples and measurements were used as a check and back up to the water quality analyzer.

All samples from all stations were placed in coolers containing wet ice and were returned to the Ventura laboratory the same day. Immediately upon return, the bacterial samples were set for total and fecal coliform and enterococcus bacteria via multiple-tube fermentation methods. Check samples were titrated for dissolved oxygen by Winkler titration and chloride (converted to salinity) by the argentometric titration. All water analyses were performed in accordance with *Standard Methods for the Examination of Water and Wastewater* (American Public Health Association, 22nd Edition).

After all analyses were completed, the five water quality analyzer variables were correlated against the check samples measured or collected in the field: thermistor probe versus mercury thermometer, conductivity probe versus chloride titration, dissolved oxygen probe versus Winkler titration, field pH probe versus hand-held pH meter, and transmissometer versus Secchi disk (see Appendix Figure 10-1 for calibration curves). The Seabird Water Quality Analyzer was downloaded and water column graphs were generated. Two tables were also prepared containing the results of the physical, chemical, bacterial, and observational water measurements. Check sample correlations, water column graphs, and data tables were joined with a narrative report and were presented to the Water Quality Control Board quarterly. The results and conclusions of all water column measurements and analyses are presented and summarized in Section 3.10 below.



3.10. Results

3.10.1. Physical and Chemical Water Quality

3.10.1.1. Temperature

Coastal water temperatures vary considerably more than those of the open ocean. This is due to the relative shallowness of the water, inflow of freshwaters from the land, and upwelling. Seawater density is important in that it is a major factor in the stratification of waters. The transition between two layers of varying density is often distinct; the upper layer, in which most wind-induced mixing takes place, extends to a depth of 10 to 50 m in southern California waters.

During the winter months, there is little difference in temperature between surface and deeper waters, while in the summer a relatively strong stratification (i.e. thermocline) is evident because the upper layers become more heated than those near the bottom do. Thus, despite little difference in salinity between surface and bottom, changes in temperature during the summer result in a significant reduction of density at the surface. Stratified water allows for less vertical mixing. This is important because bottom waters may become lower in oxygen without significant replenishment from the surface (Soule et. al. 1997).

Spatial temperature patterns. Examination of 3D contours for each quarterly survey showed that the water column was weakly isothermal in January. April had the coolest water of the any survey (10.8 °C) indicating an upwelling event. In July temperatures warmed at the surface and a moderately strong thermal gradient was established. In October the water column temperatures were warmest of the four surveys and was isothermal at the nearshore stations and stratified at the offshore stations due to the cooler deep water there (Figure 3-6 and Table 3-3). In January, water temperatures essentially the same through the water column (15.2 to 15.7 °C). The April survey had water temperatures that declined with depth, ranging from 13.4 °C near the surface to 10.8 °C at the bottom. Thermal stratification was strongest in July and October when water temperatures were ranged from 15.1 near the bottom to 21.5 °C near the surface, representing a 6.4 °C decrease from surface to bottom.

Influences of the outfall were not evident in the temperature profiles during any survey. Temperatures did not correlate with distance from the outfall in any survey. There were no significant temperature differences by t-test between near outfall and far field station groups during the four quarterly surveys.



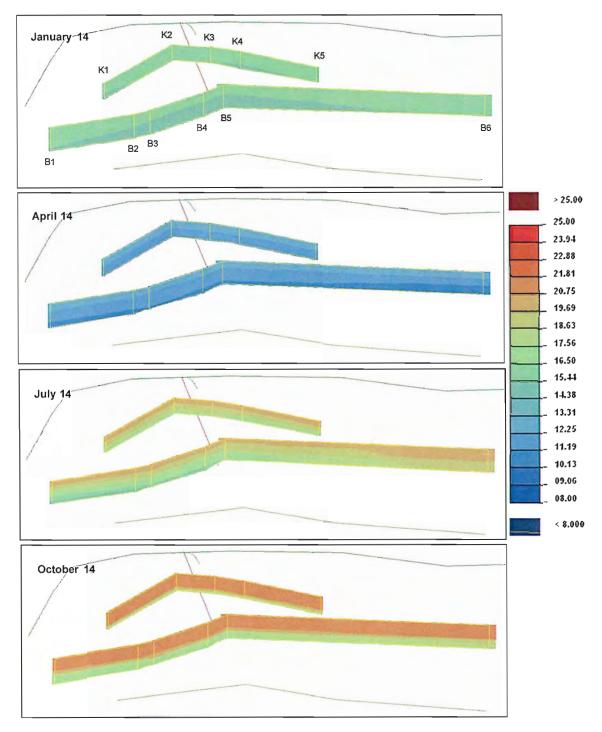


Figure 3-6. Temperature contours for the K Station (depth = 18 m) and B Station (depth = 28 m) water quality transects. The Goleta Sanitary District outfall is depicted as a red line. The color legend is presented to the right.



Table 3-3. Water quality parameter averages and ranges for all stations and depths combined for each quarterly survey. The statistical significance of quarterly measurements with distance from the outfall was tested by correlation analysis and by t-test.

Parameter	Month	Average	Range	Expected & Significant Correlation w/ Outfall?	Significant t- test w/ Outfall?
Temperature	January	15.5	15.2 - 15.7	No	No
	April	12.0	10.8 - 13.4	No	No
	July	18.6	15.1 - 20.4	No	No
	October	20.1	16.2 - 21.5	No	No
Salinity	January	33.5	33.4 - 33.5	No	No
	April	33.6	33.5 - 33.6	Yes	No
	Jul y	33.5	33.4 - 33.6	Yes	No
	October	33.5	33.2 - 33.6	No	No
рН	January	8.1	8.1 - 8.2	No	No
	April	8.1	7.8 - 8.4	No	No
	July	8.2	8.1 - 8.3	Yes	No
	October	8.3	8.2 - 8.3	No	No
DO	January	7.9	7.5 - 8.1	No	No
	April	9.1	4.4 - 12.5	No	No
	July	7.8	7.2 - 8.1	No	No
	October	7.3	6.9 - 8.0	No	No
Transmissance	January	78.6	56.2 - 88.0	No	No
	April	68.3	61.4 - 78.0	Yes	No
	July	82.7	71.9 - 87.4	Yes	No
	October	83.1	51.7 - 90.9	No	No
Transparency	January	12.4	7.0 - 17.3	Yes	No
	April	5.7	5.1 - 6.3	No	No
	July	11.3	10.0 - 12.3	Yes	Yes
	October	15.1	11.0 - 18.0	No	No



3.10.1.2. Salinity

Salinity (a measure of the concentration of dissolved salts in seawater) is relatively constant throughout the open ocean; however, it can vary in coastal waters primarily because of the inputs of freshwater from the land or because of upwelling. In a five-year study conducted by the U.S. Navy Research and Development Center, more than 1000 samples were analyzed for salinity. The mean salinity was 33.75 parts per thousand (ppt), and the range of 90% of the samples in southern California fell between 33.57 and 33.92 ppt (SCCWRP 1973).

Despite the general lack of variability, salinity concentrations can be affected by a number of oceanographic factors. During spring and early summer months, northwest winds are strongest and drive surface waters offshore. Deeper waters, which are colder, more nutrient-rich, and more saline, are brought to the surface to replace water driven offshore (Emery 1960). El Nino (ENSO) events can also affect coastal salinities. During these events northern flowing waters move into the Bight with waters that are also more saline, but are warmer and lower in nutrients than ambient water. Major seasonal currents (i.e. California current, countercurrent, or undercurrent) can also affect ambient salinity to some degree (Soule et. al. 1997).

Spatial salinity patterns. Average salinity in the survey area was nearly identical across the four surveys ranging from 33.2 ppt in October to 33.6 ppt in each of the other surveys. However, salinity provided the best opportunity to detect the effluent plume which is evident in the January, April and July surveys. In January, the water column was isohaline (same surface to bottom), except for a small patch of fresher water at the surface at station B4 near the outfall. In April, lower salinity water is seen as a surface and subsurface lens of slightly fresher water to the north of the outfall. In July, fresher water can be seen toward the surface north of the outfall. In October, when the water column was most strongly thermally stratified, salinity was also stratified with no clear indication of the plume.

<u>Salinity ranges and outfall effects.</u> Table 3-3 shows the range of salinities for the 11 water column stations over the four quarterly sampling surveys. Salinities correlated with distance from the outfall in April and July, but there were no significant salinity differences by t-test between near outfall and far field station groups for any of the four quarters.



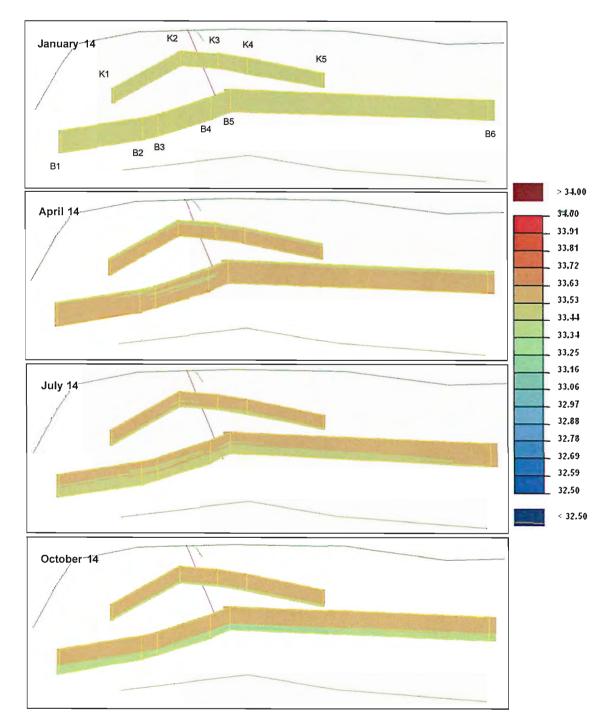


Figure 3-7. Salinity (ppt) contours for the K Station (depth = 18 m) and B Station (depth = 28 m) transects. The Goleta Sanitary District outfall is depicted as a red line.



3.10.1.3. Hydrogen Ion Concentration (pH)

pH is defined as the negative logarithm of the hydrogen ion concentration. A pH of 7.0 is neutral, values below 7.0 are acidic, and those above 7.0 are basic (Horne 1969). Seawater in southern California is slightly basic, ranging between 7.5 and 8.6, although values in shallow open-ocean water are usually between 8.0 and 8.2 (SWQCB 1965). These narrow ranges are due to the strong buffering capacity of seawater, which rarely allows for extremes in pH.

Factors that can influence pH in the ocean are freshwater inputs, upwelling, and biological activity. Since freshwater pH values tend to be about 0.5 pH units less than seawater, any inflow from a freshwater source will tend to lower the pH slightly. When photosynthesis is greater than respiration, more carbon dioxide is taken up than generated, and pH may increase to higher values in the euphotic (i.e. light penetrating) zone. When respiration is greater than photosynthesis, more carbon dioxide is released than used and pH may decrease, especially when mixing is minimal such as in the oxygen minimum zone and towards the bottom (Soule et. al. 1997).

<u>Spatial pH patterns.</u> The pH sensor failed in January, 2014 and while data are presented, they are not interpreted. Average pH across the three other quarterly surveys ranged from 7.8 to 8.4 (Figure 3-8 and Table 3-3). Low pH water was evident near the bottom in April indicating an upwelling event.

In April pH was stratified through the water column and was least near the bottom (7.9) and greatest near the surface (8.4). In July and October, pH was similar through the water column. There was no clear evidence of the effluent plume from the contours during any of the four surveys.

pH ranges and outfall effects. Table 3-3 shows the range of pH values for 11 water column stations for each of the four quarterly sampling surveys. There were no expected and significant correlations with distance to the outfall for any survey, except in July. However, the greatest average difference in pH between the ZID stations (WCZID and WC100) and those further away was 0.03 pH units. Also, there were no significant differences in pH among station groups located near and far from the outfall by t-test for any survey. Analysis of each quarterly data set showed that all pH differences between stations near and away from the outfall were very low and well within the 0.2 pH unit limit specified in the California Ocean Plan (2009).



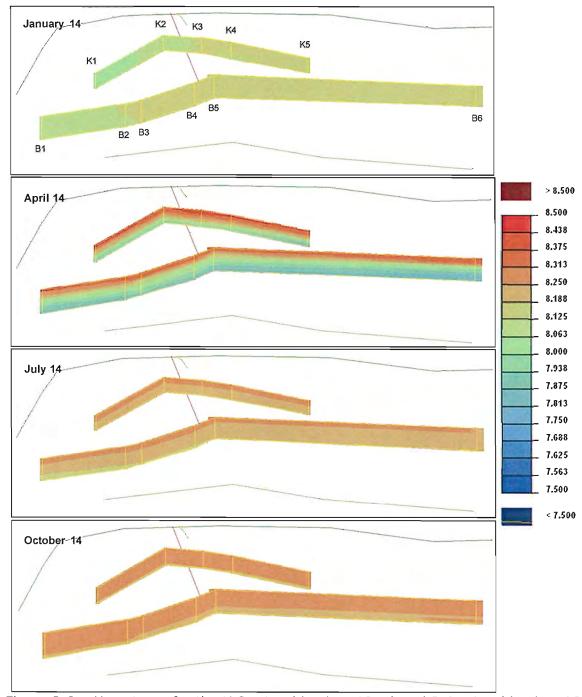


Figure 3-8. pH contours for the K Station (depth = 18 m) and B Station (depth = 28 m) transects. The Goleta Sanitation District outfall is depicted as a red line.



3.10.1.4. Dissolved Oxygen

The most abundant gases in the ocean are oxygen, nitrogen, and carbon dioxide. These gases are dissolved in seawater and are not in chemical combination with any of the materials composing seawater. Gases are dissolved from the atmosphere by exchange across the sea surface. The gases dissolved at the sea surface are distributed by mixing, advection (i.e. from currents), and diffusion. Concentrations are modified further by biological activity, particularly by plants and certain bacteria. In nature, gases dissolve in water until saturation is reached given sufficient time and mixing. The volume of gas that saturates a given volume of seawater is different for each gas and depends upon temperature, pressure, and salinity. An increase in pressure, or a decrease in salinity or temperature, causes an increase in gas solubility.

The amount of oxygen dissolved in the sea varies from zero to about 11 milligrams per liter. At the surface of the sea, the water is more or less saturated with oxygen because of the exchange across the surface and plant activity. In fact, when photosynthesis is at a maximum during a phytoplankton bloom, such as during a red tide event, it can become supersaturated (Anikouchine and Sternberg 1973). When these blooms die off, bacterial aerobic respiration during decomposition of these phytoplankton cells can rapidly reduce dissolved oxygen in the water. Dissolved oxygen typically decreases with depth due to respiration associated with the bacterial breakdown of organic material. However, if the water column is well mixed, oxygen will be fairly constant with depth. Temperature and/or salinity can affect the density structure of the water column and create barriers to vertical mixing.

<u>Spatial oxygen patterns.</u> During the January, July and October surveys, dissolved oxygen concentrations were similar through the water column and ranged from 6.9 to 8.1 mg/L (Figure 3-9 and Table 3-3). In April the water column was stratified for oxygen and ranged from 4.4 mg/L near the bottom to 12.5 mg/L near the surface. The presence of oxygen depleted deep water was further evidence of an upwelling event.

Oxygen ranges and outfall effects. Table 3-3 shows the range of oxygen concentrations for the 11 water column stations over the four quarterly sampling surveys. Dissolved oxygen did not correlate significantly with distance to the outfall for any of the four surveys and there were no significant differences by t-test among sites located near the outfall and those further away. This indicates that dissolved oxygen was not influenced by the outfall diffuser. Dissolved oxygen concentrations between stations located near and away from the outfall remained within the Ocean Plan standards (2009) throughout the year, except in April when dissolved oxygen was lower at the plume stations (7.76 mg/L) compared to sites further away where, on average, there was up to a 25% increase in dissolved oxygen. It is most likely that the depressed oxygen near the outfall was the result of the freshwater plume entraining upwelled low oxygen water and bringing it towards the surface.



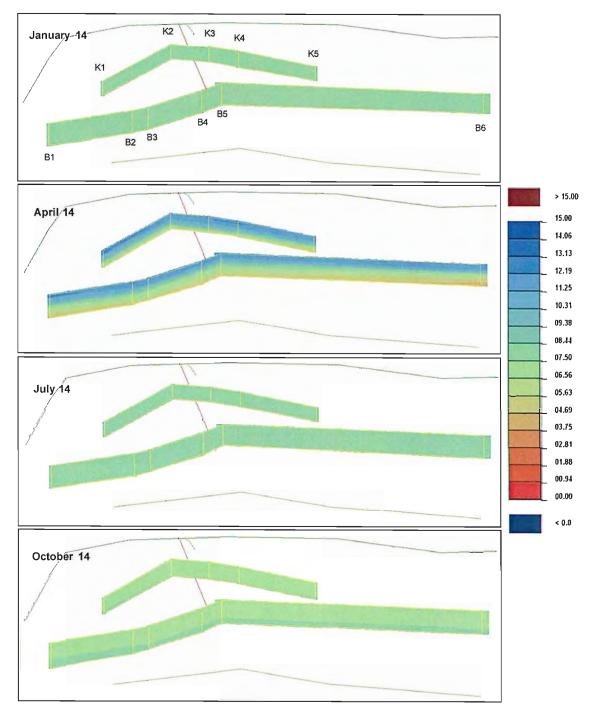


Figure 3-9. Dissolved oxygen contours for the K Station (depth = 18 m) and B Station (depth = 28 m) transects. The Goleta Sanitary District outfall is depicted as a red line.



3.10.1.5. Light Transmissance

Water clarity in the ocean is important both for aesthetic and ecological reasons. Phytoplankton, as well as multicellular marine algae and flowering plants are dependent upon light for photosynthesis and therefore growth. Since nearly all higher-level organisms are dependent upon plants for survival (except those animals living in deep-ocean volcanic vents and similar environments), the ability of light to penetrate into the ocean depths is of great importance. Seasonally, water is usually least clear during spring upwelling and winter rain. In early summer, increased day length can promote plankton growth and reduce water clarity, as well. In late summer and fall, days are shorter and the rains that bring sediments into the marine environment have yet to begin. Therefore, late summer and early fall are typically the periods of greatest water clarity. Anthropogenic influences such as wastewater effluents, storm drainage discharges, and non-point runoff can also influence water quality on a local basis.

Water clarity is determined using two completely different measuring techniques. Surface transparency is measured using a weighted, white plastic, 30 cm diameter disk (called a Secchi Disk) attached to a marked line. The disk is simply lowered through the water column until it disappears, and the depth of its disappearance is recorded. Surface transparency is a good estimate of the amount of ambient light that is available to plankton since the depth to which light is available for photosynthesis is generally considered to be about 2.5 times the Secchi disk depth.

Light transmissance is measured using a transmissometer, which is a 0.25 m open tube with an electrical light source at one end and a sensor at the other. The amount of light that the sensor receives is directly dependent upon clarity of the water between them. Results are recorded as percent light transmissance. Since transmissance is independent of ambient sunlight, it can be used at any depth and under any weather conditions. Surface light transmissance is usually positively correlated with surface transparency.

Spatial transmissance patterns. Water clarity was good throughout the water column during each of the four quarterly surveys (Figure 3-10). Average transmissance across the four surveys ranged from 51.7% in October to 90.9% in October (Table 3-3). In January and July there was a layer of lower transmissance water near the bottom and into the mid-water column, especially in July. Water clarity was uniformly lower in April, possibly due to phytoplankton growth after the influx of nutrient rich, upwelled water. In October, water clarity was the best of the entire year, except near the bottom of the offshore transect where the lowest water quality of the year was measured.

<u>Transmissance ranges and outfall effects.</u> Table 3-3 shows the range of transmissance for the 11 water column stations over the four sampling surveys. Comparisons among stations showed there was a significant correlation with distance to the outfall in April and July, but there was no significant difference among near and far field stations by t-test during any of the four surveys. In all cases, there was never a reduction in transmissance between near and far field stations that exceeded the Ocean Plan (2009) standard of 10%.



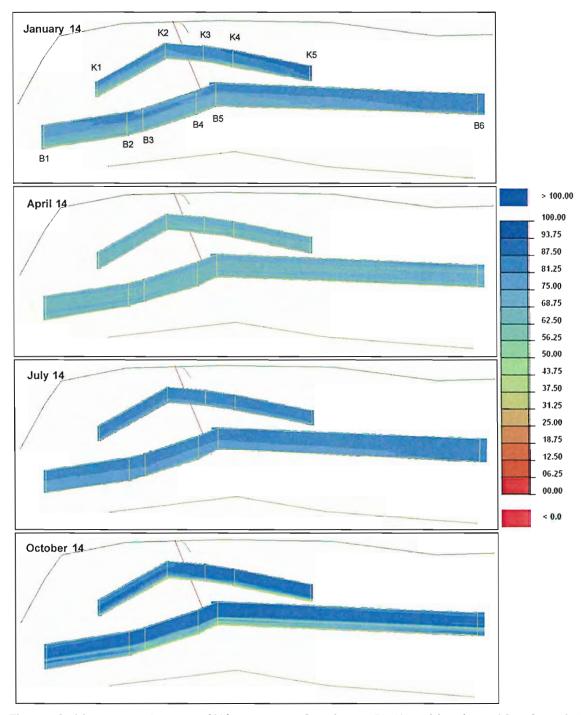


Figure 3-10. Transmissance (%) contours for the K Station (depth = $18\,$ m) and B Station (depth = $28\,$ m) transects. The Goleta Sanitary District outfall is depicted as a red line.



3.10.1.6. Surface Transparency

As discussed in more detail in Section 3.10.1.5 above, surface transparency is recorded as the depth (m) at which a weighted, 30 cm, white plastic disk (Secchi Disk) disappears from view. Since only a single quarterly measurement is taken at each station, these data are presented as a line plot of transparency vs. quarter.

<u>Transparency patterns and outfall effects.</u> Figure 3-11 shows the range of transparency measurements for the 11 water column stations over the four sampling surveys. Average surface transparency ranged from 5.7 m in April to 15.1 m in October. The lowest transparencies of the year occurred in April during the upwelling event. Transparency correlated significantly with distance from the outfall in January and July, and significantly by t-test among stations located near to the outfall compared to stations further away in July.

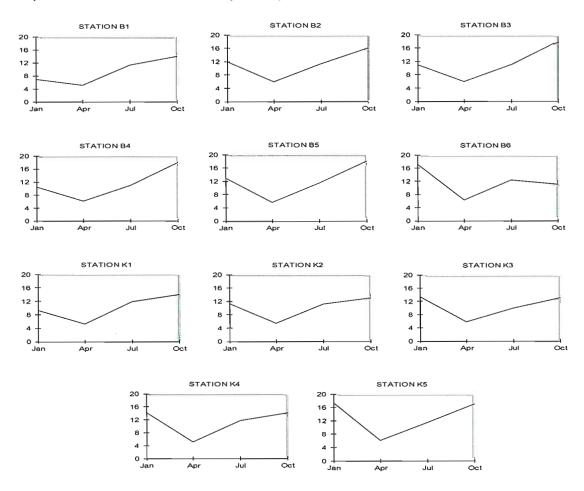


Figure 3-11. Average transparency vs. season for each of the 11 water quality stations.



3.10.2. Bacterial Water Quality

The three bacterial measurements of total coliforms, fecal coliforms and enterococcus, are used by health authorities to assess the potential risk of human exposure to pathogens in the aquatic environment (Soule 1997). The principle problem with these indicators is that analysis takes 72 hours, slowing the response of health officials to potentially hazardous conditions. Research has been underway to develop more rapid tests that are both sensitive and cost effective. Rainfall episodes have been closely associated with violations of all three bacterial standards, especially near areas where creeks or stormwater channels discharge into the ocean. At present, it is more prudent to post areas of potential or known contamination immediately following rain storm events than to wait for confirmation. Bacterial results are summarized in Tables 3-4 and 3-5.

3.10.2.1. Total Coliforms

Coliform bacteria (those inhabiting the colon) have been used for many years as indicators of fecal contamination; they were initially thought to be harmless indicators of pathogens at a time when waterborne diseases such as typhoid fever, dysentery and cholera were severe problems. Recently it was recognized that coliforms themselves might cause infections and diarrhea. However, the total coliform test is not effective in identifying human contamination because these bacteria may also occur as free living in soils, and are present in most vertebrate fecal material. The California Ocean Plan (SWRCB 2009) states that within 1,000 feet of shore, the single sample total coliform concentration cannot exceed 10,000 MPN/100 mL of water. Additionally, during a 30-day period the average concentrations cannot exceed 1,000 MPN/100 mL. Although no offshore stations are within 1000 feet of shore, this value was used as a criterion of concern.

Total coliform patterns over the year. Total coliform counts were very low during the year, ranging from <2 to 20 MPN/100 mL for all surveys (Table 3-4). In general values were very low throughout the year at all stations and depths with most samples below detection (<2 MPN/100 mL). These total coliform concentrations were far below either the single sample Ocean Plan standard (2009) of 10,000 MPN/100 mL or the monthly average total coliform standard of 1,000 MPN/100mL (Table 3-5).



Table 3-4. Annual summary of total and fecal coliforms and enterococcus bacteria (MPN/100 mL).

Sampling Station	Season	ВІ	B2	Off B3	shore B4	B5	B6	Plu WCZID 1	ıme WC100	K1	K2	Nearsh K3	nore K4	K5
SURFACE														
Total Coliform	Winter	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Spring	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Summer	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Fall	<2	<2	<2	<2	<2	<2	<2	<2 <2	<2	<2	<2	<2	<2
Fecal Coliform	Winter	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Spring	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Summer	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Fall	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
Enterococcus	Winter	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Spring	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Summer	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Fall	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
MIDDLE														
Total Coliform	Winter	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Spring	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Summer	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Fall	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
Fecal Coliform	Winter	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Spring	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Summer	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Fall	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
Enterococcus	Winter	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Spring	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Summer	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Fall	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
воттом														
Total Coliform	Winter	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Spring	<2	<2	<2	<2	<2	20	<2	<2	<2	<2	<2	<2	<2
	Summer	<2	<2	<2	<2	<2	<2	50	<2	<2	<2	<2	<2	<2
	Fall	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
Fecal Coliform	Winter	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Spring	<2	<2	<2	<2	<2	20	<2	<2	<2	<2	<2	<2	<2
	Summer	<2	<2	<2	<2	<2	<2	20	<2	<2	<2	<2	<2	<2
	Fall	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
Enterococcus	Winter	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Spring	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
II .	Summer	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
	Fall	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2



Table 3-5. Indicator bacteria geometric averages and ranges for all stations and depths combined for each quarterly survey. Measurements for the year were compared individually against single sample event, REC-1 bathing water standards.

Parameter	Month	Average	Range	Water Quality Standard	Standard Exceedances
Total Coliform	January	2	<2 - 2	10,000	0
	April	3	<2 - 20	10,000	0
	July	2	<2 - 2	10,000	0
	October	2	<2 - 2	10,000	0
Fecal Coliform	January	2	<2 - 2	400	0
	April	3	<2 - 20	400	0
	July	2	<2 - 2	400	0
	October	2	<2 - 2	400	0
Enterococcus	January	2	<2 - 2	104	0
	April	2	<2 - 2	104	0
	July	2	<2 - 2	104	0
	October	2	<2 - 2	104	0

3.10.2.2. Fecal Coliforms

The fecal coliform test discriminates primarily between soil bacteria and those in warm blooded animals such as dogs, cats, birds, horses, barnyard animals, and humans. The California Ocean Plan (SWRCB 2009) states that within 1000 feet of shore, samples from each station shall have a density of fecal coliform organisms less than 400 MPN/100 mL of water for any single sample or average less than 200 for any 30 day period. Although no offshore stations are within 1000 feet of shore, this value was used as a criterion of concern.

Fecal coliform patterns over the year. Fecal coliform counts were very low during the year, ranging from <2 to 20 MPN/100 mL for all surveys (Table 3-4). In general values were very low throughout the year at all stations and depths with most samples below detection (<2 MPN/100 mL). These fecal coliform concentrations were far below either the single sample Ocean Plan standard (2009) of 400 MPN/100 mL or the monthly average fecal coliform standard of 200 MPN/100mL (Table 3-5).



3.10.2.3. Enterococcus

Enterococcus bacteria include species that are found in human wastes and are related to the Streptococcus bacteria. At one time they were believed to be exclusive to humans, but other Streptococcus species occur in feces of cows, horses, chickens, and other birds. Enterococci die off rapidly in the environment, making them indicators of fresh contamination, but not exclusively from humans. The California Ocean Plan (SWRCB 2009) limitations within 1000 feet of shore are a 30 day average of 34 MPN/100 mL and a single sample limit of 104 MPN/100 mL.

<u>Enterococcus bacteria patterns over the year.</u> Enterococcus bacteria counts ranged from the method detection limit (<2 MPN/100 mL) to just above it (2 MPN/100 mL) during each survey (Table 3-4). Enterococcus concentrations at all stations and depths in the survey area were below the single sample Ocean Plan standard (2009) of 104 MPN/100 mL (Table 3-5).



3.11. Discussion

Quarterly water quality surveys were conducted offshore Goleta in January, April, July and November 2014. Measurements for temperature, salinity, pH, dissolved oxygen and water clarity showed that oceanographic conditions during the year were typical of nearshore areas in southern California. Rainfall for this period (11.51 inches) was 5.39 inches less than the average yearly rainfall since 1981 (18.96 inches). This lack of rainfall meant less nearshore surface runoff and may have led to the good water clarity and low bacteria counts throughout the year.

Salinity provided the best opportunity to detect the effluent plume which is evident in the January, April and July surveys. In January, the water column was isohaline (same surface to bottom), except for a small patch of fresher water at the surface at station B4 near the outfall. In April, lower salinity water is seen as a surface and subsurface lens of slightly fresher water to the north of the outfall. In July, fresher water can be seen toward the surface north of the outfall. None of the other parameters showed evidence of the effluent plume.

Physical and chemical characteristic restrictions, which apply to waters outside of the zone of initial dilution, are addressed in the California Ocean Plan (2009):

- The pH shall not be changed at any time more than 0.2 units from that which occurs naturally.
- The dissolved oxygen concentration shall not at any time be depressed more than 10 percent from that which occurs naturally, as the result of the discharge of oxygen demanding waste materials.
- Natural light shall not be significantly reduced at any point outside of the zone of initial dilution.
- Floating particulates and grease and oil shall not be visible.
- The discharge of waste shall not cause aesthetically undesirable discoloration of the ocean surface.
- Waste discharged to the ocean must be essentially free of: 1) Material that is floatable or will become floatable upon discharge.
- The waste discharged to the ocean must be essentially free of: 4) Substances that significantly decrease the natural light to benthic communities and other marine life.
- Waste discharged to the ocean must be essentially free of: 5) Materials that result in aesthetically undesirable discoloration of the ocean.

The water quality parameters measured during the four quarterly surveys indicated that the outfall plume was not altering the condition of the water mass in the vicinity of the Goleta outfall. None of the above restrictions were exceeded outside the zone of initial dilution, except in April when dissolved oxygen concentrations were up to 25% less near the outfall compared to stations further away. This discrepancy was most likely strongly influenced by the upwelling event that was occurring during this



survey. This was evidenced by the coolest water of the year ($10.8\,^{\circ}$ C), low pH near the bottom (7.96 pH units) and decreased dissolved oxygen near the bottom (4.4 mg/L), all characteristics of the intrusion of deep upwelled water. Decreases in dissolved oxygen at the outfall stations were probably due to the entrainment of oxygen poor upwelled water being brought to the surface by the buoyant freshwater plume, and not the result of decreased oxygen due to the effluent. pH and transmissance were within Ocean Plan (2009) standards during each of the four quarterly surveys.

Water color throughout the area was green, and the discharge of oil or floating particulates were never observed in the survey area. Water quality measurements taken near to and far from the outfall terminus correlated expectedly and significantly with distance from the outfall in a few instances including salinity in April and July, pH in July, transmissance in April and July and transparency in January and July. Of these, only transparency was significantly different among sites close to and far from the outfall. While statistically significant, these differences were small and not ecologically significant:

In July transparency differences from near field to far field stations were exceedingly small, with the average difference among the plume stations (11.0 m) and the station with the greatest transparency (far field station B1, 11.3 m) a 0.3 m difference.

Bacteriological standards are addressed in the Ocean Plan and NPDES discharge permit, however these standards relate primarily to shoreline waters used for recreation or shellfish harvesting (REC-1 bathing water standards). Total coliforms, fecal coliforms and enterococcus indicator bacteria concentrations were very low throughout the year in the Goleta survey area. A total of 156 samples were collected and analyzed for each indicator. None of these exceeded the single sample Ocean Plan standard (2009) during the year and over 95% of the measurements were below detection limits.

In conclusion, evidence from the four quarterly water column monitoring surveys conducted in 2014 indicate that the Goleta Sanitary District Wastewater Treatment Plant was in compliance with all water quality standards, and that the treatment plant was operating effectively.



CHAPTER 4

Physical Characteristics of the Benthic Sediments

4.1. Background

Marine sediments provide clues to the nature of the environment from which their constituent materials were derived, the transportation processes by which they arrived at the final site of deposition, and the physico-chemical and biological characteristics of the depositional environment. The Southern California Bight coastal shelf is characterized by sediments composed of varying combinations of sand, silt and clay. This is quite different in character from more northerly coastal reaches that are composed of rocky substrates. The distribution of benthic sediments can have a profound affect upon the diversity, abundance, and community structure of infaunal organisms and the accumulation of organic material and anthropogenic contaminants (Gray 1981). In general, finer sediments provide a more stable environment for benthic organisms, especially those that build tubes, burrow and feed there. Finer sediments, however, also tend to adsorb more organic and elemental contaminants than do coarser, sandier sediments. As a result, organisms that live closely associated with fine sediments can be exposed to higher concentrations of contaminants.

4.2. Materials and Methods

Benthic grab sampling was conducted in accordance with *Techniques for Sampling and Analyzing the Marine Macrobenthos* March 1978, EPA 600/3-78-030; *Quality Assurance and Quality Control (QA/QC) for 301 (h) Monitoring Programs: Guidance on Field and Laboratory Methods* May 1986, Tetra Tech; *The Southern California Bight Pilot Project Field Operations Manual* (SCCWRP 2008).

Samples were collected with a chain-rigged, tenth square-meter Van Veen Grab. At each station, the grab was lowered rapidly through the water column until near bottom, and then slowly lowered until contact was made. The grab was then slowly raised until clear of the bottom. Once on board, the grab was drained and initial qualitative observations of color, odor, consistency, etc. were recorded.

Sediments to be analyzed for physical properties were removed from the top 2 cm of the surface and placed in clean plastic Whirl-Pacs. These were analyzed for particle size distribution using a Horiba LA920 Particle Size Analyzer and in accordance with Standard Methods 2560 D (APHA, 2012). Sub-samples from each sediment sample were re-suspended in de-ionized water, and then injected into the analyzer. The analyzer is capable of measuring particle sizes ranging from silt and clay (<2 μ m) up to course sand (2,000 μ m). Results were recorded as the percentage each size distribution represented of the whole. When the LA920 detected particles in a sample that neared its upper detection limit (2,000 μ m), a portion of the sample was dried at 105 °C, weighed, then sieved through a 2,000 μ m mesh screen. Particles not passing through the screen were weighed and expressed as the percentage of particles in the sample >2,000 μ m (gravel).

Data for each station were reduced to the median particle size (μ m), percent fines and, the sorting index. The sorting index values range between sediments that have a very narrow distribution (very well sorted) to those which have a very wide



distribution (extremely poorly sorted). This index is simply calculated as the 84^{th} percentile minus the 16^{th} percentile divided by two (Gray 1981). Well sorted sediments are homogeneous and are typical of high wave and current activity (high energy areas), whereas poorly sorted sediments are heterogeneous and are typical of low wave and current activity (low energy areas).

4.3. Results

4.3.1. Station Event and SeaState Conditions

Sediment sampling, trawling and mussel retrieval was conducted on October 23rd,2014under clear skies, and calm to moderate conditions (Table 4-1). Wave height was three feet from the southwest and winds were four to six knots.

4.3.2. Particle Size Distribution

Tables 4-2 and 4-3, and Figure 4-1 illustrate the overall particle size distributions from the six sediment-sampling stations. Detailed raw and summary data for particle size are presented in Appendix 10.3. Results are presented for each size range as the percent of the whole. Two sediment characteristics can be inferred from the graphs. Position of the midpoint of the curve will tend to be associated with the median particle size (Figure 4-1). If the midpoint tends to be toward the larger micron sizes, then it can be assumed that the sediments will tend to be coarser overall. If the midpoint is near the smaller micron sizes, then it can be assumed that the sediments are mostly finer. Sediment sizes that range from 2000 to 63 µm are defined as sand, sediments ranging from 63 to 4 µm are defined as silt, and sediments that are 4 µm or less are defined as very fine silt and clay (Wentworth Sediment Scale, see Gray 1981). There are also subdivisions within the categories (e.g. very fine sand, etc., see Table 4-3). A second pattern discernible from the graph is how homogeneous the distributions of sediments are. Sediments that tend to have a narrow range of sizes are considered homogeneous or well sorted. Others, which have a wide range of sizes, are considered to be heterogeneous or poorly sorted.

4.3.2.1. General Description

At total of 36 replicate samples were successfully collected at the six sampling sites for all biological and chemical analyses (Table 4-2). The penetration depth of each grab exceeded the 5 cm minimum depth required by the Southern California Bight protocol. Surface sediments had the same descriptions at all stations. Surface sediments were composed of fine sand, the color was olive green and there was no odor, except at station B1 where the smell of hydrogen was detected in one replicate.

4.3.2.2. Median Particle Size

Median particle sizes are depicted in Table 4-3. Similar to past years, median particle sizes were categorized as very fine sand. Median particle sizes ranged from 91 to $115\mu m$.



4.3.2.3. Sorting Index & Percent Fines

Particles at all stations were poorly sorted and sorting indexes ranged from 1.16at station B6to 1.88 at B1 (Table 4-3). The percent fine sediments ranged from 20%at stations B5 and B6 to 25% at stationsB1 and B2.

4.4. Discussion

Observational and analytical evaluations of the benthos in the vicinity of the Goleta outfall show that the sediments are heterogeneous and composed of very fine sand. The percentage of fine sediments (silt and clay) ranged from 20% to 25% at each of the stations, which was in keeping with results from previous years. Hydrogen sulfide gas is a byproduct of bacterial decomposition of organic material under anoxic conditions. In 2014 the smell of hydrogen sulfide was present in one replicate at station B1.

There were no apparent differences in particle size between the outfall stations and those further away. Evidence from this analysis suggests that the discharge is not contributing finer particles to the benthos near the outfall terminus.



Table 4-1. Goleta Sanitary District locations, survey information and weather conditions during the sediment and trawling survey.

Stations	ВІ	B2	В3	B4	B5	B6	TB3	TB6
Date	23-Oct-14							
Time	10:38	10:12	9:50	9:25	8:54	8:14	11:16	13:15
Research Vessel	Hey Jude							
Survey Program	Benthic Sediment	Benthic Sediment	Benthic Sediment	Benthic Sediment	Benthic Sediment	Benthic Sediment	Trawl, Bioaccum.	Trawl, Bioaccum.
Dist. From Outfall (m)	1500	500	250	25	25	3000	250	3000
Direc. From Outfall (°M)	270	270	270	270	90	90	270	90
Depth (m)	26.5	26.5	26.8	26.5	26.5	26.5	26.8	22.9
Latitude (N)	34.58261	34.40192	34.40192	34.40192	34.40197	34.40283	34.40111	34.40339
Longitude (W)	119.84103	119.83069	119.82792	119.82547	119.82492	119.79269	119.83414	119.80194
Weather	Clear							
Tide	Outgoing	Outgoing	Outgoing	Outgoing	Incoming	Incoming	Outgoing	Outgoing
Swl. Ht. (ft)	3	3	3	3	3	3	3	3
Swl. Dir.	sw							
Wind Sp. (Kn)	5	5	6	4	4	4	4	5
Wind Dir.	w	W	NW	NW	NW	NW	SE	sw



Table 4-2. Sediment grab descriptions.

Station	Rep	Penetration (cm)	Surface Description	Surface Color	Odor	Analysis
B1	1	10.0	Fine Sand	Olive Green	None	Biology
B1	2	11.0	Fine Sand	Olive Green	Hydrogen Sulfide	Biology
B1	3	9.0	Fine Sand	Olive Green	None	Biology
B1	4	9.5	Fine Sand	Olive Green	None	Biology
B1	5	8.5	Fine Sand	Olive Green	None	
B1	6	11.5	Fine Sand	Olive Green	None	Biology Chemistry
B2	1	9.0	Fine Sand	Olive Green	None	Biology
B2	2	9.5	Fine Sand	Olive Green	None	
B2	3	10.0	Fine Sand	Olive Green	None	Biology
B2	4	10.0	Fine Sand	Olive Green	None	Biology Biology
B2	5	10.5	Fine Sand	Olive Green	None	
B2	6	6.5	Fine Sand	Olive Green	None	Biology
B3	1	6.0	Fine Sand	Olive Green	None	Chemistry
B3	2	10.5	Fine Sand	Olive Green		Chemistry
B3	3				None	Biology
		8.5	Fine Sand	Olive Green	None	Biology
B3	4	11.5	Fine Sand	Olive Green	None	Biology
B3	5	10.5	Fine Sand	Olive Green	None	Biology
B3	6	11.5	Fine Sand	Olive Green	None	Biology
B4	1	9.5	Fine Sand	Olive Green	None	Biology
B4	2	9.0	Fine Sand	Olive Green	None	Biology
B4	3	8.5	Fine Sand	Olive Green	None	Biology
B4	4	6.0	Fine Sand	Olive Green	None	Chemistry
B4	5	9.0	Fine Sand	Olive Green	None	Biology
B4	6	8.5	Fine Sand	Olive Green	None	Biology
B5	1	8.0	Fine Sand	Olive Green	None	Biology
B5	2	10.0	Fine Sand	Olive Green	None	Biology
B5	3	7.0	Fine Sand	Olive Green	None	Biology
B5	4	8.0	Fine Sand	Olive Green	None	Biology
B5	5	11.0	Fine Sand	Olive Green	None	Biology
B5	6	8.5	Fine Sand	Olive Green	None	Chemistry
B6	1	13.0	Fine Sand	Olive Green	None	Biology
B6	2	10.0	Fine Sand	Olive Green	None	Biology
B6	3	10.0	Fine Sand	Olive Green	None	Biology
B6	4	13.0	Fine Sand	Olive Green	None	Biology
B6	5	11.0	Fine Sand	Olive Green	None	Biology
B6	6	11.0	Fine Sand	Olive Green	None	Chemistry

Table 4-3. Grain size characteristics of each Goleta station.

Station	Median (microns) 1.	Category	Sorting Index 2.	Sorting	% Fines
B1	95	very fine sand	1.88	poorly sorted	25
B2	91	very fine sand	1.65	poorly sorted	25
В3	109	very fine sand	1.50	poorly sorted	21
B4	111	very fine sand	1.74	poorly sorted	24
B5	115	very fine sand	1.45	poorly sorted	20
B6	88	very fine sand	1.16	poorly sorted	20

^{1. 0-4 =} clay, 4-8 = very fine silt, 8-16 = fine silt, 16-31 = medium silt, 31-63 = coarse silt, 63-125 = very fine sand, 125-250 = fine sand, 250-500 = medium sand, 500-1000 = coarse sand.

^{2. &}lt;0.35 = very well sorted, 0.35-0.50 = well sorted, 0.50-0.71 = moderately well sorted, 0.71-1.00 = moderately sorted, 1.0-2.0 = poorly sorted, 2.0-4.0 = very poorly sorted, >4.0 = extremely poorly sorted.



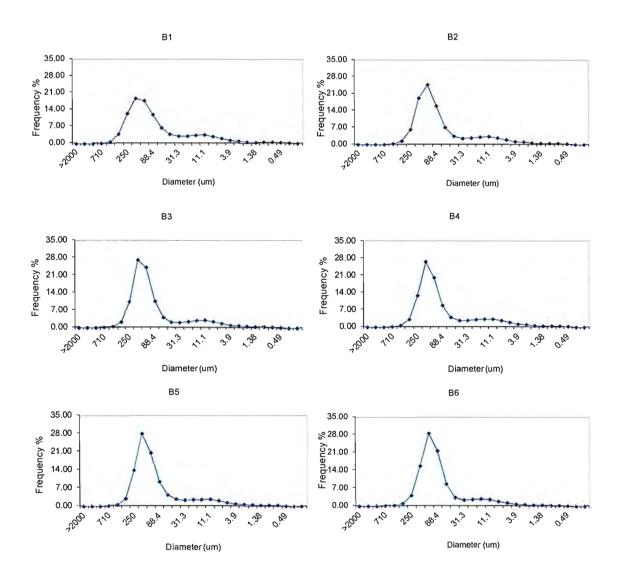


Figure 4-1. Particle size frequency (%) at each station in the Goleta survey area.



CHAPTER 5

Chemical Characteristics of Sediments

5.1. Background

Sources of potential contaminants discharged into the Southern California Bight include treated municipal and industrial wastewater, storm water runoff from urbanized areas, disposal of dredged materials, aerial fallout, oil and hazardous material spills, boating and other sources. Bottom sediments are often the fate of these contaminants, where they can reside for long periods of time, exerting effects at various levels of biological organization (SCCWRP 1998). Organic and metal contaminants tend to adsorb more readily on finer particles and can thus accumulate in areas of deposition. This accumulation of contaminants can impact resident organisms living both within the sediments and on the surface.

5.2. Materials and Methods

Field sampling for all benthic sediment components is described in Chapter 4, Section 4.2, Materials and Methods. Single sediment grabs were collected at stations B1 through B6 (Figure 5-1). Sediment portions to be chemically analyzed were removed from the top two centimeters of the grab sample with a stainless steel spatula and placed in pre-cleaned glass bottles with Teflon-lined caps. During all collections, the sides of the grab were avoided. Samples were immediately placed on ice and returned to the laboratory. PHYSIS Environmental Laboratories, located in Anaheim, California, performed all chemical analyses. Results were standardized to $\mu g/g$ dry weight for undifferentiated organics and metals and $\mu g/Kg$ dry weight for complex organics.

Since replicate field samples are not required, results were correlated against distance from the outfall diffuser. When appropriate, correlations were designated as significant ($p \le 0.05$) or marginally significant (0.05 , see Section 3.5.) and expected (negative) or unexpected (positive) (see Section 3.5.1). Since grain size can have an important effect on the ability of contaminants to adhere to particles, results were also correlated against percent fine particle size. The expected sign for particle size would be negative (increasing concentrations with smaller size).

As described in (Section 4.4.), areas west of the diffuser are known sources of natural oil seepages; therefore, results were also correlated against distance from Goleta Point. Like distance from outfall, the expected sign would be negative. Spearman's correlation was used to assess spatial trends (see Sokal and Rohlf 1981).

In order to determine long-term trends, 2014data were compared to results from monitoring surveys that began in 1991 (Brown and Caldwell 1992, 1993, 1994, 1995, 1996, 1997, 1998; Aquatic Bioassay 1999 to2012). Data were also compared to results of "reference" sediments from uncontaminated areas collected and analyzed by the Southern California Bight Regional Monitoring Program(SCBRMP) in 1998, 2003 and 2008. Finally, results were compared to the limits presented in two NOAA studies (NOAA 1990 and Long, et. al. 1995). In these studies, researchers compiled published information regarding the toxicity of chemicals to benthic organisms. The data for each compound were sorted, and the lower 10th percentile and median (50th) percentile were identified. The lower 10th percentile in the data



was identified as an Effects Range-Low (ER-L) and the median was identified as an Effects Range-Median (ER-M).

Per the NPDES permit, all contaminants were "normalized" to percent fine sediments and percent total organic carbon (TOC) at each station. NOAA scientists have determined that normalizing data from sediments that contain less than 20% silt and clay can cause erroneously high results; therefore, results from samples containing less than 20% fine components should be viewed with caution (NOAA 1990).

5.3. Results

Table 5-1 lists all of the chemical constituents measured from samples collected at each of the six benthic sediment stations. These compounds have been separated here into three main groups: undifferentiated organic compounds, heavy metals, and complex organic compounds. Complex organic compounds are further divided into chlorinated pesticides, polychlorinated biphenyls (PCB's), and polynuclear aromatic hydrocarbons (PAH's). Appendix tables10-4and 10-5 present data normalized to percent fine sediments (silt and clay fractions) and percent TOC. Appendix table 10-6 lists the constituents minimum detection limits (MDL), reporting limits (RL) and methods. Figure 5-2 shows the average (± standard deviation) concentration for all Goleta stations combined, for each constituent measured from 1991 to present. Tables 5-4 and 5-5 compare the Goleta sediment chemistry results with the 1998, 2003 and 2008SCBRMP surveys and the NOAA ER-L and ER-M values.

5.3.1 Undifferentiated Organics

The undifferentiated organics discussed in this report includes groups of compounds whose concentrations can help to determine the extent of anthropogenic contaminant loading in an area. These groups are discussed below:

- Total organic carbon (TOC) is a measure of the amount of carbon derived from plant and animal sources. It is a better measure of the portion of a sample derived from these sources than is percent volatile solids (Soule et al. 1996).
- Sources of oil and grease can be attributed to storm water runoff and ocean going vessels. The extent that people dump used motor oil into storm drains is unknown. Also, the Goleta outfall is located in an area of natural oil seeps, which may be a natural source.
- Total Kjeldahl Nitrogen (TKN) is the method used for the measure of organic nitrogen in water and sediments. Organic nitrogen is present due to the breakdown of animal products and includes such natural materials as proteins and peptides, nucleic acids, urea, and numerous synthetic organic materials (APHA 1995).
- \bullet Acid volatile sulfide (H_2S) is an indicator of organic decomposition occurring particularly in anoxic sediments and characterized by a rotten egg smell. No sediment reference values are available for sulfides.

5.3.1.1 Undifferentiated Organics Spatial Patterns

The concentrations for each of the undifferentiated organics measured for this survey are listed in Table 5-1. Similar to 2013, the concentrations of oil and grease weregreatest at Station B1 offshore Goleta Point (793 mg/L) and decreased at stations nearest to the outfall until the lowest concentration was measured at station



B5 (193mg/Kg). Total Kjeldahl nitrogen (TKN) concentrations were similar across sites, ranging from 427 mg/L to 583 mg/L.TOC concentrations were least near outfall stations B4 (4,900 mg/L) and B5 (4,300 mg/L), compared to station B2 (10,300 mg/L). Acid volatile sulfide (AVS) was similar across sites, ranging from 3.06 mg/L to 5.72 mg/L.

Each undifferentiated organic correlated unexpectedly (increased) with distance from the outfall, except AVS. None of the correlations with distance to the outfall were significant, except for TOC which was marginally significant (0.05). All of the undifferentiated organics, except TKN, correlated expectedly with distance to Goleta Point; oil and grease with marginal significance. Each undifferentiated organic constituent correlated unexpectedly (increased with increasing particle size) and non-significantly with sediment particle size.

5.3.1.2 Undifferentiated Organic Ranges Compared with Past Years

Each of the undifferentiated organics measured during this survey were within their reported range since 1991 (Figure 5-2). Acid volatile sulfideswhich were historically high in 2011, dropped to background levels in 2012 and remained low in 2014. Concentrations of oil and grease, TKN, TOC and acid volatile sulfides in 2013 were variable but within range of the past 20 years with no sustained increasing or decreasing trends evident.

5.3.1.3 Undifferentiated Organics Compared with Reference Surveys

The average concentrations of undifferentiated organics reported in this survey were compared to concentrations found during three southern California regional surveys conducted in 1998, 2003 and 2008 (Table 5-4 and 5-5). O&G, TKN and AVS were not measured during these surveys. Average TOC concentrations in the Goleta survey area were lower or similar to concentrations measured by the other surveys, except when compared to 2003 inner shelf stations which were approximately two times lower. ER-L and ER-M threshold limits are not available for these constituents.

5.3.2 Heavy Metals

Heavy metals in the marine environment are relatively ubiquitous and, with the exception of mercury, can normally be detected in sediments in low amounts. When anthropogenic sources increase sediment concentrations above levels that can be assimilated by benthic organisms, their assemblages can be impaired. For example:

- Aluminum is generally considered to be nontoxic to organisms in its elemental state and is one of the most common elements on earth.
- Antimony is used for alloys and other metallurgical purposes. The salts, primarily sulfides and oxides are employed in the rubber, textile, fireworks, paint, ceramic, and glass industries (SWRCB 1973). Acute and chronic toxicity of antimony to freshwater aquatic life occur at water concentrations as low as 9000 to 1600 ppm, and toxicity to algal species occurs at about 610 ppm. There is no saltwater criterion available for antimony (Long and Morgan 1990).
- Arsenic is carcinogenic and teratogenic (causing abnormal development) in mammals and is mainly used as a pesticide and wood preservative. Inorganic arsenic can affect marine plants at concentrations as low as 13 to 56 ppm and marine animals at about 2000 ppm (Long and Morgan 1990). The USEPA (1983) gives a terrestrial range of 1-50 ppm, with an average of 5 ppm.



- Cadmium is widely used in manufacturing for electroplating, paint pigment, batteries and plastics. Toxicity in water to freshwater animals ranges from 10 ppb to 1 ppm, as low as 2 ppm for freshwater plants, and 320 ppb to 15.5 ppm for marine animals (Long and Morgan 1990). The USEPA (1983) places the terrestrial range for cadmium at 0.01 to 0.7 ppm, with an average of 0.06 ppm.
- Chromium is widely used in electroplating, metal pickling, and many other industrial processes. Chromium typically occurs as either chromium (III) or chromium (VI), the latter being considerably more toxic. Acute effects to marine organisms range from 2,000 to 105,000 ppm for chromium (VI) and 10,300 to 35,500 ppm for chromium (III). Chronic effects range from 445 to 2,000 ppb for chromium (VI) and 2,000 to 3,200 ppb for chromium (III) (Long and Morgan 1990). The terrestrial range is 1 to 1,000 ppm with an average of 100 ppm (USEPA, 1983).
- Copper is widely used in anti-fouling paints. Saltwater animals are acutely sensitive to copper in water at concentrations ranging from 5.8 to 600 ppm. Mysid shrimp indicate chronic sensitivity at 77 ppm (Long and Morgan 1990).
- Iron is generally not considered toxic to marine organisms. Iron, in some organic forms, is a stimulator for phytoplankton blooms. Recent experiments in deep-sea productivity have shown a considerable increase in phytoplankton in normally depauperate mid-ocean waters when iron is added (Soule et al. 1996).
- Older paints and leaded gasoline are a major source of lead. Lead may be washed into the Harbor or become waterborne from aerial particulates. Adverse effects to freshwater organisms range from 1.3 to 7.7 ppm, although marine animals may be more tolerant (Long and Morgan 1990).
- Mercury is a common trace metal once used in industry and as a biocide. Acute toxicity to marine organisms in water ranges from 3.5 to 1678 ppm. Organic mercury may be toxic in the range of 0.1 to 2.0 ppm (Long and Morgan 1990).
- Nickel is used extensively in steel alloys and plating. Nickel is chronically toxic to marine organisms in seawater at 141 ppm (Long and Morgan 1990).
- Selenium is used as a component of electrical apparatuses and metal alloys and as an insecticide. Although there is no data available for selenium toxicity to marine organisms, the present protection criteria range is from 54 to 410 ppb (USEPA 1986). The normal terrestrial range is from 0.1 to 2.0 ppm with a mean of 0.3 ppm. Selenium and lead levels found and reported in Least Tern eggs from Venice Beach and North Island Naval Station in San Diego County were considered to be harmful to development (Soule et al. 1996).
- Silver has many uses in commerce and industry including photographic film, electronics, jewelry, coins, and flatware and in medical applications. Silver is toxic to mollusks and is sequestered by them and other organisms. Silver increases in the Southern California Bight with increased depth; high organic content and percent silt (Mearns et. al., 1991). The range in the rural coastal shelf is from 0.10 to 18 ppm, in bays and harbors from 0.27 to 4.0 ppm, and near outfalls 0.08 to 18 ppm (Soule et al. 1996). The normal terrestrial level ranges from 0.01 to 5.0 ppm, with a mean of 0.05 ppm.
- Soule and Oguri (1987, 1988) found the effects of tributyl tin can be toxic in concentrations as low as 50 parts per trillion in water. The terrestrial range for tin is 2 to 200 ppm, with a mean of 10 ppm. The California Department of Fish and Game considers tributyl tin to be the most toxic substance ever released in the



- marine environment. Tributyl tin may not be as bio-available in sediments as it is in seawater, and therefore may not affect the benthic biota in the same fashion.
- Zinc is widespread in the environment and is also an essential trace element in human nutrition. It is widely used for marine corrosion protection, enters the waters as airborne particulates, and occurs in runoff and sewage effluent. Acute toxicity of zinc in water to marine fish begins at 192 ppm, and chronic toxicity to marine mysid shrimp can occur as low as 120 ppm (Long and Morgan 1990). The normal terrestrial range is from 10 to 300 ppm, with a mean of 50 ppm (Soule et al. 1996).

5.3.2.1 Heavy Metal Spatial Patterns

The concentrations for each of the heavy metals measured for this survey are listed in Table 5-1. Of the fourteen metals measured, all were above detection at each of the sites. Differences in the concentrations of each metal among sites were small. Each of the fourteen metals correlated unexpectedly (increased) with distance from the outfall, except mercury and tin which did not correlate significantly. There were no significant correlations (p < 0.05) for any of the fourteen metals with either the distance from Goleta Point or sediment particle size.

5.3.2.2 Heavy Metal Ranges Compared with Past Years

Each of the heavy metals measured during this survey were within their reported range since 1991 and there were no clear increasing or decreasing concentration trends, especially in recent years (Figure 5-2).

5.3.2.3 Heavy Metals Compared with Reference Surveys

The average concentrations of 14 of the heavy metals measured in this survey were compared to concentrations found during three SCBRMP surveys in 1998, 2003 and 2008 (Tables5-4). Of the metals where comparisons could be made, several slightly exceeded concentrations measured in other surveys (aluminum, arsenic, cadmium, chromium, copper, mercury, nickel) (Table 5-5).

5.3.2.4 Heavy Metals Compared with NOAA Effects Range Thresholds

Metals concentrations measured at each station in the Goleta survey area during 2013were compared to the ER-L and ER-M threshold values (Table 5-4). All metal concentrations were below both the ER-L and ER-Mthreshold limits.

5.3.3 Complex Organics

5.3.3.1 Pesticides, PCB's and PAH's

Pesticides, PCBs and PAHs are contaminants that are widespread in the environment, are toxic to marine organisms when concentrations are increased and can cause reproductive failure in organisms at higher levels in the food chain. The sources and relative toxicity of each of these organic chemical groups are discussed below.

• DDT is a pesticide that has been banned since the early 1970's, but the presence of non-degraded DDT suggests that either subsurface DDT is being released during erosion and runoff in storms, or that fresh DDT is still in use and finding its way into coastal waters (Soule et al. 1996). DDT has been found to be chronically toxic to bivalves as low as 0.6 ppb in sediment. Toxicity of two of DDT's breakdown products, DDE and DDD, were both chronically toxic to bivalve larvae as low as about 1 ppb (Long and Morgan 1990).



- Of the non-DDT pesticides, concentrations of chlordane between 2.4 and 260 ppm in water are acutely toxic to marine organisms. Heptachlor is acutely toxic in water from 0.03 to 3.8 ppm. Heptachlor epoxide, a degradation product of heptachlor, is acutely toxic to marine shrimp at 0.04 ppm in water. Dieldrin is acutely toxic to estuarine organisms from 0.7 to 10 ppb. Endrin shows acute toxicity within a range of 0.037 to 1.2 ppb. Aldrin is acutely toxic to marine crustaceans and fish between 0.32 and 23 ppb. The EPA freshwater and saltwater criteria for aldrin are 3.0 and 1.3 ppb, respectively (Long and Morgan 1990). No toxicity data were found for any of the other chlorinated compounds measured during this survey.
- Although PCBs are not pesticides, their similarity to other chlorinated hydrocarbons makes their inclusion in this section appropriate. Before being banned in 1970, the principal uses of PCBs were for dielectric fluids in capacitors, as plasticizers in waxes, in transformer fluids, and hydraulic fluids, in lubricants, and in heat transfer fluids (Laws 1981). Arochlor 1242, a PCB congener, was acutely toxic in water to marine shrimp in ranges of 15 to 57 ppm (Long and Morgan 1990).
- The major sources of polynuclear aromatic hydrocarbons (PAH's) are believed to be the combustion of fossil fuels and petroleum or oil shales. PAH impact is characterized by altered community structure, abundance, and diversity near the pollutant source (Daily, et.al. 1993).

5.3.3.2 Pesticide, PCB, and PAH Spatial Patterns

Pesticides, PCB and PAH concentrations at the six sampling stations are listed in Table 5-1 and complex organic derivatives are listed in appendix table 10-7. Total DDTs were above detection at each station and was greatest at station B1 near Goleta Point (4.5 ug/Kg) and least at outfall station B4 (1.5 ug/Kg). Total DDT correlated unexpectedly with distance to the outfall and with marginal significance. Each of the other chlorinated hydrocarbons was below detection.

Both total PCBs and Aroclor were below detection at each site except stations B2 and B3. Similar to past years, total PAHs were above detection at each site in the survey area, with concentrations ranging from 20.7ug/Kg at station B3 to 109.8 at station B5near the outfall. Total PAHs correlated expectedly and non-significantly with the distance to the outfall, and unexpectedly and non-significantly with distance from Goleta Point.

5.3.3.3 Pesticide, PCB and PAH Ranges Compared with Past Years

Total DDT pesticides, chlorinated hydrocarbonsand PAH concentrations were within the range of previous years, although total DDT concentrations were elevated again in 2014 compared to 2010 thru 2012 (Figure 5-2). Total PCBs which had been below detection in the Goleta survey area for ten years since 2004, were slightly elevated at stations B2 and B3.

5.3.3.4 Pesticides, PCB's and PAH's Compared with Reference Surveys

The average concentrations of chlorinated pesticides (DDTs), PCBs and PAHs measured during the 2014surveywere compared to concentrations found during three southern California reference site surveys conducted in 1998, 2003 and 2008 (Table 5-4). Concentrations of each group of organics were similar to or less than



those measured on the inner shelf and near SPOTWs in during each of the SCBRMP reference surveys.

5.3.3.5 DDT Pesticides & PCB's Compared with NOAA Effects Range Thresholds

Pesticide, PCB and PAH concentrations measured in the Goleta survey area were compared to the NOAA ER-L and ER-M threshold values (Table 5-4). Each group of constituents waswell below these thresholds, except DDT which slightly exceeded the ER-L.



5.4 Discussion

Results from this survey support past studies in that the Goleta outfall discharge has little or no impact upon the chemical composition of local sediments. In order to confirm this, results from the chemical analysis of the benthos were compared among stations, compared to past surveys in the area, compared to other studies performed in southern California, and compared to levels known to have caused toxicity or other environmental impacts to resident marine infauna.

To determine if contaminant trends were significant across stations, results for each variable were correlated against three independent variables: distance from outfall diffuser, distance from Goleta Point, and median particle size. Goleta Point is a documented area of particularly heavy crude oil seepage. Since the diffuser is located relatively close to the Point (approximately 1,500 meters east) it is prudent to attempt to partition out the potential influences of seepages from the impact of the discharge. Correlation against particle size is important because it is well known that metals and other contaminants often adhere more readily to finer particles, and differences among stations may be due to differences in amount of fine material (Gray 1981).

Metal concentrations in the Goleta survey area were not as heavily influenced by distance from Goleta Point and particle size during 2014as in past years (Aquatic Bioassay 1997 to 2012). In fact, the concentrations of each of the metals were similar across sites. Of the fourteen metals measured, all correlated unexpectedly with distance to the outfall, except mercury and tin which did not correlate expectedly (increased) with distance to the outfall. .In addition, there were no significant correlations with either distance to Goleta Point or particle size.

Of the complex organic compounds measured, total DDTs and PAHs were above detection at each of the six stations, while total PCBs were only detected at stations B2 and B3. Total DDTs were greatest near Goleta Point and correlated with marginal significance with distance to the outfall. In past surveys, total PAHs were nearly always measured in greatest concentrations near Goleta Point and declined on a gradient toward the outfall. However, in 2013and 2014 (as in 2011 and 2012) this was not the case with PAH concentrations being similar across sites. The reason for the reduction in sediment PAH concentrations are unclear, but indicate that oil seepage from Goleta Point is highly variable.

This year's results were compared to past measures made in the Goleta survey area since 1991. Concentrations of sediment contaminants haveremained relatively stable over time and in 2014were within the ranges of past years. Acid volatile sulfides (AVS) which were greater on average in 2011 compared to any survey in the past 20 years, returned to normal background concentrations in 2012 and remained low thru 2014. Metals and organic contaminants remained either low or below detection in 2014. Total DDTs were again elevated in 2014 after being below detection from 2010 thru 2012.

This year's results were compared to sediment contaminant concentrations measured during the 1998, 2003 and 2008SCBRMP surveys on the inner shelf (depth < 30m) and near SPOTWs (SCBRMP 1998, 2003 and 2008). Of the metals where comparisons could be made, several slightly exceeded concentrations measured in other surveys (aluminum, arsenic, cadmium, chromium, copper, mercury, nickel). Concentrations of each group of organics were similar to or less than those



measured on the inner shelf and near SPOTWs in during each of the SCBRMP reference surveys.

The Goleta data were also compared to NOAA's Effects Range Low (ER-L) and Effects Range Median (ER-M) criteria. Based upon historical research, sediments with levels of chemical contaminants exceeding ER-L values have a "potential" of affecting sensitive benthic infauna or the sensitive live stages of the more tolerant organisms. Sediments containing contaminants that exceed ER-M values will "probably" have a negative impact upon several groups of infauna organisms. In 2014each constituent was well below the ER-L thresholds and far below the ER-M thresholds. The only exception to this was total DDT which slightly exceeded the ER-L. This indicates that Goleta sediments were not likely to have had an adverse effect on the benthic infauna community.

In summary, of the 22 constituents measured in Goleta sediments during the 2014survey, none correlated expectedly and significantly with distance from the outfall. Since the concentration of the pollutants emanating from the plant are very low or below detection, the detection of contaminants in the vicinity of the outfall is likely due to other anthropogenic inputs such as runoff from Goleta Slough, areal deposition or naturally occurring processes such as the release of oil from the seeps located offshore of Goleta Point. Comparison of Goleta sediments with historical reference data from the southern California Bight showed that most constituents were similar to or below baseline concentrations. Additionally, all sediment chemical concentrations were below those levels thought to cause toxicity to sensitive infauna organisms.



Figure 5-1. Benthic sediment sampling locations (Stations B1 – B6) in the Goleta survey area.

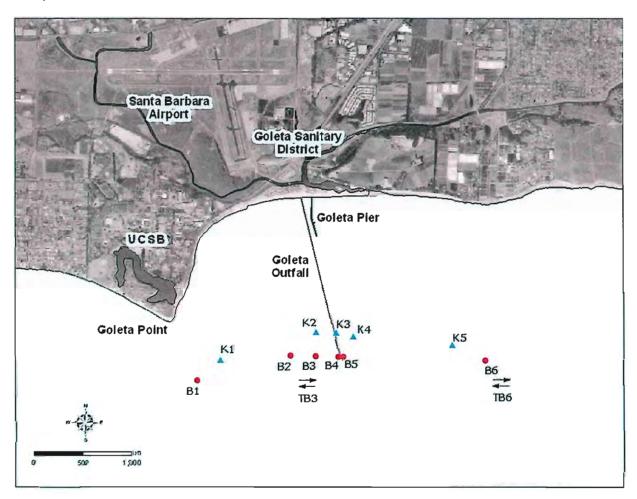




Table 5-1. Sediment contaminant concentrations (dry weight) in the Goleta survey area.

Undifferentiated Organics Oil and Grease (detention = 100 μg/g) ^{2.} 793 605 372 223 193 332 420 234 0.64 -0.83 0.71 770 (detection = 0.6 μg/g) 449 550 427 528 482 583 503 61 0.46 0.71 0.70 770 (detection = 100 μg/g) ^{3.} 9800 10300 5000 4900 4300 5500 6833 2679 0.75 -0.60 0.75 0.60 0.75			Sedi	ment Stat	ions					<u>C</u>	orrelation	<u>1S</u>
Oil and Grease (detention = 100 μg/g) ³ . 793 605 372 223 193 332 420 234 0.64 -0.83 0.0 TKN (detection = 0.6 μg/g) 449 550 427 528 482 583 503 61 0.46 0.71 0.0 TC (detection = 100 μg/g) ³ . 9800 10300 5000 4900 4300 5500 6633 2679 0.75 -0.60 0.0 AVS (detection = 0.05 μg/g) 4.69 4.15 5.72 4.08 3.06 3.29 4.17 0.97 -0.26 -0.50 0.0 AVS (detection = 1.0 μg/g) 9051 9934 6968 8726 10785 10605 9345 1423 0.37 0.41 -0.0 Arsenic (detection = 0.025 μg/g) 0.13 0.11 0.09 0.11 0.10 0.11 0.107 0.011 0.40 -0.38 0.0 Arsenic (detection = 0.025 μg/g) 0.34 0.47 0.40 0.37 0.35 0.49 0.40 0.06 0.51 0.73 -0.60 0.0 Arsenic (detection = 0.0025 μg/g) 24.82 28.04 21.04 24.87 29.15 30.56 26.41 3.50 0.47 0.57 -0.60 0.0 Arsenic (detection = 0.0025 μg/g) 8842 10511 7713 9204 9575 10667 9385 1125 0.44 0.59 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.01 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.01 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 0.00 0.00 0.00 0.00 0.00 0.00	Constituent ^{1.}	B1	B2	В3	B4	B5	86	Mean	S.D.	Outfall	Point	Prt.Sz.
Oil and Grease (detention = 100 μg/g) ³ . 793 605 372 223 193 332 420 234 0.64 -0.83 0.0 TKN (detection = 0.6 μg/g) 449 550 427 528 482 583 503 61 0.46 0.71 0.0 TC (detection = 100 μg/g) ³ . 9800 10300 5000 4900 4300 5500 6633 2679 0.75 -0.60 0.0 AVS (detection = 0.05 μg/g) 4.69 4.15 5.72 4.08 3.06 3.29 4.17 0.97 -0.26 -0.50 0.0 AVS (detection = 1.0 μg/g) 9051 9934 6968 8726 10785 10605 9345 1423 0.37 0.41 -0.0 Arsenic (detection = 0.025 μg/g) 0.13 0.11 0.09 0.11 0.10 0.11 0.107 0.011 0.40 -0.38 0.0 Arsenic (detection = 0.025 μg/g) 0.34 0.47 0.40 0.37 0.35 0.49 0.40 0.06 0.51 0.73 -0.60 0.0 Arsenic (detection = 0.0025 μg/g) 24.82 28.04 21.04 24.87 29.15 30.56 26.41 3.50 0.47 0.57 -0.60 0.0 Arsenic (detection = 0.0025 μg/g) 8842 10511 7713 9204 9575 10667 9385 1125 0.44 0.59 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.01 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.01 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0.0 Arsenic (detection = 0.025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 0.00 0.00 0.00 0.00 0.00 0.00	Hadiffe we make to all Overs = 1											
TKN (detection = 0.6 μg/g) TOC (detection = 100 μg/g) ³ 449 550 427 528 482 583 503 61 0.46 0.71 0. TOC (detection = 100 μg/g) ³ 469 4.15 5.72 4.08 3.06 3.29 4.17 0.97 -0.26 -0.50 0. Heavy Metals Aluminum (detection = 1.0 μg/g) Antimony (detection = 0.025 μg/g) Arsenic (detection = 0.025 μg/g) Cadmium (detection = 0.025 μg/g) Cadmium (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) A.11 6.00 3.67 4.35 4.16 4.79 4.51 0.81 0.16 0.21 0. Mercury (detection = 0.0025 μg/g) Architection = 0.0025 μg/g) Architection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) A.11 6.00 3.67 4.35 4.16 4.79 4.51 0.81 0.16 0.21 0. Mercury (detection = 0.0001 μg/g) ³ Nickel (detection = 0.0001 μg/g) ³ Nickel (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Architection		793	605	372	223	193	332	420	234	0.64	-0.83	0.66
TOC (detection = 100 μg/g) ³ AVS (detection = 0.05 μg/g) 4.69 4.15 5.72 4.08 3.06 3.29 4.17 0.97 -0.26 -0.50 0. Heavy Metals Aluminum (detection = 1.0 μg/g) Antimony (detection = 0.025 μg/g) Arsenic (detection = 0.025 μg/g) Cadmium (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Lead (detection = 0.0025 μg/g) Lead (detection = 0.0025 μg/g) Lead (detection = 0.0025 μg/g) Arsenic (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Arsenic (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Lead (detection = 0.0025 μg/g) Lead (detection = 0.0025 μg/g) Nickel (detection = 0.00025 μg/g) Selenium (detection = 0.001 μg/g) Selenium (detection = 0.01 μg/g) Selenium (detection = 0.025 μg/g) Tin (detection = 0.025 μg/g) Arsenic (detection = 0.025 μg/g) Arsenic (detection = 0.0025 μg/g) Arsenic (de												0.07
AVS (detection = 0.05 μg/g) Heavy Metals Altımınum (detection = 1.0 μg/g) Antimony (detection = 0.025 μg/g) Arsenic (detection = 0.025 μg/g) Cadmium (detection = 0.025 μg/g) Cadmium (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) A1.11 6.00 3.67 4.35 4.16 4.79 4.51 0.81 0.16 0.21 0.81 0.73 0.75 0.40 0.75 0.75 0.75 0.75 0.75 0.75 0.75 0.7		9800	10300	5000	4900	4300	5500	6633	2679	0.75	-0.60	0.60
Heavy Metals Aluminum (detection = 1.0 μg/g) 9934 6968 8726 10785 10605 9345 1423 0.37 0.41 -0.08 0.11 0.107 0.011 0.40 -0.38 0.41 -0.08 0.13 0.11 0.09 0.11 0.10 0.11 0.107 0.011 0.40 -0.38 0.41 -0.28 0.45 0.		4.69	4.15	5.72	4.08	3.06	3.29	4.17	0.97	-0.26	-0.50	0.22
Aluminum (detection = 1.0 μg/g) Antimony (detection = 0.025 μg/g) Antimony (detection = 0.025 μg/g) Arsenic (detection = 0.025 μg/g) Cadmium (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Cadmium (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Antimony (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Antimony (detection = 0.0000 μg/g) Antimony (detection = 0.00000 μg/g) Antimony (detection = 0.00000 μg/g) Antimony (detection = 0.00000 μg/g) Antimony (detection = 0.000000 μg/g) Antimony (detection = 0.0000000000000000000000000000000000	1000											
Antimony (detection = 0.025 μg/g) Arsenic (detection = 0.025 μg/g) Cadmium (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Lead (detection = 1.0 μg/g) Lead (detection = 0.0025 μg/g) Ansenic (detection = 0.0025 μg/g) Lead (detection = 1.0 μg/g) Lead (detection = 1.0 μg/g) Ansenic (detection = 0.0025 μg/g) Ansenic (detection = 1.0 μg/g) Lead (detection = 0.0025 μg/g) Ansenic (detection = 1.0 μg/g) Ansenic (detection = 0.0025 μg/g) Ansenic (detection = 1.0 μg/g) Ansenic (detection = 0.0025 μg/g) Ansenic (detection = 1.0 μg/g) Ansenic (detection = 0.0025 μg/g) Ansenic (detection = 0.0000000000000000000000000000000000	Heavy Metals											
Arsenic (detection = 0.025 μg/g) Cadmium (detection = 0.025 μg/g) Chromium (detection = 0.0025 μg/g) Lead (detection = 1.0 μg/g) Lead (detection = 0.0025 μg/g) Mercury (detection = 0.0001 μg/g) Selenium (detection = 0.001 μg/g) Selenium (detection = 0.0025 μg/g) Dissiver (detection = 0.0025 μg/g) A.72 Silver (detection = 0.0025 μg/g) A.73 A.74 A.72 A.75 A.75 A.70 A.75	Aluminum (detection = 1.0 μg/g)	9051	9934	6968	8726	10785	10605	9345	1423	0.37	0.41	-0.11
Cadmium (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Chromium (detection = 0.0025 μg/g) Copper (detection = 0.0025 μg/g) Lead (detection = 1.0 μg/g) Lead (detection = 0.0025 μg/g) Mercury (detection = 0.00001 μg/g) Nickel (detection = 0.0001 μg/g) Selenium (detection = 0.01 μg/g) Selenium (detection = 0.01 μg/g) Nickel (detection = 0.025 μg/g) O.017 O.052 O.018 O.018 O.024 O.022 O.022 O.026 O.013 O.014 O.26 O.059 O.059 O.018 O.017 O.050 Selenium (detection = 0.01 μg/g) Diri (detection = 0.0025 μg/g) O.09 O.09 O.09 O.10 O.09 O.09 O.00	Antimony (detection = 0.025 μg/g)	0.13	0.11	0.09	0.11	0.10	0.11	0.107	0.011	0.40	-0.38	0.70
Chromium (detection = 0.0025 μg/g) Copper (detection = 0.0025 μg/g) Iron (detection = 1.0 μg/g) Lead (detection = 0.0025 μg/g) Mercury (detection = 0.00025 μg/g) Nickel (detection = 0.00001 μg/g) Selenium (detection = 0.01 μg/g) Selenium (detection = 0.01 μg/g) Silver (detection = 0.01 μg/g) Tin (detection = 0.025 μg/g) Complex Organics (ng/g dry welght)² Chlordane 24.82 28.04 21.04 24.87 29.15 30.56 26.41 3.50 0.47 0.57 -0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.	Arsenic (detection = 0.025 μg/g)	4.72	5.62	4.58	5.10	4.95	5.24	5.04	0.37	0.14	0.37	0.33
Copper (detection = 0.0025 μg/g)	Cadmium (detection = 0.0025 μg/g)	0.34	0.47	0.40	0.37	0.35	0.49	0.40	0.06	0.51	0.73	-0.12
Iron (detection = 1.0 μg/g)	Chromium (detection = 0.0025 µg/g)	24.82	28.04	21.04	24.87	29.15	30.56	26.41	3.50	0.47	0.57	-0.18
Lead (detection = 0.0025 μg/g) Mercury (detection = 0.00001 μg/g) ^{3.} Nickel (detection = 0.01 μg/g) Selenium (detection = 0.025 μg/g) Silver (detection = 0.01 μg/g) Zinc (detection = 0.025 μg/g) Complex Organics (ng/g dry weight) ² Chlorinated Pesticides DDTs ^{3.} Lead (detection = 0.0025 μg/g) 3.71 3.84 3.14 3.67 3.64 3.65 3.61 0.24 0.20 -0.03 0 0.017 0.052 0.018 0.024 0.022 0.022 0.026 0.013 -0.14 0.26 0 0.020 0.023 0.18 0.29 13.06 16.12 13.58 1.85 0.59 0.65 0 0.20 0.23 0.18 0.20 0.18 0.28 0.21 0.04 0.81 0.77 -0.00 0.00 0.10 0.09 0.10 0.09 0.10 0.09 0.10 0.09 0.10 0.09 0.01 0.36 0.38 0 Tin (detection = 0.01 μg/g) 0.09 0.10 0.09 0.10 0.08 0.10 0.09 0.01 0.36 0.38 0 Zinc (detection = 0.025 μg/g) 21.84 27.56 20.70 24.59 24.46 28.79 24.65 3.13 0.47 0.71 -0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.	Copper (detection = 0.0025 µg/g)	4.11	6.00	3.67	4.35	4.16	4.79	4.51	0.81	0.16	0.21	0.45
Mercury (detection = 0.00001 μg/g)³. 0.017 0.052 0.018 0.024 0.022 0.022 0.022 0.026 0.013 -0.14 0.26 0 Nickel (detection = 0.01 μg/g) 12.51 15.51 11.30 12.96 13.06 16.12 13.58 1.85 0.59 0.65 0 Selenium (detection = 0.025 μg/g) 0.20 0.23 0.18 0.20 0.18 0.28 0.21 0.04 0.81 0.77 -0 Silver (detection = 0.01 μg/g) 0.09 0.10 0.09 0.10 0.08 0.10 0.09 0.01 0.36 0.38 0 Tin (detection = 0.025 μg/g) 0.55 0.59 0.44 0.57 0.77 0.62 0.59 0.11 -0.20 -0.60 0 Zinc (detection = 0.025 μg/g) 21.84 27.56 20.70 24.59 24.46 28.79 24.65 3.13 0.47 0.71 -0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.	lron (detection = 1.0 μg/g)	8642	10511	7713	9204	9575	10667	9385	1125	0.44	0.59	0.05
Nickel (detection = 0.01 μg/g) Selenium (detection = 0.025 μg/g) Silver (detection = 0.01 μg/g) Tin (detection = 0.025 μg/g) Zinc (detection = 0.025 μg/g) Complex Organics (ng/g dry weight)² Chlorinated PestIcIdes DDTs³. HCHs Chlordane 12.51 15.51 11.30 12.96 13.06 16.12 13.58 1.85 0.59 0.65 0 0.20 0.23 0.18 0.20 0.18 0.28 0.21 0.04 0.81 0.77 -0 0.09 0.10 0.09 0.10 0.08 0.10 0.09 0.01 0.36 0.38 0 0.09 0.10 0.09 0.10 0.08 0.10 0.09 0.01 0.36 0.38 0 0.55 0.59 0.44 0.57 0.77 0.62 0.59 0.11 -0.20 -0.60 0 21.84 27.56 20.70 24.59 24.46 28.79 24.65 3.13 0.47 0.71 -0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.	Lead (detection = 0.0025 μg/g)	3.71	3.84	3.14	3.67	3.64	3.65	3.61	0.24	0.20	-0.03	0.58
Selenium (detection = 0.025 μg/g) 0.20 0.23 0.18 0.20 0.18 0.28 0.21 0.04 0.81 0.77 -0.05	Mercury (detection = 0.00001 µg/g) ^{3.}	0.017	0.052	0.018	0.024	0.022	0.022	0.026	0.013	-0.14	0.26	0.20
Silver (detection = 0.01 µg/g) Tin (detection = 0.025 µg/g) Zinc (detection = 0.025 µg/g) Zinc (detection = 0.025 µg/g) Complex Organics (ng/g dry weight)² Chlorinated Pesticides DDTs³. HCHs 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.	Nickel (detection = 0.01 μg/g)	12.51	15.51	11.30	12.96	13.06	16.12	13.58	1.85	0.59	0.65	0.04
Tin (detection = 0.025 μg/g) Zinc (detection = 0.025 μg/g) Complex Organics (ng/g dry weight) ² Chlorinated Pesticides DDTs ^{3.} HCHs Chlordane 0.55 0.59 0.44 0.57 0.77 0.62 0.59 0.11 -0.20 -0.60 0.00 0.00 0.00 0.00 0.00 0.00 0.	Selenium (detection = 0.025 μg/g)	0.20	0.23	0.18	0.20	0.18	0.28	0.21	0.04	0.81	0.77	-0.07
Zinc (detection = 0.025 μg/g) Complex Organics (ng/g dry weight) ² Chlorinated Pesticides DDTs ^{3.} HCHs 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.	Silver (detection = 0.01 μg/g)	0.09	0.10	0.09	0.10	80.0	0.10	0.09	0.01	0.36	0.38	0.43
Complex Organics (ng/g dry weight) ²	Tin (detection = 0.025 μg/g)	0.55	0.59	0.44	0.57	0.77	0.62	0.59	0.11	-0.20	-0.60	0.60
Chlorinated Pesticides 4.5 2.5 2.2 1.5 2.3 2.6 2.60 1.01 0.84 -0.26 0.0	Zinc (detection = 0.025 μg/g)	21.84	27.56	20.70	24.59	24.46	28.79	24.65	3.13	0.47	0.71	-0.02
DDTs³. 4.5 2.5 2.2 1.5 2.3 2.6 2.60 1.01 0.84 -0.26 0 HCHs 0.0												
Chlordane 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.00 0.00		4.5	2.5	2.2	1.5	2.3	2.6	2.60	1.01	0.84	-0.26	0.09
	HCHs	0.0	0.0	0.0	0.0	0.0	0.0	0.00	0.00	0.00	0.00	0.00
Aldrin (detection = 1.0 unif(a)	Chlordane	0.0	0.0	0.0	0.0	0.0	0.0	0.00	0.00	0.00	0.00	0.00
Akinin (detection = 1.0 µg/ng) 1.0 1.0 1.0 1.0 1.0 1.0 0.00 0.00 0	Aldrin (detection = 1.0 μg/Kg)	1.0	1.0	1.0	1.0	1.0	1.0	1.00	0.00	0.00	0.00	0.00
Dieldrin (detection = 1.0 μg/Kg) 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.00 0.00	Dieldrin (detection = 1.0 µg/Kg)	1.0	1.0	1.0	1.0	1.0	1.0	1.00	0.00	0.00	0.00	0.00
Heptachlor (detection = 1.0 μg/Kg) 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.00 0.00	Heptachlor (detection = 1.0 μg/Kg)	1.0	1.0	1.0	1.0	1.0	1.0	1.00	0.00	0.00	0.00	0.00
Heptachlor epoxide (detection = 1.0 μg/Kg) 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.00 0.00	Heptachlor epoxide (detection = 1.0 μg/Kg)	1.0	1.0	1.0	1.0	1.0	1.0	1.00	0.00	0.00	0.00	0.00
Mirex (detection = 1.0 μg/Kg) 1.0 1.0 1.0 1.0 1.0 1.0 0.00 0.00 0.00	Mirex (detection = 1.0 μg/Kg)	1.0	1.0	1.0	1.0	1.0	1,0	1.00	0.00	0.00	0.00	0.00
Hexachiorobenzene (detection = 1.0 µg/Kg) 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.00 0.00	Hexachlorobenzene (detection = 1.0 μg/Kg)	1.0	1.0	1.0	1.0	1.0	1.0	1.00	0.00	0.00	0.00	0.00

Bold = Marginally significant (0.05

Bold = Significant (p < 0.05)



^{1.} Minimum detection limits, reporting limits and methods are listed in Appendix 10.4

^{2.} Complex organic derivatives are listed in Appendix 10.4.

^{3.} Non-normal data. Correlations by nonparametric Spearman's rho.

Table 5-1.continued

		Sed	iment Sta	tions					C	orrelation	ns
Constituent ^{1.}	B1	B2	В3	B4	B5	В6	Mean	S.D.	Outfall	Point	Prt.Sz.
Polychlorinated Biphenyls											
PCBs ^{3.}	0.0	1.2	1.4	0.0	0.0	0.0	0.43	0.67	-0.03	-0.37	0.30
Aroclors ^{3.}	0.0	10.0	11.0	0.0	0.0	0.0	3.50	5.43	-0.03	-0.37	0.30
Polycyclic Aromatic Hydrocarbons											
PAHs ^{3.}	42.7	40.0	20.7	44.6	109.8	31.3	48.18	31.45	-0.58	0.09	0.09
1-Methylnaphthalene (detection = 1.0 μg/Kg)	1.0	1.0	1.0	1.1	1.1	1.0	1.03	0.05	-0.57	-0.06	-0.07
1-Methylphenanthrene (detection = 1.0 µg/Kg) ^{3.}	2.3	1.4	1.2	1.0	2.1	1.5	1.58	0.51	0.43	-0.14	-0.09
2,3,5-Trimethylnaphthalene (detection = 1.0 μg/Kg)	1.0	1.0	1.0	1.0	1.0	1.0	1.00	0.00	0.00	0.00	0.00
2,6-Dimethylnaphthalene (detection = 1.0 μg/Kg)	1.3	1.4	1,1	1.0	2.0	1.5	1.38	0.35	0.08	0.20	-0.46
2-Methylnaphthalene (detection = 1.0 μg/Kg)3.	1.1	1.6	1.0	1.2	1.8	1.1	1.30	0.32	-0.47	0.17	0.14
Acenaphthene (detection = 1.0 μg/Kg)	1.0	1.0	1.0	1.0	1.0	1.0	1.00	0.00	0.00	0.00	0.00
Benz[a]anthracene (detection = 1.0 μg/Kg) ^{3.}	2.6	2.1	1.3	4.8	9.8	1.8	3.73	3.21	-0.58	0.09	0.09
Benzo[b]fluoranthene (detection = 1.0 μg/Kg) ^{3.}	5.2	5.4	2.9	4.7	11.2	3.7	5.52	2.94	-0.29	-0.14	0.31
Benzo[e]pyrene (detection = 1.0 μg/Kg) ^{3.}	12.7	10.2	6.0	2.6	11.2	6.2	8.15	3.83	0.32	-0.37	0.20
Benzo[g,h,i]perylene (detection = 1.0 μg/Kg) ^{3.}	6.3	4.7	2.7	1.0	9.8	3.6	4.68	3.08	0.12	-0.14	0.03
Biphenyl (detection = 1.0 μg/Kg)	1.0	1.0	1.0	1.0	1.0	1.0	1.00	0.00	0.00	0.00	0.00
Fluoranthene (detection = 1.0 µg/Kg) ^{3.}	7.1	5.8	1.7	2.6	20.1	4.9	7.03	6.71	0.03	-0.09	0.09
Naphthalene (detection = 1.0 μg/Kg)	1.6	2.1	1.2	1.5	2.4	1.8	1.77	0.43	-0.05	0.13	-0.08
Perylene (detection = 1.0 μg/Kg) ^{3.}	34.1	29.1	12.7	2.5	14.1	19.7	18.70	11.54	0.72	-0.49	0.37



^{2.} Complex organic derivatives are listed in Appendix 10.4.

^{3.} Non-normal data. Correlations by nonparametric Spearman's rho.

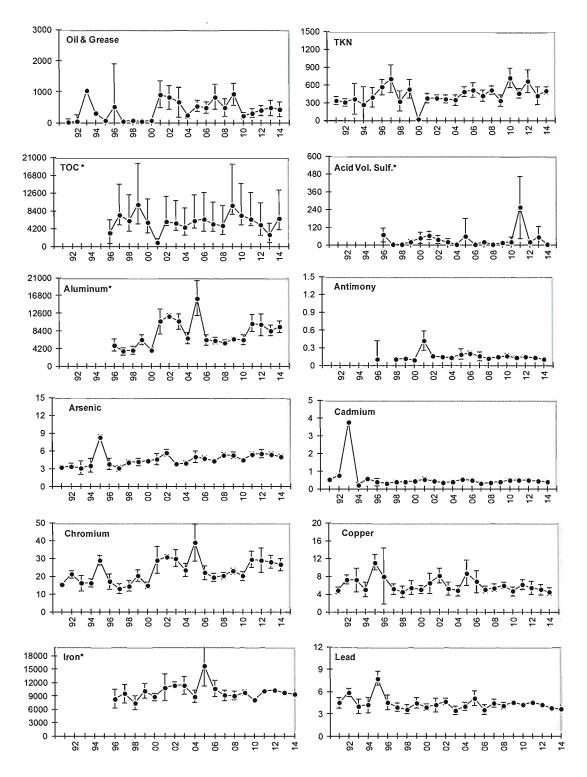


Figure 5-2. Average concentrations (±SD) of sediment contaminants measured between 1991 and 2014 in the Goleta survey area. TOC, acid volatile sulfide, aluminum, iron, selenium and tin were not measured from 1991 to 1995.



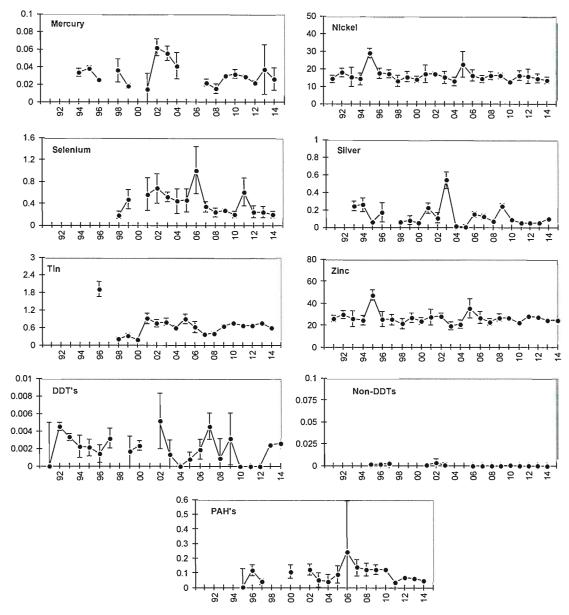


Figure 5-2. (continued)



Table 5-4. Comparison of sediment contaminants found in the Goleta survey area to the Southern California Bight Regional Monitoring Program (SCBRMP) data from1998, 2003 and 2008; and, the NOAA status and trends ERL and ERM threshold values. The SCBRMP survey includes comparisons against stations located near SPOTWs and shallow water reference sites.

		1		TANCOCO.	COBBAND (2000)			SCODD##D /2002/2	/2002/2		C/CCC/4000/3	(4000)3.	AIOA A /100014.	400014
	ופ	OLEIA S.D.		ייייייייייייייייייייייייייייייייייייייי	(2000)	ì		ייין איין	(2003)	-	TWINDS OF	(986)	770	(4000)
Constituent	Mean	Range	inne Mean	inner Sheif an 95% Cl	So Ca Mean	So Cal Bight an 95% CI	inner sheif Mean 95%	Sheif 95% CI	Small FOTW Mean 95%	-01W 95% CI	Mean	Shallow	Long,et.al. (1995) ³ : ER-L ER-M	. (1995)* ER-M
Undifferentiated Organics														
Oil and Grease	420	193 - 793	ł	1	ł	1	ŀ	ł	}	<u> </u>	ł	i	ţ	1
TKN	503	427 - 583	1	1	1	ı	ļ	1	ì	ł	ŀ	1	!	I
700	6633	4300 - 10300	00099	4100	30000	100	2700	800.00	5400	1600	2200	4200	;	1
AVS	4.2	3.1 - 5.7	l 	1	1	1	I	1	į	ł	ł	:	1	ŧ
Heavy Metals						_								
Aluminum	9345	6968 - 10785	5256	726	15372	1594	9212	2233	13244	3585		l	1	1
Antimony	0.11	0.09 - 0.13	0.12	0.02	0.28	0.04	0.14	0.04	0.15	0.02	1.09	1.59	2	25
Arsenic	5.04	4,58 - 5,62	4.3	1.2	6.70	1.20	4.2	1.4	4.6	0.67	79.7	4.39	8.2	70
Cadmium	0.40	0.34 - 0.49	0.23	0.03	0.88	0.12	0.20	90'0	0.22	0.05	0.28	0.36	1.2	9.6
Chromium	26.41	21.04 - 30.56	16	3.8	26.0	6.6	27	6.8	27	5.6	24.72	19.02	81	370
Copper	4.51	3,67 - 6.00	4.4	0.8	23.00	5.80	9.9	1.8	9.0	2.5	17.41	6.82	34	270
íron	9385	7713 - 10667	10239	2233	26218	3125	12952	2784	16255	3655	į	ļ	}	ļ
Lead	3.61	3.14 - 3.84	5.0	1.3	12.00	1.40	4.7	[:	4.90	0.81	15.92	10.14	46.7	218
Mercury	0.026	0.017 - 0.052	0.02	0.01	1.600	2.800	0.03	0.01	0.05	0.03	0.050	0.036	0.15	0.71
Nickel	13.58	11.30 - 16.12	თ	1.7	27.00	2.80	13	3.8	1	2.0	13.85	15.50	20.9	51.6
Selenium	0.21	0.18 - 0.28	0.44	0.11	3.50	2.60	69'0	0.22	0.55	0.12	0.97	0.47	ŀ	ŀ
Silver	60'0	0.08 - 0.10	0.12	90.0	0.91	0.40	0.13	90.0	0.14	90.0	0.12	0.19	1.0	3.7
ᄪ	0.59	0.44 - 0.77	I	:	1	!	ŀ	-	I	ì	!	1	ŀ	1
Zinc	24.65	20.70 - 28.79	25	6.8	71.00	5.90	34	7.8	40	8.0	52.14	33.59	150	410
Complex Organics														
DDTs	0.0026	0.0015 - 0.0045	5 0.0023	0.0004	0.1260	0.0970	0.0023	0.0004	0.0012	0.0002	0.020	0.036	0.00158	0.0461
HCHs	0.0000	0.0000 0.0000	1	ŀ	1	ŀ	i	i	!	1	}	ŀ	1	1
Chlordane	0.0000	0.0000 - 0.0000	0.0005	0.0001	0.0016	0.0008	0.00001	0.00001	0.0000.0	0.0000.0	!	i	ŀ	ì
PCBs	0.0035	0.0000 - 0.0110	0.0002	0.0000	0.1700	0.0067	0.0024	0.00001	0.0001	0.00001	0.004	0.005	0.0227	0.18
PAHs	0.0482	0.0207 - 0.1098	3 0.0512	0.0449	0.2860	0.0390	0.0512	0.0449	0.0249	0.0087	0.118	0.073	4.022	44.792
			-											

1. SCCWRP, 2012; 2. SCCWRP, 2006; 3.SCCWRP 2003; 4.Long and Morgan, 1990; 5.Long et al., 1995.



Table 5-5. Summary of sediment contaminant spatial trends and concentrations found in the Goleta survey area to the Southern California Bight Regional Monitoring Program (SCBRMP) data from1998, 2003 and 2008; and,the NOAA status and trends ERL and ERM threshold values.

	Expected Correlation	Expected & Significant			ceeds Refer				Exce	eeds
Constituent	w / Dist from Outfall	Correlation	2008 Inner Shelf	2008 So Cal Bight	2003 Inner Shelf	2003 SPOTW	1998 SPOTW	1998 Shallow	ER-L?	ER-M?
Oil and Grease	No	No								
TKN	No	No								
TOC	No	No	No	No	Yes	Yes	Yes	Yes		
AVS	Yes	No								
Aluminum	No	No	Yes	No	Yes	No				
Antimony	No	No	No	No	No	No	No	No	No	No
Arsenic	No	No	Yes	No	Yes	Yes	No	Yes	No	No
Cadmium	No	No	Yes	No	Yes	Yes	Yes	Yes	No	No
Chromium	No	No	Yes	No	No	No	No	Yes	No	No
Copper	No	No	Yes	No	No	No	No	No	No	No
Iron	No	No	No	No	No	No				
Lead	No	No	No	No	No	No	No	No	No	No
Mercury	Yes	No	Yes	No	No	No	No	No	No	No
Nickel	No	No	Yes	No	Yes	No	No	No	No	No
Selenium	No	No	No	No	No	No	No	No	No	No
Silver	No	No	No	No	No	No	No	No	No	No
Tin	Yes	No								
Zinc	No	No	No	No	No	No	No	No	No	No
DOTs	No	No	No	No	No	Yes	No	No	Yes	No
HCHs	No	No								
Chlordane	No	No	No	No	No	No			No	No
PCB'S	Yes	No	No	No	Yes	Yes	No	No	No	No
PAHS	Yes	No	No	Yes	No	Yes	No	No	No	No



CHAPTER 6

Benthic Infauna

6.1. Background

The benthic infauna community is composed of those species living in or on the bottom (benthos). This community is very important to the quality of the habitat because it provides food for the entire food web including juvenile and adult fishes who are bottom feeders. Usually polychaete annelid worms, molluscans, and crustaceans dominate the benthic fauna in shallow, silty, sometimes unconsolidated, habitats. In areas where sediments are contaminated or frequently disturbed by natural events such as storms or by manmade events, nematode round worms, oligochaete worms, or tolerant polychaetes or mollusks may dominate the fauna temporarily. Storms can cause organisms to be washed away or buried under transported sediment, or can cause changes in the preferred grain size for particular species. Excessive runoff may lower normal salinities, and thermal regime changes offshore may disturb the composition of the community. Some species of benthic organisms with rapid reproductive cycles or great fecundity can out-compete other organisms in recolonization, at least temporarily after disturbances, but competitive succession may eventually result in replacement of the original colonizers with more dominant species.

6.2. Materials and Methods

Field sampling for all benthic sediment components is described in Chapter 4. Sediments to be analyzed for infauna content were sieved through 1.0 millimeter screens. The retained organisms and larger sediment fragments were then washed into four-liter plastic bottles, relaxed with a magnesium sulfate solution, and preserved with 10% buffered formalin. Five replicates were collected from six benthic infauna stations (B1, B2, B3, B4, B5, and B6; see Figure 3-1). Screened and preserved sediments collected in the field were delivered to the Ventura laboratory for counting, sorting, and identification. Infauna were sorted out by Aquatic Bioassay staff biologists and separated into five groups: echinoderms, mollusks, polychaetes, crustaceans, and miscellaneous. For each station, organisms were counted per group in accordance with Techniques for Sampling and Analyzing the Marine Macrobenthos EPA 600/3-78-300, March 1978; Quality Assurance and Quality Control (QA/QC) for 301(h) Monitoring Programs: Guidance on Field and Laboratory Methods, Tetra Tech 1986; and Southern California Bight Pilot Project Field Operations Manual, 2008. Each sorted sample was re-checked by a second biologist for representatives not found during the first inspection. Infauna was identified by SCAMIT taxonomists Tony Phillips for and polychaetes, mollusks and other phyla, Dean Pasko for crustaceans and Megan Lily of the City of San Diego for echinoderms. A complete list of infauna is included in Appendix 10.6. Bioassay maintains and updates standardized type collections and voucher specimens for most southern California infauna.

Following enumeration of infauna organisms by species, the total and phyla group numbers of individuals, and numbers of separate species were compiled for each station replicate. In addition, several required biological indices were calculated: Shannon Weiner species diversity (H'), Margelef's richness index (d), Simpson's species diversity (SI), Schwartz's dominance (D), the infauna trophic index (ITI) and



Benthic Response Index (BRI). Analysis of Variance (ANOVA) was used to compare average metrics values among stations. Species compositions were compared using numerical classification and ordination. Brief descriptions of the indices are presented below.

<u>Shannon Diversity</u>. The Shannon Diversity Index (H') (Shannon and Weaver 1963) is defined as:

s
$$H' = -\sum \{(nj/N) \text{ Ln } (nj/N)\}, \\ j=1$$

where:

nj = number of individuals of the jth species,

s = number of species in the sample,N = number of individuals in the sample.

Margalef's Richness. Margalef's Species Richness Index (d) (Margalef 1958) is:

$$d = s-1 / Ln N,$$

where:

s = number of species in the sample, N = number of individuals in the sample.

Simpson's Diversity. The Simpson' Diversity Index (SI) (Simpson 1949) is:

$$SI = 1 - \sum_{i=1}^{S} (p_i)^2,$$

i=1

where:

 p_i = proportion of individuals of the *i*th species in the community.

<u>Schwartz' Dominance</u>. Schwartz's Dominance Index (D) is defined as the minimum number of species required accounting for 75% of the individuals in a sample (Schwartz 1978).

<u>Infauna Trophic Index.</u> This index measures the prevailing feeding modes of benthic infauna. Higher values denote southern California species assemblages dominated by suspension feeders, which are more characteristic of unpolluted environments. Lower index values denote assemblages dominated by deposit feeders more characteristic of areas near major outfalls (Word 1980):

ITI = -33.33
$$\{n_2 + (2)(n_3) + (3)(n_4) / n_1 + n_2 + n_3 + n_4\},$$

where: $n_1,...,n_4$ = numbers of individuals in species trophic groups 1,...,4, respectively.

Benthic Response Index. The BRI is the abundance-weighted average pollution tolerance of species occurring in a sample (Smith *et al.* 2001). The general index formula is:



$$BRI_{s} = \frac{\sum_{i=1}^{n} a_{si}^{f} p_{i}}{\sum_{i=1}^{n} a_{si}^{f}}$$
(1)

where BRI_s is the BRI value for sampling unit s, n is the number of species in s, p_i is the pollution tolerance of species i, a_{si} is the abundance of species i in s, and f is an exponent used to transform the abundance values. The primary objective of BRI development is to assign pollution tolerance scores p_i to species based on their position on a pollution gradient. Once assigned, the scores can be used to assess the condition of the benthic community by calculating the BRI. A reference threshold, below which natural benthic assemblages normally occur, was identified at an index value of 31, the point on the pollution vector where pollution effects first resulted in a net loss of species. Three additional thresholds of response to disturbance were defined at index values of 42, 53 and 73, representing points at which 25%, 50%, and 80% of the species present at the reference threshold were lost.

Analysis of Variance (ANOVA). ANOVA's were used to compare population variables and sediment chemistry concentrations among stations. ANOVA analysis requires two steps. In the first step, differences in a variable among stations are evaluated to determine if they are sufficiently large to be statistically significant (p \leq 0.05). If they are, then a second test must be performed to determine which stations are significantly different from another station or stations. In this report, this second step is called the comparison of means. For example, a comparison of means stating: OS1 > OS2, OS3 > OS4, indicates that, for that particular variable, Station OS1 is significantly larger than Stations OS2, OS3, and OS4, and Stations OS2 and OS3 are also significantly larger than Station OS4. For chemical contaminants, if stations near the outfall are significantly higher than stations farther away, that compound should be evaluated further. For population variables, the opposite is true.

<u>Cluster Analysis.</u>Cluster analysis was used to define groups of samples, based on species presence and abundance, which belong to the same community without imposing an *a priori* community assignment. Identified clusters were then evaluated to define the habitat to which they belong. In cluster analysis, samples with the greatest similarity are grouped first. Additional samples with decreasing similarity are then progressively added to the groups. The percentage dissimilarity (Bray-Curtis) metric (Gauch, 1982; Jongman et al., 1995) was used to calculate the distances between all pairs of samples. The cluster dendogram was formed using the unweighted pair-groups method using arithmetic averages (UPGMA) clustering algorithm (Sneath and Sokal, 1973). All steps were completed using the computer program MVSP (Multivariate Statistical Package, v3.12, 2000). Only the most commonly occurring species were used in the analysis, in this case only those that occurred at more than one station and season.

For normal (station by station) classifications, the Bray-Curtis Index is:

B.C. =
$$\sum_{i=1}^{s} \min (P_{ij}, P_{ik}),$$

where: P_{ij} = proportion of species i collected at station j,



 P_{ik} = proportion of species i collected at station k, s = number of species.

For inverse (species group by species group) classifications:

B.C. =
$$\sum_{i=1}^{N} \text{min } (P_{ij}, P_{ik}),$$

where now: P_{ij} = proportion collected at station i of species j,

P_{ik} = proportion collected at station i of species k,

N = number of stations.

Ordination analysis. Ordination analysis displays the sampling stations as points in a multidimensional space. The distances between the stations (points) in the space are proportional to the dissimilarity of the communities found at the respective stations. The different dimensions of the ordination space, called axes, define independent gradients of biological change in the community data. The projections of the station points onto the various axes are called scores. The axes are ordered so that the first axis displays a maximal amount of community change; the second axis defines a maximal amount of the remaining community change, and so on for subsequent axes. Often most of the relevant community changes are displayed in a few ordination axes.

6.3. Results

6.3.1. Benthic Infauna

6.3.1.1. Infauna Abundance

The simplest measure of resident animal health is the abundance of infauna collected per sampling effort. Measures of abundance include biomass and numbers of individuals, which is partially dependent upon the volume of sediment collected in the grab. For this survey, abundance was determined to be all of the non-colonial animals collected from one replicate Van Veen Grab (0.1 square meter surface area) and retained on a 1.0 mm screen (note that abundance per square meter can be easily calculated by multiplying individuals per grab by ten). Five replicates were collected from six sediment stations.

<u>Spatial infauna abundance patterns.</u>Infauna abundances at the six sediment sampling stations are listed in Table 6-1. Numbers of individuals were similar across sites and greatest at B6 (average = 695) furthest from the Goleta outfall and least at B5 (average = 527) nearest the outfall. Numbers of individuals correlated expectedly and with marginal significance with distance from the outfall, expectedly and non-significantly with distance from Goleta Point, and expectedly with particle size.

<u>Infauna abundance patterns compared with past years.</u> Figure 6-1 illustrates biologicalmetric trends over time in the Goleta survey area during the past twenty years. The average numbers of individuals increased between 1990 and 1994 and then steadily declined through 1999. Low values during 1998 and 1999 may reflect



the El Nino conditions present then. In 2000, values began to increase through 2002 (average = 700), dipped in 2003, and then nearly doubled to historic highs during the period between 2004 and 2006 (average = 1566). Infauna abundances declined in 2007 and 2008 to levels similar to the years previous to 2004. From 2009 thru 2013, abundances remained relatively stable (average \sim 1,000). In 2014 abundances dropped to levels similar to years prior to 2004.

<u>Infauna abundance values compared with other surveys.</u> Table 6-2 compares abundance and other variables with reference control stations from the Southern California Bight Regional Monitoring Program (SCBRMP) surveys conducted in 1998, 2003 and 2008. Average numbers of individuals collected in the Goleta survey areawere greater than the averages measured at reference site locations in each of the SCBRMP surveys.

6.3.1.2. Infauna Species

Another simple measure of population health is the number of separate infauna species collected per sampling effort (i.e. one Van Veen Grab). Because of its simplicity, numbers of species is often underrated as an index. If the sampling effort and area sampled are the same for each station, however, this index can be one of the most informative. In general, stations with higher numbers of species per grab tend to be in areas of healthier communities.

<u>Spatial infauna species patterns.</u> Infauna species at the six sediment sampling stations are listed in Table 6-1. Numbers of species were similar across sites, ranging from least at station B2 (90) to greatest at station B1 (115) near Goleta Point. Numbers of species correlated expectedly and non-significantly with distance from the outfall, unexpectedly and non-significantly with Goleta Point, and expectedly and non-significantly with particle size.

<u>Infauna species patterns compared with past years.</u> Figure 6-1 illustrates biologicalmetric trends over time in the Goleta survey area during the past twenty years. Similar to numbers of individuals, numbers of species increased between 1991 and 1994 and then steadily declined through 1999 possibly owing to an El Nino effect. Since 2000 the average number of species has steadily increased through 2006 when it reached a historic high (average = 181). Since 2006 the average number of species has steadily declined thru 2014 (average = 101).

<u>Infauna species values compared with other surveys.</u> Table 6-2 compares numbers of species and other variables with reference control stations from SCBRMP surveys conducted in 1998, 2003 and 2008. Ranges for Goleta species counts were greater than ranges measured in each of the SCCWRP reference site surveys.



6.3.1.3. Infauna Diversity

Species diversity indices are similar to numbers of species; however they often contain an evenness component, as well. For example, two samples may have the same numbers of species and the same numbers of individuals. However, one station may have most of its numbers concentrated into only a few species while a second station may have its numbers evenly distributed among its species. The diversity index would be higher for the latter station. The diversity indices required in the Goleta permit are the Shannon Diversity Index, Margalef Richness Index, and Simpson Diversity Index. Since all of these indices are calculated from the same measures (numbers of individuals and numbers of species), they often show the same patterns, and are, thus, probably somewhat redundant (Table 6-1). Infauna population metrics are presented by station. Comparisons are made using correlation analysis and ANOVA.

<u>Spatial infauna diversity patterns.</u> Infauna diversities at the six sediment-sampling stations are listed in Table 6-1. Diversity, as measured by Shannon's, Margalef's, and Simpson's indiceswere similar across sites and uniformly elevated in the survey area. Each was significantly greatest at station B1 near Goleta Point by ANOVA (p < 0.05). None of the correlations with distance to the outfallwere significant for Shannon, Margalef's or Simpson's Diversity.

Infauna diversity patterns compared with past years. Figure 6-1 illustrates biologicalmetric trends over time in the Goleta survey area during the past twenty years. Shannon Diversity has been high in the Goleta survey area during the entire time period, with averages ranging between 3.5 to over 4.0. Diversity was just below 4.0 through the 1990's and then began a slight decrease to a low in 2005. In 2006 diversity began to increase thru 2007 and 2008, and reached a historic high 2009 and 2010, before decreasing again in 2011 and 2012. In 2014 diversity was similar to 2013.

Infauna diversity values compared with other surveys. Table 6-2 compares the Shannon Diversity Index reference stations from the SCBRMP surveys conducted in 1998, 2003 and 2008. Shannon Diversity measured in the Goleta survey area was similar when compared to each of the SCBRMPreference site surveys. Neither Margalef's nor Simpson's indices were calculated during the two SCCWRP programs.

6.3.1.4. Infauna Dominance

The Schwartz Dominance Index is defined as the minimum number of species required to account for 75% of the individuals in a sample. The infauna environment tends to be healthier when the dominance index is high, and it tends to correlate with species diversity.

<u>Spatial infauna dominance patterns.</u> Dominance at the six sediment-sampling stations is listed in Table 6-1. Dominance was marginally significantly different among sites by ANOVA. Dominance correlated unexpectedly and non-significantly with distance from the outfall, unexpectedly and with marginal significance with distance from Goleta Point, and expectedly and with marginal significance with sediment particle size.

<u>Infauna dominance patterns compared with past years.</u> Figure 6-1 illustrates biologicalmetric trends over time in the Goleta survey area during the past twenty



years. Dominance has been high in the Goleta survey area during the entire time period, ranging between 23 and 40. Dominance ranged between 35 and 40 through the 1990's and then began a slight decrease to a low in 2005. In 2006 dominance began to increase to an historic high (average = 36) in 2010, before decreasing back to 2005 levels in 2013 and 2014.

<u>Infauna dominance values compared with other surveys.</u> Table 6-2 compares the dominance at reference sites from the SCBRMP surveys conducted in 1998, 2003 and 2008. Dominance in the Goleta survey area in 2014 was similar to the SCBRMP reference site surveys.

6.3.1.5. Infauna Trophic Index

The Infauna Trophic Index (SCCWRP 1978, 1980) was developed to measure the feeding modes of benthic infauna. Higher values denote California species assemblages dominated by suspension feeders, which are more characteristic of unpolluted environments. Lower index values denote assemblages dominated by deposit feeders more characteristic of sediments high in organic pollutants (e.g. near major ocean outfalls). SCCWRP has also provided definitions for ranges of infauna index values. Values that are 60 or above indicate "normal" bottom conditions. Values between 30 and 60 indicate "change", and values below 30 indicate "degradation". The infauna trophic index is based on a 60-meter depth profile of open ocean coastline in southern California. Therefore, its results should be interpreted with some caution when applied to Goleta's shallower stations (24 m).

Spatial Infauna Trophic Index patterns. Infauna Trophic Index (ITI) scores at the six sediment-sampling stations is listed in Table 6-1. ITI scores were marginally significantly greatest at stations B3 and B4 near the outfall (average = 75 each). ITI values correlatedunexpectedly and non-significantly with distance from the outfall, expectedly and with marginal significance with distance from Goleta Point, and unexpectedly and non-significantly with particle size. ITI scores at all stations were well above levels defining benthic communities that are changed (60) and far above levels defining benthic communities that are degraded (30).

<u>Infauna Trophic Index patterns compared with past years.</u> Figure 6-1 illustrates biologicalmetric trends over time in the Goleta survey area during the past twenty years. Average ITI values have remained stable across years and were similar in 2014 to past surveys.

<u>Infauna Trophic Index values compared with other surveys.</u> The ITIwas not calculated for the SCBRMP (1998, 2003 and 2008). This index has been replaced as a measure of biological condition by the Benthic Response Index (BRI).

6.3.1.6 Benthic Response Index

The Benthic Response Index (BRI) measures the condition of a benthic assemblage, with defined thresholds for levels of environmental disturbance (Smith et al. 2001). The pollution tolerance of each species is assigned based upon its distribution of abundance along a pre-established environmental gradient. To give index values an ecological context and facilitate their interpretation, four thresholds of biological response to pollution were identified. The thresholds are based on changes in biodiversity along a pollution gradient. A reference threshold, below which natural



benthic assemblages normally occur, was identified at an index value of 31, the point on the pollution vector where pollution effects first resulted in a net loss of species. Three additional thresholds of response to disturbance were defined at index values of 42, 53 and 73, representing points at which 25%, 50%, and 80%, respectively, of the species present at the reference threshold were lost.

<u>Spatial BRI patterns.</u> Average BRI scores were marginally significantly greatest by ANOVA at station B1 (average = 30) compared to outfall station B5 (average = 25) (Table 6-1). BRI scores correlated unexpectedly (increased) and with marginal significance with distance to the outfall, expectedly and non-significantly with distance to Goleta Point, and with marginal significance with particle size. Scores were below 31 for each station indicating there was no net loss of reference species in the survey area. This indicates that the sites in the Goleta survey area are similar to other shallow reference site locations in the Southern California Bight.

This was the fifth year the BRI was calculated for Goleta and therefore was not compared against past survey years. The BRI was calculated using reference site data collected throughout southern California, therefore the BRI results for the 2014 survey are comparable to reference site conditions.

6.3.1.6. Cluster & Ordination Analysis

Patterns of species composition in the receiving environment's infauna community were evaluated by comparing normal (station x station) and inverse (species group x species group) classifications using the Bray-Curtis pair-wise similarity index. As Bray-Curtis Index values between station groups approach zero, the population of animals that make up the community at those sites becomes more the same. A station dendrogram was constructed from the resulting pattern matrix (Figure 6-2). For the 2014 survey, rare species were excluded from the analysis so that 170 species that occurred at > three sites were retained for analysis (94% of the total number of individuals collected).

Stations clustered into threegroups that were very similar to one another (Figure 6-2). The greatest Bray-Curtis distance between any two station nodes was approximately 30%, which indicates very small differences in species abundances and composition between sites. Station group 1 included station B6 and B2, group 2 included near outfall stations B3, B4 and B5, andgroup 3 included Goleta Point station B1.

Of the twenty most relatively abundant species collected in each cluster group, seven were shared across cluster groups, underscoring the community similarities among stations (Table 6-3). The most common species in the survey area were those typically found in coastal nearshore waters. In 2014 the seed shrimp, Euphilomedescarcharodonta, was the most relatively abundant species and was represented in each station group.

When the biological metrics for each station cluster group were averaged together they showed that the infauna population in cluster groups 1 and 2 had similar abundances, numbers of species, diversity and BRI scores. In comparison, cluster group 3 (Goleta Point station B1) had similar abundances and numbers of species, but a slightly greater BRI score and slightly lower ITI score.



6.4. Discussion

Results from this infauna survey support past studies that indicated that the ocean outfall discharge does not appear to be strongly impacting the resident benthic infauna community. This was confirmed by statistically comparing results among stations both near and far from the diffuser, comparing results with historical surveys, comparing results with other studies performed in Southern California, and comparing stations by cluster analyses.

Evaluation of the biological metrics for the 2014 survey showed that there were significant differences among sites for numbers of taxa, diversity, dominance, ITI and BRI. Each of the standard metrics (excluding ITI and BRI) were greatest at outfall station B1 andlower at the other stations. This is in contrast to the 2012 survey when the infauna populations were slightly depressed near the outfall.It appears that Goleta Point plays a role in the distribution of infauna in the Goleta survey area. This pattern of increased infauna taxa near Goleta Point may be due to the increased availability of organic material emanating from the oil seeps that are present there (Pearson and Rosenberg 1978). These results indicate the difficulty with interpreting the results of hypothesis testing on infauna abundance data. To try to elucidate these patterns and assess what, if any, impacts might be occurring to the infauna community, two indices were calculated and cluster analysis was employed.

The Infaunal Trophic Index (ITI) assesses the health of the benthic community using trophic level feeding strategies. In 2014 ITI scores at all stations were well above levels defining benthic communities that are changed (60) and far above levels defining benthic communities that are degraded (30). ITI scores in the survey area ranged from least (69) at station B1 to greatest at the outfall stations B3 and B4 (75 each). The ITI has been employed to assess the health of benthic communities since the early 1980's. However, its use to assess communities residing at depths less than 60 m has been criticized.

The averaged Benthic Response Index (BRI) scores (Smith et al. 2001) were below 31 indicating that there was no net loss of reference species in the survey area. There was an unexpected andmarginally significant correlation with distance to the outfall and among stations by ANOVA, with Goleta point station B1 having significantly greater (poorer) BRI scores compared to all outfall station B5. The BRI approach differs from other multimetric techniques in using multivariate ordination as the basis for assigning pollution tolerance scores. The primary objective of the BRI is to assign pollution tolerance scores to species based on their position on a pollution gradient. Once assigned, the scores can be used to assess the condition of the benthic community. The BRI was developed using hundreds of infauna samples collected from throughout the southern California bight, at sites that were both degraded and in reference condition.

Biological metrics calculated for the 2014 survey were compared to results of past surveys at the same sampling locations since 1990. Each of the metrics measured in 2014 were within the ranges of past surveys.

Cluster analysis showed that the dissimilarity among both station and species groups were very low across the survey area. The three station clusters identified were at most 30% different from one another based on infauna abundances and taxa composition. Of the top twenty most abundant species in the survey area,



sevenwere shared by the three cluster groups, underscoring the community similarities among stations.

To further investigate the potential influence of the Goleta outfall on the infauna community, cluster analysis and ordination were conducted on infauna data sets collected from 2004 to 2014 (Figure 6-3). Ordination analysis showed that the largest portion of the variation in the infauna community during the time period could be described by ordination axis 1 (27%) which was closely associated with survey year. Stations clustered together on axis 1 by year with 2004 thru 2010 infauna communities (cluster groups 1, 2 and 3) furthest from stations collected during 2011 thru 2014 (cluster groups4 and 5). This indicates that larger oceanographic conditions are defining the abundances and composition of species in the survey area. There was no clear outfall related gradient on either axis 1 or axis 2 which described 11% of the variation in the community.

The biological metrics for each site and survey were averaged by historic cluster group and showed there was very little difference across cluster groups indicating a relatively stable infauna population through time (Table 6-5). This was especially true of numbers of species which ranged from 238 to 287, Shannon Diversity which ranged from 3.9 to 4.2 and the BRI which ranged from 27 to 31.

Finally, Goleta results were compared to measurements made of the inner continental shelf throughout southern California. All infauna population variables were comparable to or greater than those measured in regional surveys conducted by the SCBRMP in 1998, 2003 and 2008.

Although there are no specific numerical limitations regarding infauna animals, the California Ocean Plan (SWRCB 2007) states that:

The rate of deposition of inert solids and the characteristics of inert solids in the ocean shall not be changed such that benthic communities are degraded.

The dissolved sulfide concentration of waters in and near sediments shall not be significantly increased above that present under natural conditions.

The concentration of substances set forth in Chapter IV, Table B, in marine sediments shall not be increased to levels which would degrade indigenous biota.

The concentration of organic materials in marine sediments shall not be increased to levels which would degrade marine life.

Nutrient materials shall not cause objectionable aquatic growths or degrade indigenous biota.

Marine communities, including vertebrate, invertebrate, and plant species, shall not be degraded.

Waste management systems that discharge to the ocean must be designed and operated in a manner that will maintain the indigenous marine life and a healthy and diverse marine community.



Waste discharged to the ocean must be essentially free of: "2) Settleable material or substances that may form sediments which will degrade benthic communities or other aquatic life."

Waste discharged to the ocean must be essentially free of: "3) Substances which will accumulate to toxic levels in marine waters, sediments or biota."

Based upon spatial and temporal comparisons and analogies with other studies, the results of the infauna survey indicate that the discharge is in compliance with the general limitations and that it causes no adverse impact.



Table 6-1. Infauna population indices by replicate for each of the six Goleta survey area stations. Comparisons are made using correlation analysis and ANOVA (p < 0.05).

			Winhara Okali			
Constituent	B4		Offshore Statio		05	D0
Constituent	B1	B2	B3	B4	B5	B6
						<u> </u>
INDIVIDUALS 1.						
Repl. 1	556	444	475	403	483	1033
Repl. 2	812	602	442	576	655	546
Repl. 3	614	483	636	608	446	752
Repl. 4	411	639	651	558	641	537
Repl. 5	623	615	548	537	409	609
Mean =	603	557	550	536	527	695
Std. Dev. =	144	87	93	79	114	207
Lower Conf. Int. =	477	480	469	467	427	514
Upper Conf. Int. =	730	633	632	606	627	877
Overail Mean =		r (outfall)=		r (point) =		r (prt.sz.) = -0.05
Overall S.D. =		, ,	0.41	Comp. of means		· (p/1.02.)
313/23/3.2.	100,0		0.41	COTTP: OF TECHNO		
SPECIES ¹						
Repl. 1	106	82	74	86	78	121
·						
Repl. 2	123	89	82	106	118	91
Repl. 3	113	86	106	102	94	114
Repl. 4	107	93	108	101	118	96
Repl. 5	128	102	107	105	105	89
Mean =	115	90	95	100	103	102
Std. Dev. =	10	8	16	8	17	14
Lower Conf. Int. =	107	84	81	93	88	90
Upper Conf. Int. =	124	97	110	107	118	115
Overall Mean =	101.0	r (outfall) =	0.21	r (point) =	-0.17	r (prt.sz.) = 0.22
Overall S.D. =	14.0	F=	2.1 9	Comp. of means	= NA	
SHANNON DIVERSITY						
Repl. 1	3.79	3.28	3.24	3.57	3.28	3.45
Repl. 2	3.62	3.41	3.5	3.66	3.71	3.33
Repl. 3	3.79	3.36	3.51	3.51	3.49	3.51
Repl. 4	3.99	3.47				
·			3.54	3.59	3.83	3.64
Repl. 5	44	3.52	3.66	3.67	3.78	3.29
Mean =	3.84	3.41	3.49	3.60	3.62	3.44
Std. Dev. =	0.16	0.09	0.15	0.07	0.23	0.14
Lower Conf. Int. =	3.70	3.33	3.36	3.54	3.42	3.32
Upper Conf. Int. =	3.98	3.49	3.62	3.66	3.82	3.57
Overall Mean =		r (outfall) =		r (point) =	_	r (prt.sz.) = 0.42
Overall S.D. =		, ,	5.50	Comp. of means		
	U.EU		0.00	Comp. or means	- DI - DZ, D	, LA
MADOAL DE DIO NECO						
MARGALEF RICHNESS	40.04	42.00	44.04	44.47	40.40	47.00
Repl. 1	16.61	13.29	11.84	14.17	12.46	17.29
Repl. 2	18.21	13.75	13.30	16.52	18.04	14.28
Repl. 3	17.45	13.75	16.27	15.76	15.25	17.06
Repl. 4	17.61	14.24	16.52	15.81	18.10	15.11
Repl. 5	19.74	15.73	16.81	16.55	17.29	13.73
Mean =	17.92	14.15	14.95	15.76	16.23	15.49
Std. Dev. =	1.16	0.94	2.24	0.97	2.40	1.62
Lower Conf. Int. =	16,90	13.33	12.99	14.91	14.12	14.08
Upper Conf. Int. =	18.94	14.98	16.91	16.61	18.34	16.91
Overall Mean =		r (outfall) =		r (point) =		r (prt.sz.) = 0,26
		. ,				(pr52) - 0,20
Overall S.D. =	1.92	F=	2.96	Comp. of means	s = B1 > B2	

Bold = Marginally Significant (0.05 < p < 0.10)

Bold & Gray = Significant (p < 0.05)

- 1. The van Veen Grab collects samples one tenth of one square meter in area. To determine individuals per meter, multiply by ten.
- 2. Non-normal data: correlation coefficients and ANOVA's from non-parametric tests (Spearman's rho and Kruskal-Wallace H, respectively).



Table 6-1. continued

			Offshore Statio	nne -		
Constituent	B1	B2	B3	яь В4	B5	B6
557,54,145,11						50
SIMPSON DIVERSITY						
Repl. 1	0.97	0.92	0.94	0.95	0,93	0.93
Repl. 2	0.94	0.94	0.95	0.95	0.95	0.93
1	0.96		0.95			
Repl. 3		0.94		0.94	0.93	0.95
Repl. 4	0.97	0.95	0.95	0.96	0.96	0.96
Repl. 5	0.97	0.94	0.95	0.96	0.97	0.94
Mean =	0.96	0.94	0.95	0.95	0.95	0.94
Std. Dev. =	0.01	0.01	0.01	0.01	0.02	0.01
Lower Conf. Int. =	19.20	22.03	20.29	19.88	24.65	21.34
Upper Conf. Int. =	26.20	24.95	25.23	21.70	27.68	22.39
Overall Mean =	0.948	r (outfall) =	-0.04	r (point) =	-0.37	r (prt.sz.) = 0.30
Overall S.D. =	0.014	F=	2.4	Comp. of means	= NA	
SCHWARTZ DOMINANCE						
Repl. 1	30	18	16	24	18	21
Repl. 2	28	21	22	27	28	18
Repl. 3	30	20	21	22	23	22
Repl. 4	36	23	25	25	31	27
1						
Repl. 5	38	22	26	26	32	16
Mean =	32	21	22	25	26	21
Std. Dev. =	4	2	4	2	6	4
Lower Conf. Int. =	19	22	20	20	25	21
Upper Conf. Int, =	26	25	25	22	28	22
Overall Mean =		r (outfall) =		r (point) =		r (prt.sz.) = 0.46
Overall S.D. =	5.49	F=	6.39	Comp. of means	= B1 > B2, B3	3, B6
	T					
INFAUNAL INDEX	70		7.4	70	70	74
Repl. 1	70 64	68 71	74 76	76 76	73 71	71 73
Repl. 2 Repl. 3	68	71 72	76 73	76 74	71 74	73 73
Repl. 4	72	72 70	73 77	71	73	73 77
Repl. 5	71	70	74	76	77	73
Mean =	69	70	75	75	74	73
Std. Dev. =	3.0	1.7	1.6	2.5	1.9	2.3
Lower Conf. Int. =	19	22	20	20	25	21
Upper Conf. Int. =	26	25	25	22	28	22
Overall Mean =		r (outfall) =		r (point) =		r (prt.sz.) = -0.58
Overall S.D. =	3.00	F=	5.84	Comp. of means	= B3, B4 > B	1, B2; B5 > B1
BENTHIC RESPONSE INDEX						<u> </u>
Repl. 1	28	25	28	28	23	30
Repl. 2	35	30	25	26	27	26
Repl. 3	29	28	30	26	24	28
Repl. 4	31	29	26	28	28	28
Repl. 5	28	30	30	29	26	30
Mean =	30	28	28	27	25	29
Std. Dev. = Lower Conf. Int. =	2.8 19	2.1 22	2.3 20	1.3	2.1 25	1.8
Upper Conf. Int. =	26	22 25	20 25	20 22	25 28	21 22
Overall Mean =		r (outfall) =		r (point) =		r (prt.sz.) = -0.15
Overall S.D. =		, ,	2.5	Comp. of means		· (prc32.) = -0.10
2.5.25.0.					~	

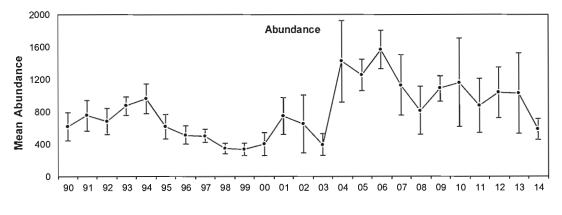
Bold = Marginally Significant (0.05

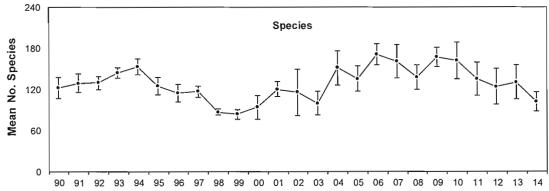
Bold & Gray = Significant (p < 0.05)

- 1. The van Veen Grab collects samples one tenth of one square meter in area. To determine individuals per meter, multiply by ten.
- Non-normal data: correlation coefficients and ANOVA's from non-parametric tests (Spearman's rho and Kruskal-Wallace H, respectively).



Figure 6-1. Infauna community variables, station (n = 6) means and standard deviations since 1990.





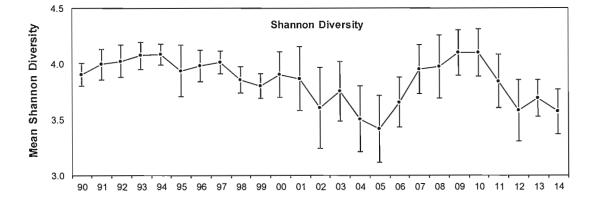
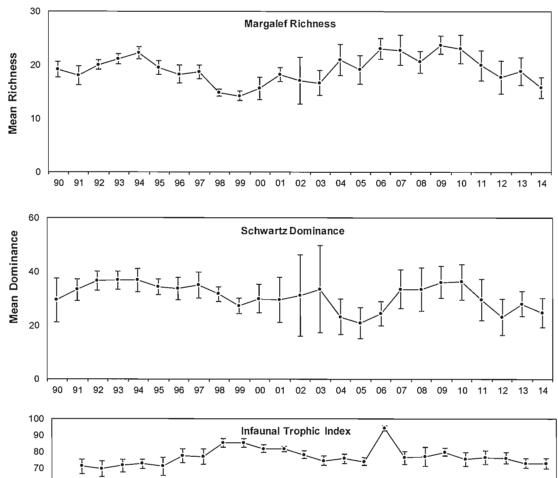




Figure 6-1. (continued).



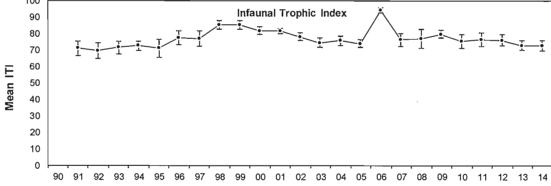




Table 6-2. Comparison of Goleta infauna variables with results from other studies (per 0.1 m^2).

	G	oleta 2012	so	BRMP 1998		MP 2003 Shelf	SCBRM Inner	
Variable	Mean	Range	Mean	Range	Mean	±95% CI	Mean	SE
Number of Individuals Number of Species	578 101	403 - 1033 74 - 128	385 85	35 - 1696 18 - 162	283 62	30 5	346 85	22 4
Shannon Diversity Index Dominance	3.6 24.5	3.2 - 4.0 16.0 - 38.0	3.60	2.00 - 4.40	3.48 23	0.09 2	3.63 27	0.06
Dominance	24.5	16.0 - 38.0	_		23	2	27	1

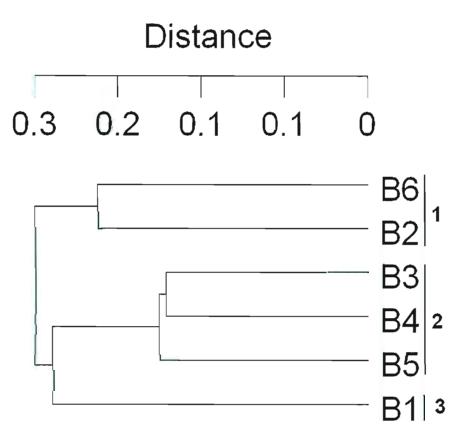


Figure 6-2. Station dendrogram based on cluster analysis (UPGMA, Sneath and Sokal 1973). The Bray-Curtis dissimilarity index was used to calculate the distances among stations and species (Gauch 1982, Jongman et. al. 1995).



Table 6-3. Average abundances of the top twenty species for each cluster group in 2014.

	Cl	uster Gro	цр
Species	1	2	3
	4.5=		
Euphilomedes carcharodonta	167	177	22
Cossura sp A	99	64	24
Mediomastus sp	78 76	100	
Poecilochaetus martini	76	4.47	
Glottidia albida	74	147	
Ampelisciphotis podophthalma	65	18	
Tellina modesta	64	116	8
Spiophanes duplex	46	102	27
Amphiodia sp	21	27	8
Nuculana taphria	21		
Macoma yoldiformis	20	42	11
Monticellina cryptica	17		14
Parvilucina tenuisculpta	15	19	23
Paraprionospio alata	15		10
Leptochelia dubia Cmplx	15		
Rhepoxynius stenodes	14	38	
Spiochaetopterus costarum Cmplx	14	20	
Pectinaria californiensis	14		
Magelona berkeleyi	13	19	7
Foxiphalus golfensis	12		
Prionospio (Prionospio) jubata		32	46
Levinsenia gracilis		24	31
Foxiphalus obtusidens		21	15
Platynereis bicanaliculata		20	
Leptosynapta sp		19	16
Ampelisca cristata cristata		19	
Arachnanthus sp A		18	8
Mediomastus acutus			64
Oligochaeta			11
Monticellina siblina			6
Phoronis sp			6
Kurtiella tumida			5



Table 6-4. Biological metrics for each station in 2014 averaged by cluster group.

Station	Cluster Group	Number of Species	Total Abundance	BRI	ιπ	Evenness	Margalef Richness	Schwartz Dominance	Shannon Diversity	Simpson Diversity
B2	1	199	556	27.78	70.1	0.707	31.32	24	3,74	0.947
B6	1	211	695	27,99	73.1	0.689	32,088	25	3.69	0.95
	average	205	626	27.9	72	0.698	31.704	25	3.72	0.949
В3	2	182	550	27.82	74.9	0.719	28.682	26	3.74	0.954
B4	2	214	536	26.54	74.3	0.731	33.891	32	3.92	0.956
B5	2	207	526	26.26	73.2	0.737	32.872	32	3.93	0.957
	average	201	537	26.9	74	0.729	31.815	30	3.86	0.956
В1	3	231	603	29.29	68.6	0.775	35.925	46	4,22	0.968

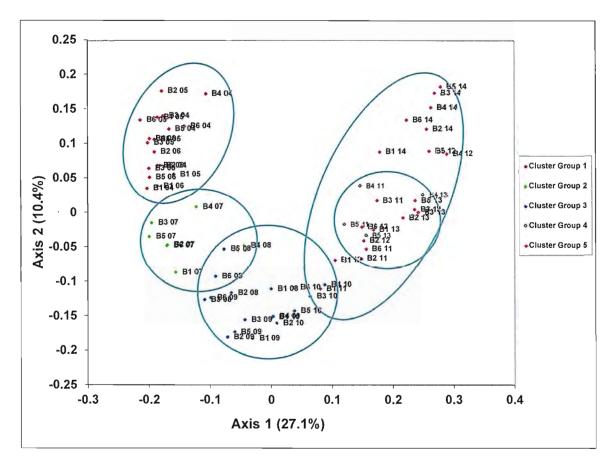


Figure 6-3. Plot of ordination scores for infauna communities at stations measured from 2004 to 2014.



Table 6-5. Biological metrics for each station for each year individually from 2004 thru 2014 and averaged by cluster group.

ragea		ster gr	oup.			
01-11	Cluster	Number of	Total			Shannon
Station/Year	Group	Species	Abundance	BRI	ITI	Diversity
B1 04	1	369	2159	32.86	72.8	3.62
B1 05	1	315	1246	29.94	78.3	3.92
B1 06	1	302	1386	30.05	74.3	3,8
B2 04	1	330	1616	30.61	76.6	3.88
B2 05	1	246	1302	32.94	71.6	3.12
B2 06	1	300	1580	31.59	72.7	3.43
B3 04	1	247	1430	30.71	74	3.31
B3 05		284	1499	30.5	74.3	3.54
	1					
B3 06	1	299	1775	31,05	73.6	3.66
B4 04	1	238	1132	31.31	71.5	2.99
B4 05	1	261	1112	31.51	71	3.25
B4 06	1	291	1580	32.24	71.5	3.48
B5 04	1	257	1220	27.89	74.9	3.47
B5 05	1	300	1221	29.73	73.5	3.53
B5 06	1	312	1804	29.95	72.1	3.65
B6 04	1	258	945	26.56	78.9	3.8
B6 05		265	1124	30.7	74.9	3,43
	1					
B6 06	1	290	1270	30.47	73.9	3.61
	Average	287	1411	30.6	73.9	3,53
B1 07	2	317	1022	31.05	75.9	4.33
B2 07	2	248	729	31.66	79.8	4.25
B3 07	2	261	1400	32.83	73.4	3.72
B4 07	2	247	1023	31.98	71.2	3.84
B4 08	2	237	854	32.26	63.4	3.95
B5 07		278	1220	31.18	75.2	3.95
	2					
B6 07	2	315	1349	30.53	78.3	3.93
	Average	272	1085	31.6	73.9	4
		_	_			
B1 08	3	252	582	26.13	80.6	4.53
B1 09	3	311	1203	28.13	77.9	4.17
B1 10	3	299	1210	28.09	74.5	4.26
B2 08	3	224	677	30.64	80.5	4.35
B2 09	3	287	1024	29.82	76.4	4.31
B2 10		291	920	27.35	75	4.42
B3 08	3	261	1093		77.2	4.06
	3			30.21		
B3 09	3	275	1124	29.24	79.9	4.05
B3 10	3	275	985	28.33	75.3	4.35
B4 09	3	247	950	27.17	80.5	3.98
B4 10	3	268	995	30.39	71.5	4.13
B5 08	3	247	741	29.43	76.6	4.08
B5 09	3	290	1154	27.16	81.4	4
B5 10	3	349	1973	28.6	79.6	4.05
B6 08	3	258	910	27.79	76.1	4.01
						4.36
B6 09	3	265	1037	26.75	80.8	
B6 10	3	265	855	26.27	79.3	4.26
	Average	274	1025	28.3	77.8	4.2
B4 11	4	241	735	29.42	72.7	4.12
B4 13	4	266	1047	28.92	74.1	4.2
B5 11	4	242	738	31.75	75.1	4.15
B5 13	4	300	1796	32.15	73.1	3.85
	Average	262	1079	30.6	73.8	4.08
B1 11	5	324	1343	28.7	72.8	4.22
B1 12		332	1457	28.69	72.8	4.34
B1 13	5					4.34
	5	277	956	28.84	71.1	
B1 14	5	231	603	29.29	68.6	4.22
B2 11	5	241	973	28.02	76.7	3.94
B2 12	5	261	1179	28.82	73.7	4.04
B2 13	5	250	936	28.38	69.2	3.93
B2 14	5	199	557	27.78	70.1	3.74
B3 11	5	228	816	26.5	82.1	3.81
B3 12	5	257	1114	27.96	75.3	3.94
B3 13		244	844	27.7	72.3	4.09
	5					
B3 14	5	182	550	27.82	74.9	3.74
B4 12	5	214	773	25.48	80.9	3.7
B4 14	5	214	536	26.54	74.3	3.92
B5 12	5	222	830	24.38	78.9	3,65
B5 14	5	207	527	26.26	73.2	3.93
B6 11	5	244	617	26.24	77.3	4.29
B6 12	5	228	842	28.77	73.5	3.68
B6 13	5	203	572	25.83	74.2	3,99
B6 14	5	211	695	27.99	73.1	3.69
	Average	238	836	27.5	74.3	3.96
	-					



CHAPTER 7

Trawled Fish and Invertebrate Populations

7.1. Background

Demersal fishes and megabenthic invertebrates (species living closely associated with the seafloor) are widely distributed on the soft-bottom habitats along the southern California shelf. This diverse community is composed of approximately 100 species of fish and several hundred species of invertebrates (Allen 1982, Allen et al. 1998, Moore and Mearns 1978). Since these populations are generally sedentary, they can act as indictors of human impacts on the soft bottom habitat. As a result, trawl programs have been part of the monitoring activities of both large and small municipal dischargers for nearly thirty years. The goal of the Goleta Sanitary District's trawl program is to look for population changes in the vicinity of the ocean outfall.

7.2. Materials and Methods

Trawl sampling was conducted in accordance with *Use of Small Otter Trawls in Coastal Biological Surveys*, EPA 600/3-78/083, August 1978; *Quality Assurance and Quality Control (QA/QC) for 301(h) Monitoring Programs: Guidance on Field and Laboratory Methods*, Tetra Tech 1986; and the *Southern California Bight Project Field Operations Manual*, 2008. Duplicate ten-minute trawls were taken at a uniform speed of 2.0 - 2.5 knots with a 7.6 m Marinovich otter trawl. Care was taken to not trawl over previous transects or grab sampling sites. For each trawl, all fish and macroinvertebrates were identified, counted, measured, and weighed. Collection observations, such as algae or cobble in the trawl, were recorded. Fish abnormalities, such as fin rot, parasites, or tumors, were also noted. Species abundance lists were compiled for all trawl samples. All fish and invertebrates were identified by Jim Mann and Karin Patrick. All animals collected for tissue dissection were placed in plastic zip-lock bags in coolers over ice during transit.

Following enumeration of trawl organisms by species, the total and animal group biomasses, numbers of individuals, and numbers of separate species were compiled for each station replicate. In addition, several required biological indices were calculated: Shannon-Weiner species diversity (H'), Margalef's richness index (d), Simpson's species diversity (SI), and Schwartz's dominance (D). These indices are described in detail in Chapter 6, in Section 6.2, Materials and Methods. Since there were only two stations sampled, no clustering or numerical classification analyses could be calculated. Stations were compared by t-test (see Materials and Methods section above).

7.3. Results

The demersal fish and macrobenthic invertebrate community was compared among two trawl stations using measures of population abundance and diversity. These included numbers of individuals, numbers of species, species diversity, and species dominance. In addition, ranges of these variables were compared to surveys conducted in past years. Duplicate trawls were taken at two locations, one near Station B3 (TB3) and the other near Station B6 (TB6) (Figure 6-1).



7.3.1. Trawled Fish

7.3.1.1. Fish Community Metrics

The averaged fish community metrics and biomass for replicate trawls are presented in Table 7-1, with results by replicate presented in Appendix 10.7 (Tables 10-9 and 10-10). A total of 435 individual fish were collected from both stations combined during the 2014 survey, with the average numbers of individuals at TB3 (96) less than the average numbers collected at TB6 (122) (Table 7-1). There was no statistically significant difference in average abundances between sites (p > 0.05; Table 7-1). The average numbers of species collected at the two stations was the same (11). Average biomass was half that at TB3 (1.23 Kg) compared to TB6 (2.21 Kg), but there was no significant difference between sites. Shannon Diversity, Simpsons Diversity, Margalet's Richness and Dominance were low at each site and were not significantly different between sites.

7.3.1.2. Species Composition

As with past years, the fish caught in the 2014 trawls were typical of those found on most southern California near shore soft bottom habitats (Table 7-2). A total of 16 and 15 unique taxa were collected at stations TB3 and TB6, respectively. The most abundant species collected in the Goleta survey area was the speckled sanddab (Citharichthys stigmaeus), which also dominated the catch at TB3 (n = 63). The next most abundant species at this site California lizard fish ($Synodus\ lucioceps$, n = 8). At TB6 pink seaperch ($Zalembius\ rosaceus$, n = 30) were nearly as abundant as speckled sanddabs (n = 34), followed by kelp bass ($Paralabrax\ clathratus$, n = 13), shiner perch ($Cymatogaster\ aggregate$, n = 13) and giant kelpfish ($Heterostichus\ rostratus$, n = 12).

7.3.1.3. Fish Community Metrics Compared to Past Surveys

Fish assemblage community metrics for 2014 were compared to previous Goleta area surveys starting in 1991 (Figure 7-1). The numbers of individuals collected in 2014 was within the range of past surveys. Fish biomass was again very low during 2014 and similar to the past 20 years. Numbers of species was similar in 2014 compared to past years. In 2014, Shannon Diversity and dominance were low and similar to past surveys.

7.3.1.4. Fish Community Metrics Compared to Reference Surveys

Fish community metrics for the 2014 Goleta survey were compared to fish assemblage data collected in the northern region on the inner continental shelf in the southern California bight during the 2008 Southern California Bight Regional Monitoring Survey (SCBRMP) (SCCWRP 2011; Table 7-3). Number of individuals, number of species, Shannon Diversity and biomass were all well within the range fish assemblages found in the vicinity of the northern region inner shelf.

7.3.1.5. Fish Length

<u>Fish size class distributions</u>. The size frequency distributions for all fish collected from trawl samples are presented in Appendix 10.7 (Table 10.7-1). The size frequency distributions for one of the historically most abundant species in the survey area (speckled sanddabs, *Citharicthys stigmaeus*) are presented in Figure 7-2. Across years, sanddab lengths ranged from 3 to 13 cm at both stations, with 2014 having slightly more individuals in the 6 cm size class at both stations. At TB3, near the outfall, the numbers of fish collected were relatively evenly spread across size classes for all years, except in 2007 and 2012 when large numbers of individuals in the 7 and 8 cm size classes were captured. The majority of sanddabs collected 2004, 2007, 2009 and 2012 at TB6 were in 6 to 8 cm size classes.



Table 7-1. Trawled fish - Summary of biological metrics of fish collected at Stations TB3 and TB6. Comparison between sites by two sample T-test (p < 0.05).

		Fis	sh			
Station	TE	33	TE	36	T-t	est
Metric	Avg	SD	Avg	SD	t score	p =
Individuals	96	8	122	33	-1.12	0.38
Species	11	3	11	1	0.24	0.83
Biomass (kg)	1.23	0.49	2.21	1.70	-0.78	0.52
Shannon Diversity	1.28	0.37	1.54	0.32	-0.74	0.54
Simpson Diversity	0.53	0.14	0.66	0.15	-0.93	0.45
Margalef Richness	2.19	0.58	1.98	0.26	0.43	0.71
Schw artz Dominance	3	1	3	1	-0.45	0.70

Bold - Marginally Significant (0.05 < p < 0.10)

Bold - Significant (p < 0.05)

1. Non-normal data: T-test by Mann-Whitney U test.

Table 7-2. Trawled fish abundance and biomass sorted from most to least abundant.

	Trawl TB3				Trawl TB6		
Scientific Name	Common Name	Mean Abundance	Mean Weight (kg)	Scientific Name	Common Name	Mean Abundance	Mean Weight (kg)
Citharichthys stigmaeus	speckled sanddab	63	0.29	Citharichthys stigmaeus	speckled sanddab	34	0.17
Synodus lucioceps	California lizardfish	8	0.45	Zalembius rosaceus	pink seaperch	30	0.50
Citharichthys sordidus	Pacific sanddab	6	0.08	Paralabrax clathratus	kelp bass	13	0.06
Pleuronichthys decurrens	curlfin sole	5	0.23	Cymatogaster aggregata	shiner perch	13	0.54
Heterostichus rostratus	giant kelpfish	4	<0.1	Heterostichus rostratus	giant kelpfish	12	<0.1
Syngnathus califomiensis	kelp pipefish	4	<0.1	Synodus lucioceps	California lizardfish	8	0.53
Hypsurus caryi	rainbow seaperch	1	<0.1	Hypsurus caryi	rainbow seaperch	6	0.14
Icelinus quadriserialus	yellowchin sculpin	1	<0.1	Syngnathus californiensis	kelp pipefish	3	< 0.1
Zalembius rosaceus	pink seaperch	1	<0.1	Gibbonsia montereyensis	crevice kelpfish	2	<0.1
Odontopyxis trispinosa	pygmy poacher	1	<0.1	Paralabrax nebulifer	barred sand bass	1	0.07
Phanerodon furcalus	white seaperch	1	<0.1	Pleuronichthys coenosus	C-O sole	1	0.07
Sebastes auriculatus	brown rockfish	1	<0.1	Pleuronichthys decurrens	cuttin sole	1	<0.1
Sebastes miniatus	vermilion rockfish	1	<0.1	Porichthys notatus	plainfin midshipman	1	<0.1
Symphurus atricaudus	California tonguefish	1	<0.1	Sebastes dallii	calico rockfish	1	<0.1
Xystreurys liolepis	fantail sole	1	0.06	Squatina californica	Pacific angel shark	1	0.16
Gibbonsia evides	spotted kelpfish	1	<0.1		composite weight*		<0.1
	composite weight*		0.13		,		

^{*}Species <0.1 kg are weighed together as a composite weight.



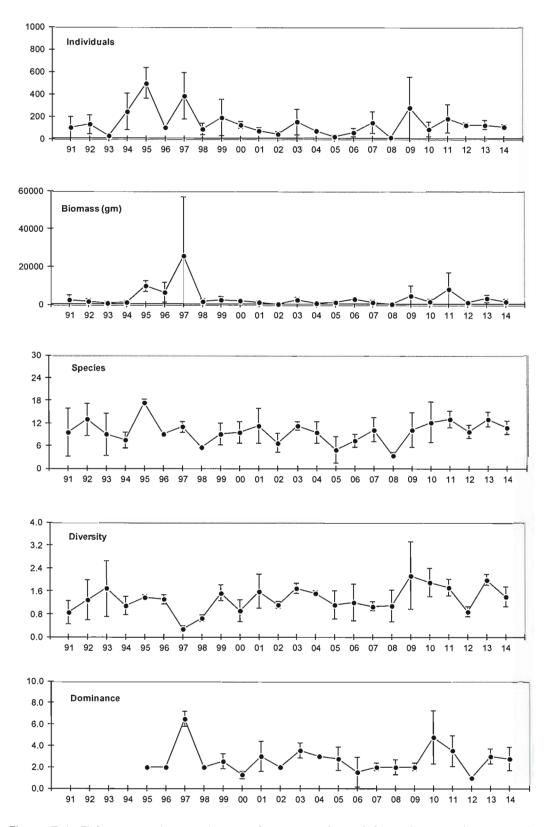


Figure 7-1. Fish community metric annual averages (\pm SD) for Goleta trawl transect data (n=2) since 1991.

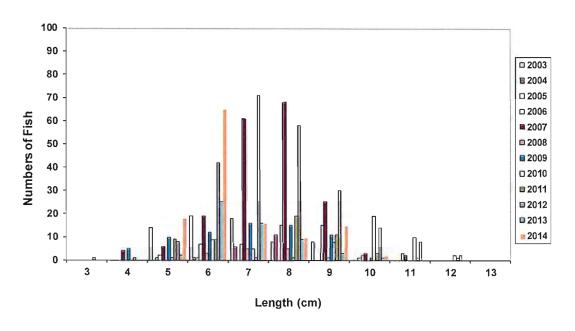


Table 7-3. Comparison of trawl fish metrics with results from the Southern California Regional Survey, Bight 2008 (SCCWRP 2011).

		Trawl Fish	
Metric	Goleta Range	Bight '08 Northern Region Inner Shelf	Below Range?
Biomass (kg)	1.23 - 2.21	0.7 - 4.7	No
Individuals	96 - 122	24 - 467	No
Species	11 - 11	5 - 22	No
Shannon Diversity	1.28 - 1.54	0.5 - 2.31	No



Station TB3



Station TB6

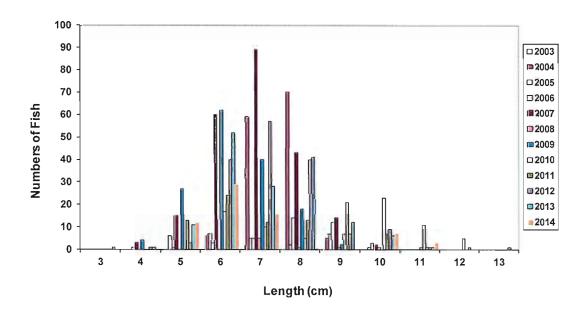


Figure 7-2. Length (cm) frequency distributions for speckled sanddabs (*Citharicthys stigmaeus*) collected from 2003 to 2014 from stations TB3 and TB6 in the Goleta survey area.



7.3.2. Trawl Macroinvertebrates

7.3.2.1. Macroinvertebrate Community Metrics

The averaged macroinvertebrate community metrics and biomass for replicate trawls are presented in Table 7-4, with results by replicate presented in Appendix 10.7 (Tables 10-11 and 10-12). A total of 67 individual invertebrates were collected from both stations combined during the 2014 survey. An average of 22 macroinvertebrates was collected at station TB3 compared to 12 at TB6 and there was a significant difference between sites (Table 7-4). Numbers of species collected averaged 3 at station TB3 and 6 at station TB6, with no significant difference between sites. Biomass was 0.77 Kg at TB3 and 1.24 Kg at TB6 and there was no significant difference. Shannon Diversity, Simpson Diversity and Margalef Richness were low at both stations and there were no significant differences between sites. Dominance was 2 at TB3 and 4 at TB6 and was significantly different.

7.3.2.2. Species Composition

As with past years, the invertebrates in the 2014 trawls were typical of those found on most southern California near shore soft bottom habitats (Table 7-5). A total of 10 unique taxa were collected in the survey area. The most abundant species collected in the survey area were the peanut rock shrimp (Sicyonia penicillata) and sheep crab (Loxorhynchus grandis).

7.3.2.3 Macroinvertebrate Community Metrics Compared to Past Surveys

Macroinvertebrate community metrics for 2014 were compared to previous Goleta area surveys starting in 1991 (Figure 7-2). The numbers of individuals and average biomass in 2014 was similar to previous surveys. Numbers of species, Shannon Diversity and Dominance were also similar to the previous decade. These three metrics declined in 1998 from historic highs and have been relatively stable since. The reasons for these reductions are unclear.

7.3.2.4. Macroinvertebrate Community Metrics Compared to Reference Surveys

Macroinvertebrate community metrics for the 2014 Goleta survey were compared to invertebrate assemblage data collected in the northern region on the inner continental shelf in the southern California bight during the 2008 Southern California Bight Regional Monitoring Survey (SCBRMP) (SCCWRP 2011; Table 7-6). Biomass, numbers of individuals, numbers of species and Shannon Diversity were all within the range or fish assemblages found in the northern region inner shelf.



Table 7-4. Trawled inverts - Summary of biological metrics of invertebrates collected at Stations TB3 and TB6. Comparison between sites by two sample T-test (p > 0.05).

		Inverteb	rates		-1:12	
Station	TE	33	TE	36	T-to	est
Metric	Avg	SD	Avg	SD	tscore	p =
Individuals	22	1	12	1	8.50	0.01
Species	3	1	6	0	-3.00	0.10
Biomass (kg)	0.77	0.86	1.24	0.01	-0.79	0.51
Shannon Diversity	0.62	0.43	1.70	0.03	-3.46	0.07
Simpson Diversity	0.33	0.23	0.80	0.01	-2.93	0.10
Margalef Richness	0.65	0.45	2.01	0.10	-4.18	0.05
Schw artz Dominance ^{1.}	2	1	4	0	-5.00	0.04

Table 7-5. Trawled invertebrate abundance and biomass sorted from most to least abundant.

5 3 5 5 6 10	Trawl TB3		1		Trawl TB6		
Scientific Name	Common Name	Mean Abundance	Mean Weight (kg)	Scientific Name	Common Name	Mean Abundance	Mean Weight (kg)
Sicyonia penicillata Octopus rubescens Loxorhynchus grandis	peanut rock shrimp red octopus sheep crab	17 2 2	0.13 <0.1 0.55	Loxorhynchus grandis Sicyonia penicillata Aplysia califomica	sheep crab peanut rock shrimp purple sea hare	4 3 1	0.72 <0.1 0.40
Metacarcinus gracilis	graceful rock crab	1	0.09	Conus californicus Crangon nigromaculata Lytechinus pictus	California cone blackspotted bay shrimp white sea urchin	1 1 1	<0.1 <0.1 <0.1
				Octopus rubescens Pugettia producta Pyromaia tuberculata	red octopus northem kelp crab tuberculate pear crab composite weight*	1 1 1	<0.1 0.13 <0.1 <0.1

^{*}Species <0.1 kg are weighed together as a composite weight.



 $[\]begin{array}{l} \mbox{Bold - Marginally Significant (0.05$

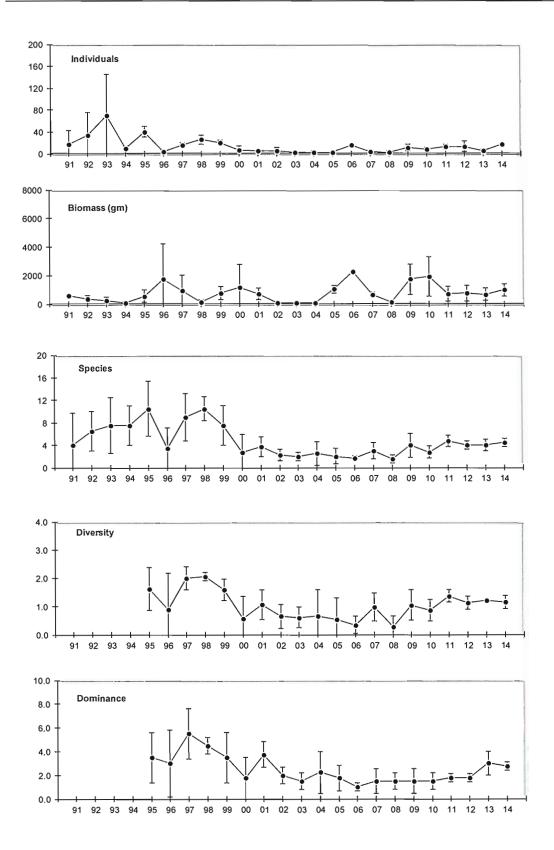


Figure 7-2. Invertebrate community metric annual averages (\pm SD) for Goleta trawl transect data (n=2) since 1991.



Table 7-6. Comparison of trawl invertebrate metrics with results from the Southern California Regional Survey, Bight 2008 (from SCCWRP, 2011).

	Tra	wl Invertebrate	
Metric	Goleta Range	Bight '08 Northern Region Inner Shelf	Below Range?
Biomass (kg)	0.77 - 1.24	0.0 - 3.0	No
Individuals	12 - 22	3 - 135	No
Species	3 - 6	2 - 20	No
Shannon Diversity	0.62 - 1.70	0.64 - 2.30	No



7.4. Discussion

Results from this trawl survey support past studies that indicated that the discharge from the Goleta Sanitary District's ocean outfall does not appear to be impacting the resident fish or macroinvertebrate communities. This was confirmed by comparing results among stations both near and far from the diffuser, comparing results with historical surveys, and comparing results with other studies being performed in southern California.

A total of 435 individual fish and 67 individual invertebrates were collected from both stations combined during the 2014 survey. There were no statistically significant differences (p < 0.05) between stations near to and far from the outfall when metrics for fish total abundance, number of species, biomass, diversity and dominance were compared. For invertebrates each of these metrics was similar between sites, except for abundance and dominance which were significantly different between sites. Individual invertebrates were nearly twice as abundant at TB3 (n = 22) compared to TB6 (n = 12). Conversely, dominance at TB6 (4) was twice that at TB3 (2). Both fish and invertebrate population indices measured in 2014 (including abundance, numbers of species and biomass) were within the range of reference sites sampled during the 2008 Southern California Bight Regional Monitoring Program.

As with past years, the fishes and macroinvertebrates caught in the 2014 trawls were typical of those found on most southern California near shore soft bottom habitats. A total of 16 and 15 individual fish taxa were collected at stations TB3 and TB6, respectively. The most abundant species collected at station TB3 and TB6 was the speckled sanddab (*Citharichtys stigmaeus*). A total of 10 unique invertebrate taxa were collected in the trawl area. The most abundant species collected in the survey area were the peanut rock shrimp (*Sicyonia penicillata*) and sheep crab (*Loxorhynchus grandis*).

When the 2013 trawled fish and invertebrate results were compared against past surveys, average abundances, numbers of species, biomass, diversity and dominance were within the ranges of the previous twenty years. This was especially true of the trawled fish community. In contrast, the trawled invertebrate community has been very similar for each biological metric over the past ten years, but prior to 2001 the numbers of invertebrate taxa and diversity were much greater. The reasons for the decrease in trawled invertebrate diversity are unclear. Since an outfall related impact has never been detected, it is probable that some larger oceanographic condition has influenced this community. Frequent cold water upwelling events which are typical of this coastal region, coupled with warm water El Nino events over the past 15 years may be playing a significant role in the recruitment to and stability of this community.

Although there are no specific numerical limitations regarding trawl animals, the California Ocean Plan (1997) states that:

- The rate of deposition of inert solids and the characteristics of inert solids in the ocean shall not be changed such that benthic communities are degraded.
- The concentration of substances set forth in Chapter IV, Table B, in marine sediments shall not be increased to levels which would degrade indigenous biota.
- The concentration of organic materials in marine sediments shall not be increased to levels which would degrade marine life.
- Nutrient materials shall not cause objectionable aquatic growths or degrade indigenous biota.
- Marine communities, including vertebrate, invertebrate, and plant species, shall not be degraded.



- Waste management systems that discharge to the ocean must be designed and operated in a manner that will maintain the indigenous marine life and a healthy and diverse marine community.
- Waste discharged to the ocean must be essentially free of: "2) Settleable material or substances that may form sediments which will degrade benthic communities or other aquatic life."

Based upon spatial and temporal comparisons and analogies with other studies, results of the trawl survey indicate that the discharge is in compliance with the general limitations and that it causes no adverse impact.



CHAPTER 8

Fish and Bivalve Tissue Bioaccumulation

8.1. Background

Outfall discharges can potentially increase contaminant concentrations in sediments and the water column to the extent that marine plant and animal communities are altered, reduced, or eliminated. Harvested fish or invertebrate flesh may become contaminated and unfit for human consumption. Bioaccumulation is a process whereby contaminants are assimilated by organisms, retained and bioconcentrated over time. The degree of bioconcentration is different among species and among toxicants. Biomagnification may also occur when predators eat organisms, resulting in the concentration of contaminants in higher levels of the food chain. In this way, higher-level predators, such as large fish, birds, and mammals can experience chronic toxicity, reproductive failure, or even mortality.

8.2. Materials and Methods

The measure of contaminants in animal tissues was performed with both fish (Pacific sanddabs, *Citharichthyssordidus*) and invertebrates (California bivalves, *Mytiluscalifornianus*) using two completely different collection procedures. Prior to 2014, Pacific sanddabs(*Citharichthyssordidus*) were collected in sufficient numbers for analysis. During this survey insufficient Pacific sanddabs were collected for chemical analysis and, as a result, a decision was made to use Pacific sanddabs which were more abundant. This species is included as a target species in Goleta's NPDES permit.

Pacific sanddabs were collected by otter trawl procedures, which are described in Section 7 above. Sanddabs collected in the population trawls were kept, and additional trawls were continued until sufficient total biomass for tissue analysis had been collected. Animals from each of two stations (TB3 between the diffuser and Goleta Point and TB6 at the down coast field control) were placed in plastic zip-lock bags and covered with ice in coolers. Immediately upon return to the laboratory, dorsal muscle and livers were removed from each animal, using standard clean room techniques, and placed in new pre-cleaned glass jars with Teflon-lined caps. All tissue samples were then stored in a freezer until ready to be shipped to the chemistry laboratory (PHYSIS Laboratories in Anaheim, California). Analytical methods were similar to sediments, except that special extraction and clean-up techniques were used to eliminate lipid interferences commonly found in marine animal tissues.

Bivalves were collected from Anacapa Island, California, an area anticipated to be very low in anthropogenic contamination. Prior to deployment these bivalves were cleaned of all debris and growth and held in a pre-cleaned seawater tank at 15° C until use. Bivalves were deployed using three arrays, each composed of a float, line, and anchor. Bivalve cages, made of plastic mesh netting, were attached to the middle of the arrays, so that the bivalves could be suspended at about mid-depth (16 m). The arrays were deployed in duplicate at Stations B3, B4, and B6; located 250, 25, and 3000 m (respectively) from the diffuser. The duplicate array at each station was suspended on a sub-surface buoy and attached to the first array with a 100 meter long line that was weighted to the bottom. Prior to deployment of the arrays in July, laboratory control bivalves were randomly selected and tissues were



resected and frozen. In October, each of the three bivalve arrays was successfully retrieved.

Once bivalves were removed from the array, they were placed on ice and returned to the laboratory. Exposed bivalves, as well as bivalves from the original population were cleaned, measured, and weighed. Their tissues were resected, stored, and analyzed, as above.

For the purposes of statistical analysis, all analytes from each of four groups (DDT and its derivatives (i.e. DDD and DDE), PCB's, PAH's, and non-DDT chlorinated pesticides) were combined. Results for individual analytes are presented in Appendix 10-16 and 10-17. All data were converted to mg/Kg or μ g/Kg, dry weight and statistically compared among stations using either t-test for two stations or analysis of variance (ANOVA) for three or more stations (see Section 3.4). When assumptions of parametric statistics could not be met (such as non-normality or excessive variability), the tests were replaced with nonparametric analogues (Aspin-Welch Unequal Variance Test, Mann-Whitney U, and Kruskal-Wallace Rank Test, respectively). Significance was noted when $p \le 0.05$ and marginal significance was noted when 0.05). A posteriori tests were utilized for significant ANOVA results to determine which stations were significantly different (see Zar 1996 or Sokal and Rohlf 1981 for a general description of statistical testing).

To compare tissue concentrations to the Office of Environmental Health Hazard Assessment (OEHHA) thresholds (OEHAA 2008) and NOAA Status and Trends mussel watch historical surveys (Kimbrough et al. 2008), Goleta tissue data were converted to wet weight units.

8.3. Results

Table 8-1 lists the physical and general descriptions of the animals utilized in the Goleta bioaccumulation study. Appendix Tables 10-13 and 10-14 lists lengths and weights of organisms, as well as tissue weights. Tables 8-2 to 8-4 and Figures 8-1 and 8-2 present average concentrations for each chemical constituent measured in the three types of animal tissues at each Station. Appendix Table 10-15 lists each constituent by replicate and averages by stations. Figures 8-3 through 8-5 compare historical contamination trends in the three tissue types. Tables 8-5to 8-6compare the Goleta tissue chemistry results with reference surveys and state OEHHA thresholds and NOAA status and trends tissue levels. Appendix 10-16 and 10-17 lists the concentrations of the derivatives of total DDT, non-DDT chlorinated hydrocarbons, total PCBs, and total PAHs. General descriptions of all chemical constituents have been presented earlier in Chapter 5, and so will not be repeated here.

8.3.1. Spatial contaminant patterns in tissues

Pacific sanddabs

A total of 96Pacific sanddabs (Citharichthyssordidus) were collected for tissue dissections from trawl transects TB3 (n = 59) and TB6 (n = 37), respectively. Average standard lengths (70and 78 mm, respectively) were similar between sites, while average weight was less at TB3 (6.5 g) compared to TB6 (10.0 g). Dissected tissue weights were greater for muscle tissue (1.0 and 1.7wet g, respectively) compared to liver (0.1g each).



Of the ten metals measured in sanddab muscle tissue all were above detection (Table 8-2 and Figure 8-1). Arsenic, nickel and zinc were each significantly greater by t-test (p< 0.05) at station TB6furthest from the outfall, compared to concentrations at TB3. Of the groups of complex organic compounds measured in sanddab muscle tissue, total chlordaneand total HCHs were below detection at both stations. Total PCB concentrations were significantly greater at TB6 (7.2 ug/L) compared to TB3 (4.5 ug/L) (p < 0.05). Total DDT concentrations were marginally significantly greater at TB6 (22.7 ug/L) compared to TB3 (20.0ug/L) (1.0 > p > 0.05).

Of the ten metals measured in sanddab liver, all were above detection (Table 8-3 and Figure 8-1). There were no significant differences in liver metal concentrations between the two sites by t-test (p < 0.05). Of the complex organic compounds, HCHswere below detection. There were no significant differences in liver organic concentrations between the two sites by t-test (p < 0.05).

Bivalves

Of the ten metals measured in bivalve (Mytiluscalifornianus) tissue, all were above detection(Table 8-4, Figure 8-1). Zinc wassignificantly greater at B6 compared to B4 by ANOVA (p < 0.05), and both of these stations were greater than B3. Of the complex organic compounds measured in bivalve tissue, chlordane,total HCHs and arochlorswere below detection at all stations. Total PCBs were only detected at B3. There were no significant differences among stations by ANOVA (p <0.05).

8.3.2 Tissue contaminant concentrations compared with past years

Pacific Sanddabs

The average concentration of contaminants in sanddab muscle and liver tissues remained within range of previous years (Figures 8-3 and 8-4). Increases in sanddab muscle concentrations of chromium, nickel and silver reported for the 2009 survey returned to lower concentrations in 2010 and remained low thru2014. Arsenic concentrations increased seven fold in muscle tissue from 2010(2 mg/dry Kg) to 2011 (15 mg/dry Kg), but dropped to 6 mg/dry Kg thru 2014. PCB concentrations in fish liver tissue have been highly variable since 2007 (range = 0 to 0.3 ug/L) and were elevated again in 2014.

Bivalves

The average concentration of each contaminant in bivalve tissues remained the same in 2014 (Figure 8-5). Arsenic concentrations were greater than in previous years.

8.3.3 Tissue contaminant concentrations compared with other surveys, State Thresholds & EPA Ranges

The concentrations of the contaminants measured in sanddab and bivalve tissues during the 2014 survey were compared to the concentrations measured at other sites throughout southern California (Table 8-5 and 8-6). Where comparisons were available, sanddab muscle and liver tissues, and mussel tissues were below or within the range of contaminant concentrations reported from other surveys(see references



in Table 8-5 and 8-6 footnotes). Sanddab and muscle tissue concentrations of metals and organic constituents did not exceed OEHHA consumption thresholds. Finally, mussel tissue concentrations were in the 'low' range reported by the NOAA Status and Trends Mussel Watch program.



8.4. Discussion

Results from this survey support past studies showing that the Goletaoutfall discharge appears not to effect the concentrations of contaminants in the tissues of fish and invertebrates residing in the survey area. Results from the chemical analysis of tissues were compared among stations, compared to past surveys in the area, compared to other studies performed in southern California, and compared to State thresholds and Federal ranges for concentrations of contaminants in animal tissue. Results for each variable were statistically compared among stations by either t-test or analysis of variance (ANOVA).

The sampling design for fish differed from the design for bivalve arrays. The bivalve sampling plan included a laboratory control (unexposed bivalves from Anacapa Island, CA) and bivalves exposed at three site locations: one station down coast (field control), one station nearest the outfall, and one station up coast and nearest Goleta Point. For fish, there was no laboratory control, and fish were collected from only two locations: one station down coast of the outfall corresponding to the field control, and one up coast of the outfall corresponding to the station nearest Goleta Point.

A total of 15 chemical compounds or groups of compounds were analyzed in Pacific sanddab muscle tissue from the two trawl locations. Sanddab muscle tissue metals were all above detection, whiletotal chlordane and total HCHs were each below method detection. Among the remaining compounds, arsenic, nickel, zinc and total PCBs were each significantly greater at station TB6,furthest from the outfall. In sanddab liver tissues each metal was above detection at each site, while HCHs were below detection. There were no significant differences between stations for any of the metals or organic compounds.

A total of 15 chemical compounds or groups of compounds were analyzed in the whole body tissues of bivalves. Zinc was significantly greater at B6 compared to B4 and both of these stations were greater than B3. Of the complex organic compounds measured in bivalve tissue, chlordane, total HCHs and arochlors were below detection at all stations. Total PCBs were only detected at B3. There were no significant differences among stations.

Comparison of the 2014 tissue concentrations from the Goleta survey area against results from the past nineteen years revealed that in all cases contaminant concentrations were similar to or less than in past years. Increases in sanddab muscle chromium, nickel and silver reported for the 2009 survey returned to lower concentrations in 2010 and remained low thru2014. Arsenic concentrations increased seven fold in sanddab muscle tissue from 2010 (2 mg/dry Kg) to 2011 (15 mg/dry Kg), and then dropped to 6 mg/dry Kg thru 2014.

The concentrations of the contaminants measured in sanddab and bivalve tissues during the 2014 survey were compared to the concentrations measured at other sites throughout southern California. Where comparisons were available, sanddab muscle and liver tissues, and mussel tissues were below or within the range of contaminant concentrations reported from other surveys. Sanddab and bivalve tissue concentrations of metals and organic constituents did not exceed OEHHA consumption thresholds. Since the Pacific sanddab is not caught for human consumption due to its small size, comparison of its tissue burdens against the OEHHA standard is included to provide context. Finally, bivalve tissue concentrations



were in the 'low' range reported by the NOAA Status and Trends Mussel Watch program (Kimbrough et al. 2008).

Although there are no specific numerical limitations regarding trawl animals, the California Ocean Plan (1997) states that:

The natural taste, odor, and color of fish, shellfish, or other marine resources used for human consumption shall not be altered.

The concentration of organic materials in fish, shellfish or other marine resources used for human consumption shall not bioaccumulate to levels that are harmful to human health.

Based upon spatial and temporal patterns and comparisons with other studies, results of the bioaccumulation survey indicate that the discharge is in compliance with the general limitations that it causes no adverse impact.



Table 8-1. Numbers of animals, length (mm), weight (g) and tissues weight (g) in fish and bivalve tissue collected in the Goleta survey area.

		Fish N	luscle	Fish	Liver		Biva	lves	
Constituent	Replicate	Т3	Т6	T3	T6	Control	B3	B4_	B6
Number of Animals		59	37	59	37	45	45	45	45
Average Standard Length (mm)	Mean =	69.8	78.4	69.8	78.4	66.6	70.4	69.6	74.4
	S.D. =	12.0	15.3	12.0	15.3	5.3	5.3	5.5	6.8
Average Weight/Animal (g)	Mean =	6.5	10.0	6.5	10.0	26.5	37.7	37.4	40.6
	S.D. =	3.2	5.7	3.2	5.7	6.5	10.5	8.1	7
Average Tissue Weight (g)	Mean =	1.0	1.7	0.1	0.1	5.8	11.6	12.0	13
	S.D. =	0.4	1.0	0.1	0.1	1.7	4.2	3.7	3.4



Table 8-2. Mean concentrations of Pacific sanddab (Citharichthyssordidus) muscle collected in the Goleta survey area. Comparisons of means determined by T-test (p < 0.05).

						Muscle			
Constituent	mean	TB3 ±	SD	mean	TB6 ±	SD	n	T-Test t	р
Metals (μg/dry g)									
Arsenic	5.776	±	0.289	7.081	±	0.085	3	-7.49	<0.01
Cadmium ²	0.025	±	0.000	0,037	±	0.001	3	-2.12	0.06
Chromium	0.055	±	0.010	0,071	±	0.022	3	-1.16	0.31
Copper	0.025	±	0.000	0.025	±	0.000	3	NA	NA
Lead	0.029	±	0.007	0.051	±	0.023	3	-1.58	0.19
Mercury	0.093	±	0.007	0.094	±	0.002	3	-0.44	0.68
Nickel	0.027	±	0.003	0.042	±	0.006	3	-3.72	0.02
Selenium	1.159	±	0.070	1,132	±	0.053	3	0.53	0.63
Silver	0.025	±	0.000	0.025	±	0.000	3	NA	NA
Zinc	10.392	±	0.470	15.012	±	2.269	3	-3.45	0.03
Complex Organics (ng/dry Kg)									
DDTs ¹	20.0	±	1.0	22.7	±	1.9	3	-2.15	0.10
Chlordane ¹	0.0	±	0.0	0.0	±	0.0	3	NA	NA
нснs¹	0.0	±	0.0	0.0	±	0.0	3	NA	NA
Aldrin	1.0	±	0.0	1.0	±	0.0	3	NA	NA
Dieldrin	1.0	±	0.0	1.0	±	0.0	3	NA	NA
Heptachlor	1.0	±	0.0	1.0	±	0.0	3	NA	NA
Hexachlorobenzene	1.0	±	0.0	1.0	±	0.0	3	NA	NA
Mirex	1.0	±	0.0	1.0	±	0.0	3	NA	NA
PCBs ¹	4.5	±	1.3	7.2	±	0.4	3	-3.30	0.03
Arochlors 1.2	0.0	±	0.0	3.4	±	5.8	3	-0.67	0.50
PAHs ¹	4.3	±	4.0	10.2	±	3.7	3	-1.88	0.13
1-Methylnaphthalene	1,0	±	0.0	1.1	±	0,1	3	-0.67	0.50
1-Methylphenanthrene	1.0	±	0.0	1.0	±	0.0	3	NA	NA
2-Methylnaphthalene	1.5	±	0.5	2.9	±	1.3	3	-1.69	0.17
2,3,5-Trimethylnaphthalene ²	1.5	±	0.5	1.0	±	0.0	3	1.29	0.20
2,6-Dimethylnaphthalene	3.7	±	1.5	2.4	±	0.6	3	1.35	0.25
Acenaphthene ²	1.0	±	0.0	2.1	±	0.1	3	-1.89	0.06
Biphenyl	1.0	±	0.0	1.0	±	0.0	3	NA	NA
Benz[a]anthracene	1.0	±	0.0	1.0	±	0.0	3	NA	NA
Benzo[b]fluoranthene	1.0	±	0.0	1.0	±	0.0	3	NA	NA
Benzo[e]pyrene	1.0	±	0.0	1.0	±	0.0	3	NA	NA
Benzo[g,h,i]perylene	1.0	±	0.0	1.0	±	0.0	3	NA	NA
Fluoranthene	1.0	±	0.0	1.0	±	0.0	3	0,43	0.69
Napthalene	3.3	±	2.0	2.5	±	1.3	3	0.59	0.59
Perylene	1.0	±	0.0	2.1	±	0.6	3	-1,86	0.06

^{1.} Complex Organic derivatives are listed in Table 10-16.

^{2.} Non-normal data. Statistics by Mann-Whitney U Test.



Table 8-3. Mean concentrations of Pacific sanddab (Citharichthyssordidus) liver collected in the Goleta survey area. Comparisons of means determined by T-test (p < 0.05).

		TD2				h Liver		T Y	
Constituent	mean	TB3 ±	SD	mean	TB6 ±	SD	n	T-Test t	P
Metals (μg/dry g)				-					<u>·</u> _
Arsenic ²	7.160	±	0.226	8.929	±	0.086	2	-1.16	0.25
Cadmium ²	5.991	±	0.103	13.608	±	0.544	2	-1.16	0.25
Chromium ²	0.583	±	0.206	0.263	±	0.200	2	1.16	0.25
Copper ²	8.720	±	0.308	15.643	±	1.182	2	-1.16	0.25
Lead ²	0,685	±	0.121	0.852	±	0.115	2	-0.82	0.41
Mercury ²	0.067	±	0.006	0.076	±	0.006	2	-1.16	0.25
Nickel ²	0.345	±	0.012	0.254	±	0.098	2	1.16	0.25
Selenium ²	5.201	±	0.030	4.326	±	0.178	2	1.16	0.25
Silver ²	0.251	±	0.032	0.187	±	0.013	2	1.16	0.25
Zinc²	56.243	±	3.417	64.532	±	1.359	2	-1.16	0.25
Complex Organics (ng/dry Kg)									
DDTs ^{1,2}	785.4	±	32.9	1016.0	±	66.9	2	-1.16	0.25
Chlordane ^{1,2}	38.2	±	2.6	36.4	±	5.5	2	0.39	1.00
HCHs ¹	0.0	±	0.0	0.0	±	0.0	2	NA	NA
Aldrin	1.0	±	0.0	1.0	±	0.0	2	NA	NA
Dieldrin	1.0	±	0.0	1.0	±	0.0	2	NA	NA
Heptachlor	1.0	±	0.0	1.0	±	0.0	2	NA	NA
Hexachlorobenzene ²	5.1	±	0.5	1.0	±	0.0	2	1.22	0.22
Mirex	1.0	±	0.0	1.0	±	0.0	2	NA	NA
PCBs ^{1,2}	230.2	±	2.1	328.8	±	7.5	2	-1.16	0.25
Arochlors ²	289.7	±	2.6	414.1	±	9.6	2	-1.16	0.25
PAHs ¹²	86.4	±	8.0	144.2	±	15.2	2	-1.16	0.25
1-Methylnaphthalene ²	10.2	±	0.3	8.6	±	1,4	2	1.16	0.25
1-Methylphenanthrene ²	1.0	±	0.0	8.2	±	1,1	2	-1.22	0.22
2-Methylnaphthalene ²	31.6	±	6.0	44.5	±	14.4	2	-0.39	0.70
2,3,5-Trimethylnaphthalene ²	16.2	±	1.0	24.1	±	2.5	2	-1.16	0.25
2,6-Dimethylnaphthalene ²	10.3	±	1.3	14.6	±	4.3	2	-1.16	0.25
Acenaphthene	1.0	±	0.0	1.0	±	0.0	2	NA	NA
Biphenyl ²	1.4	±	0.5	3.7	±	2.3	2	-1.16	0.25
Benz[a]anthracene²	1.0	±	0.0	1.0	±	0.0	2	NA	NA
Benzo[b]fluoranthene	1.0	±	0.0	1.0	±	0.0	2	NA	NA
Benzo(e]pyrene	1.0	±	0.0	1.0	±	0.0	2	NA	NA
Benzo(g,h,i)perylene	1.0	±	0.0	1,0	±	0.0	2	NA	NA
Fluoranthene ²	1.0	±	0.0	40.7	±	15.8	2	-1.22	0.22
Napthalene ²	36.0	±	5.1	49.8	±	14.4	2	-0.82	0.41
Perylene ²	35.8	±	4.9	57.4	±	2.3	2	-1.16	0.25

^{1.} Complex Organic derivatives are listed in Table 10-16. 2. Non-normal data. Statistics by Mann-Whitney U Test.



Table 8-4. Heavy metals and complex organics in California bivalve (Mytiluscalifornianus) tissues. Comparisons of means by ANOVA (p < 0.05).

		В3			B4	Bival	ve Tissue	B6			ANOV/A	
Constituent	mean	±	SD	mean	±	SD	mean	±	SD	n	ANOVA F	р
Metals (μg/dry g)												
Arsenic	17.779	±	0.995	15.727	±	0.992	15.801	±	1.023	3	4.04	0.0
Cadmium	2.892	±	0.357	2.692	±	0.039	2.795	±	0.244	3	0.48	0.6
Chromium	1.066	±	0.085	1.087	±	0.107	1.025	±	0,177	3	0.18	0.8
Copper	5.597	±	0.398	5.247	±	0.663	5.679	±	0.800	3	0.38	0.7
Lead	1.166	±	0.064	1.125	±	0.065	1.198	±	0.095	3	0.70	0.5
Mercury	0.028	±	0.002	0.027	±	0.002	0.027	±	0.003	3	0.21	0.8
Nickel	0.777	±	0.118	0.787	±	0.096	0.747	±	0.119	3	0.10	0.9
Selenium ²	2.764		0.050	2.574								
\ \ \		±			±	0.013	2,934	±	0.400	3	2.40	0.3
Silver	0.174	±	0.052	0.131	±	0.015	0.167	±	0.046	3	0.95	0.4
Zinc	113.771	±	2.581	107.919	±	2.848	124,395	±	8.612	3	7.06	0.0
Complex Organics (ng/dry Kg)												
DDTs ^{1.}	5,0	±	8,0	3.9	±	0.3	4.1	±	0.4	3	3.73	0.0
Chlordane ^{1.}	0.0	±	0.0	0.0	±	0.0	0.0	±	0.0	3	NA	N
HCHs ^{1.}	0.0	±	0.0	0.0	±	0.0	0,0	±	0.0	3	NA	N
Aldrin	1.0	±	0.0	1.0	±	0.0	1.0	±	0.0	3	NA	Ν
Dieldrin	1.0	±	0.0	1.0	±	0.0	1.0	±	0.0	3	NA	N
Heptachlor	1.0	±	0.0	1.0	±	0.0	1.0	±	0,0	3	NA	N
Hexachlorobenzene	1.0	±	0.0	1.0	±	0.0	1.0	±	0.0	3	NA	N
Mirex	1.0	±	0.0	1.0	±	0.0	1.0	±	0.0	3	NA	Ν
PCBs ^{1, 2}	0.4	±	0.8	0.0	±	0.0	0.0	±	0.0	3	0.60	0.
Arochlors 1.	0.0	±	0.0	0.0	±	0.0	0.0	±	0.0	3	NA	Ν
PAHs ^{1.}	0.8	±	1.3	1.3	±	1.4	0.4	±	0.6	3	0.51	0.
1-Methylnaphthalene	2.4	±	0.4	3,2	±	0.6	3.5	±	0.5	3	4.48	0.
1-Methylphenanthrene	6.4	±	0.4	6.7	±	0.7	5.8	±	0.1	3	3.12	0.
2-Methylnaphthalene	5.8	±	2.1	7.9	±	0.3	7.6	±	2.2	3	1.30	0.:
2,3,5-Trimethylnaphthalene	1.3	±	0.4	1.2	±	0.2	1.2	±	0.3	3	0.08	0.
2,6-Dimethylnaphthalene	1.0	±	0.0	1.0	±	0.0	1.0	±	0,0	3	NA	N
Acenaphthene	1.0	±	0.0	1.0	±	0.0	1.0	±	0.0	3	NA	N
Biphenyl	1.0	±	0.0	1.0	±	0.0	1.0	±	0.0	3	NA	٨
Benz(a)anthracene	1.0	±	0.0	1.0	±	0.0	1.0	±	0.0	3	NA	٨
Benzo[b]fluoranthene	1.0	±	0.0	1.0	±	0.0	1.0	±	0.0	3	NA	٨
Benzo[e]pyrene	1.0	±	0.0	1.0	±	0.0	1.0	±	0.0	3	NA	N
Benzo[g,h,i]perylene	1.0	±	0.0	1.0	±	0.0	1.0	±	0.0	3	NA	N
Fluoranthene	1.0	±	0.0	1.0	±	0.0	1.0	±	0.0	3	NA	
Napthalene	1.8	±	0.5	2.0	±	0.6	1.4	±	0.1	3	1.27	0.
Perylene ²	1.0	±	0.0	1.1	±	0.2	1.0	±	0.0	3	0.60	0

Complex Organic derivatives are listed in Table 10-17.
 Non-normal data. Statistics by Kruskal-Wallis Test.



Figure 8-1. Metal concentrations (mg/dry Kg) measured in fish muscle and liver tissues (Stations TB3 and TB6), and bivalves(Stations B3, B4, B6 and lab control).

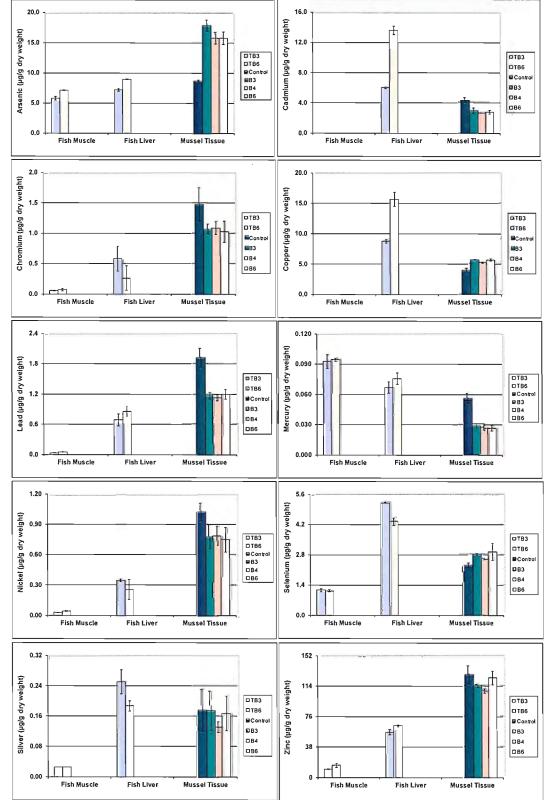




Figure 8-2. Organic concentrations ($\mu g/dry$ Kg) measured in fish muscle and liver tissues (Stations TB3 and TB6), and mussels (B3, B4, B6 and lab control).

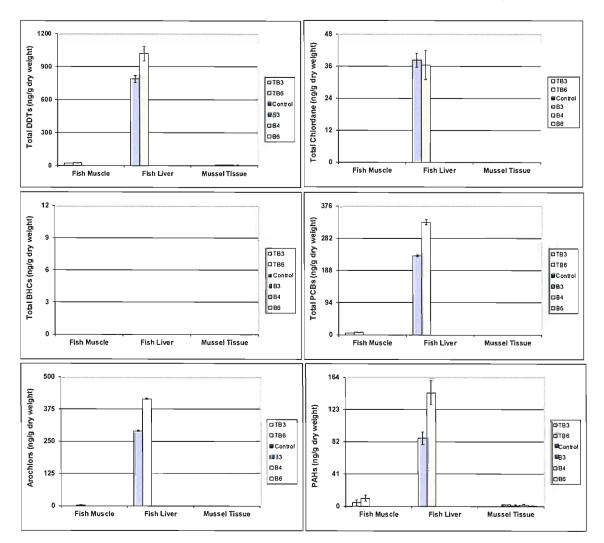




Figure 8-3. Contaminants (mg/dry Kg) measured in Pacific sanddab muscle (Citharichthyssordidus) from Goleta since 1991 (mean \pm SD, n=6).

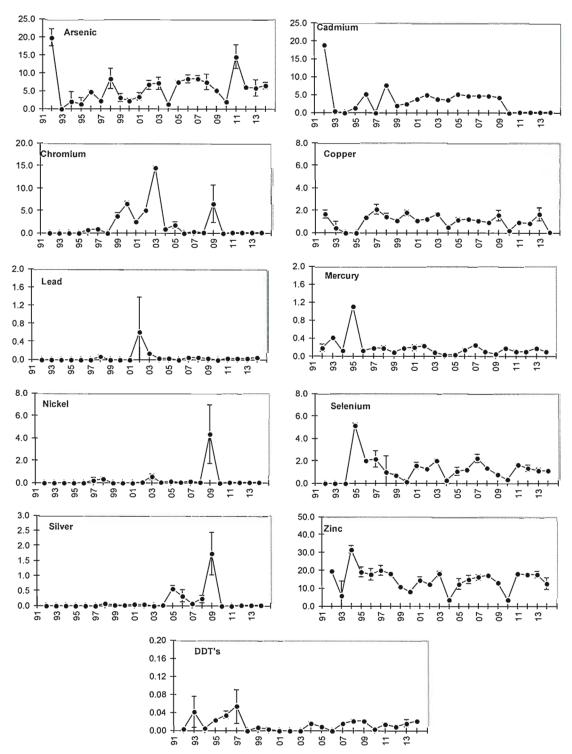




Figure 8-4. Contaminants (mg/dry Kg) measured in Pacific sanddab liver (Citharichthyssordidus) from Goleta since 1991 (mean \pm SD).

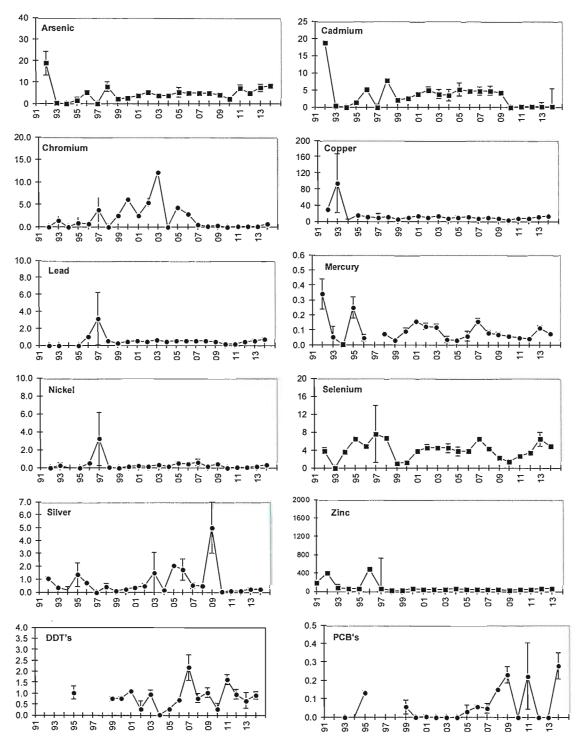




Figure 8-5. Contaminants (mg/dry Kg) measured in whole bivalves (Mytiluscalifornianus) from Goleta since 1991 (mean \pm SD, n = 3).

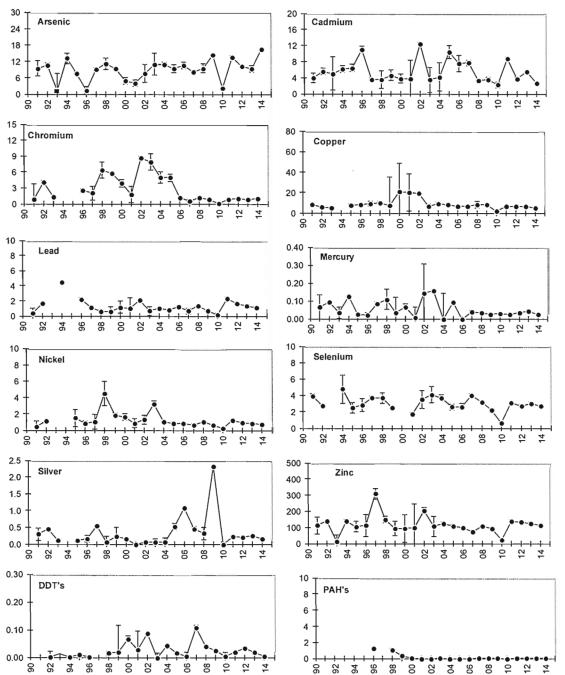




Table 8-5. Comparison of Goleta tissue chemistry with results from other studies (ug/wet g) and state and federal limits.

	<u>G</u>	OLETA S.D.		Reference	OEH	HA ²
	μg/	g Wet Weigh	t	μg/g Wet Weight		t Weight
Constituent	Means	Rang	es	Stations ^{1.}	FCG 3.	ATL ⁴.
Fish Muscle						
Arsenic	1.276	1.080 -	1.424	42.2 - 57.8		
Cadmium	0.006	0.005 -	800.0	<0.01 - 0.045		
Chromium	0.012	0.009 -	0.018	0.08 - 2.8		
Copper	0.005	0.005 -	0.005	0.45 - 2.4		
Lead	0.008	0.005 -	0.014	1.2		
Mercury	0.019	0.017 -	0.020	0.36 - 0.78	0.22	≤0.07 ^{5.}
Nickel	0.007	0.005 -	0.010	0.4 - 5.1		
Selenium	0.227	0.213 -	0.239	2.8 - 3.95	7.4	≤2.5
Silver	0.005	0.005 -	0.005	<0.005 - 1.4		
Zinc	2.521	1.974 -	3.497	12.4 - 30.5		
DDTs	0.004	0.004 -	0.005	0.005 - 2.15	0.021	≤0.52
Chlordane	0.000	0.000 -	0.000		0.0056	≤0.052
PCBs	0.001	0.001 -	0.001	0.005 - 2.7	0.0036	≤0.021
PAHs	0.001	0.000 -	0.003			
Fish Liver						
Arsenic	2.49	2.16 -	2.78			
Cadmium	3.03	1.83 -	4.32			
Chromium	0.13	0.04 -	0.22	0.5		
Copper	3.76	2.63 -	5.09			
Lead	0.24	0.19 -	0.29			
Mercury	0.02	0.02 -	0.02			
Nickel	0.09	0.06 -	0.11			
Selenium	1.47	1.30 -	1.61			
Silver	0.07	0.05 -	80.0			
Zinc	18.66	16.63 -	20.24			
DDTs	0.278	0.235 -	0.329	28		
Chlordane	0.012	0.010 -	0.012			
PCBs	0.086	0.071 -	0.103	4		
PAHs	0.036	0.025 -	0.048			

^{1.} Sources: SWRCB 1978, 1988 (EDL 85); SCCWRP 1975, 1976, 1977, 1982, 1998c; Short & Harris 1996; Brown & Caldwell 1997; NOAA 1991, OEHHA 1991

^{4.} Advisory Tissue Levels (ATLs), most conservative tissue consumption threshold based on cancer or non-cancer risk. 5. Mercury ATL for women aged 18-45 years & children aged 1-17 years (OEHHA 2008).





^{2.} OEHHA, 2008

^{3.} Fish Contamination Goal (FCG)

Table 8-6. Comparison of Goleta tissue chemistry with results from other studies (ug/wet g).

	9	GOLETA S.D.	Reference	핑	OEHHA 2.	NOAA Sta	NOAA Status & Trends, 1986 to 2005	16 to 2005
Constituent	µg/g Means	ug/g Wet Weight Ranges	ug/g Wet Weight Stations ¹	µg/g Wel FCG ^{3.}	µg/g Wet Weight FCG 3. ATL 4.	low	µg/g Wet Weight medium	high
Mussel Tissue								
Arsenic	2.77	1.62 - 3.52	16.0 - 23.8	ļ	1	5 - 11	12 - 22	23 - 41
Cadmium	0.61	0.48 - 0.90	1.9 - 54	l	I	0-3	4 - 9	10 - 20
Chromium	0.22	0.16 - 0.32	1.23 - 3.9	I	ı	ļ	i	i
Copper	0.97	0.67 - 1.26	4.0 - 21.8	ł	1	5 - 16	17 - 39	40 - 857
Lead	0.26	0.20 - 0.40	1.09 - 11	i	ŀ	0-3	4 - 6	7 - 13
Mercury	0.01	0.00 - 0.01	0.01 - 0.4	0.22	≤0.07 ^{5.}	0.00 - 0.17	0.18 - 0.35	0.36 - 1.28
Nickel	0.16	0.12 - 0.21	3.2 - 5.3	i	ł	0-5	6 - 14	15 - 44
Selenium	0.50	0.42 - 0.62	2.70 - 4.57	7.4	≤2.5	1	1	i
Silver	0.03	0.02 - 0.05	0.36 - 0.7	ŀ	ļ	1	1	1
Zinc	22.60	20.04 - 26.45	133 - 336	ŀ	ł	48 - 139	140 - 320	321 - 11500
DDTs	6000.0	0.0007 - 0.0012	0.017 - 0.35	0.21	≤0.52	0 - 0.112	0.113 - 0.286	0.287 - 0.520
Chlordane	0.0000	0.0000 - 0.0000	!	0.0056	≤0.19	0 - 0.008	0.009 - 0.020	0.021 - 0.049
PCBs	0.0001	0.0000 - 0.0002	0.017 - 0.35	0.0036	≤0.021	0.003 - 0.153	0.154 - 0.478	0.479 - 1.413
PAHs	0.0002	0.0000 - 0.0005	0.81	}	ţ	0.063 - 1.187	1.118 - 4.434	4.435 - 7.561

Sources: SWRCB 1978, 1988 (EDL 85); SCCWRP 1975, 1976, 1977, 1982, 1998c; Short & Harris 1996; Brown & Caldwell 1997; NOAA 1991, OEHHA 1991
 OEHHA, 2008
 Fish contaminant goals; based on cancer and non-cancer risk using an 8 oz/week consumption rate.
 Advisory tissue levels; based on cancer and non-cancer risk using an 8 oz/week consumption rate (OEHHA 2008).



9.0 Introduction

Aquatic Bioassay biologists conducted underwater dive surveys and underwater videos of the outfall pipe and diffuser from the Goleta Sanitary District Wastewater Treatment Plant on October 29th, 2014. The purposes of the survey were to inspect the physical integrity of the outfall pipe and associated armor rock and note any impediments to flow from the 36 diffuser ports. Aquatic Bioassay biologists also assessed the presence of attached and mobile marine organisms that were associated with the outfall and the diffuser.

9.1 Materials and Methods

Five divers, using Sony 2100 Camcorders enclosed in Gates underwater housings with attached NiteRider underwater lights, conducted the survey. Once the outfall had been located by global positioning (GPS) and bottom finder, a buoy, attached to a line and a weight, was deployed over the side. Divers entered the water, descended down the line, swam to the diffuser terminus, and began filming. At the end of each dive, a lift float was deployed as a marker for the subsequent dive. On deck between dives, the camera was removed from the housing, the footage was inspected, batterieswere replaced, and the housing was reassembled. A total of seven dives were completed for the video: diffuser, west and east ports (100 ft. to 70 ft.); deep outfall (70 ft. to 40 ft.); middle outfall (40 ft. to 20 ft.), and shallow outfall (20 ft. to surf zone).

The footage was downloaded to computer files, edited using *Adobe Premiere* software, and then transferred to DVD. DVDs were then reviewed by the survey team to assess conditions of the outfall. The video is arranged from the deepest part of the dives (outfall terminus) to the shallowest part of the dives (outfall beginning).

9.2 Results

Outfall dive surveys were conducted between approximately 0800 and 1645hours on October 29^{th} , 2014aboard the research vessel *Hey Jude*. Weather conditions were clear with a 10knot wind from the west, southwest (225°) and 2 to 4ft swell from the southwest (245°). Water color was greenwith high turbidity. Visibility at the terminus of the diffuser (100 feet) was \leq 1meter. There was a thermocline at approximately 7meters.

9.2.1 Diffuser Section (Depth: 100 TO 70ft)

9.2.1.1 Physical Description

The pipe survey was conducted in the October in hopes that water quality would be optimal for taking video footage of the pipe. This year's visibility was poor,ranging from 0 to 1 meter. The diffuser section contains 34 lateral and two terminal discharge ports. The lateral ports are alternately arranged 17 on each side of the diffuser. The end of the pipe is closed except for the two terminal ports, which are situated one above the other. There were no obstructions on the upperport of the terminus cap, and the flow from both the upper and lower terminal ports was strong.



Lateral ports were observed and videotaped, starting at the terminus and moving shoreward on the east side of the pipe, thenfrom the terminus down the west side until the most shoreward east port was occupied at the beginning of the diffuser. All of thelateral ports were flowing freely. Along the length of the diffuser pipe, no evidence of leaks, damage, erosion, holes, or cracks were observed.

An approximately one meter high bed of armor rock supports the diffuser section. Intermittent observations of the supporting armor rock revealed a stable bed of rock with little displacement throughout the diffuser section. Probably during initial construction, the diffuser section appears to have been rotated counter-clockwise (as if one were facing the terminus). Thus, the line across east and west diffuser ports is not parallel to the sea floor, and west ports are about 30 cm lower than east ports. Armor rock covers the outfall from the shoreward beginning of the diffuser to the shoreward beginning of the outfall in very shallow water. The thickness of the armor rock is about one meter.

9.2.1.2 Biological Description

Because of the depth and relative low light at the diffuser (100 ft), algal species are typically scarce. Algae that were present included the kelp *Desmarestialigulataa* tubular and leafy red alga (Rhodophyta), and the Turkish Towel (*Gigartina sp.*). Among invertebrates; brown cup coral (*Paracyathussternsi*), colonial strawberry anemones (*Corynactiscalifornica*), red gorgonian (*Lophogorgiachilensis*) and various species of colonial hydroids and bryozoans dominated. Tube worms and especially the strawberry anemones were commonly observed surrounding the diffuser ports. Batstars (*Patiriaminiata*), Octopus (*Octopus*sp), sheephead (*Semicossyphuspulcher*) and kelp bass (*Paralabraxclathratus*) were observed either on the pipe, or in its immediate vicinity.

9.2.2 Deep Outfall Section (Depth: 70 TO 40ft)

9.2.2.1 Physical Description

Throughout the dive survey, the outfall was completely covered by approximately one-meter layer of armor rock. Visibility was very poor in this section. The rock covered pipe extended vertically from the sea floor for about 2 to 3 meters and laterally for about 6 to 7 meters. The armor rock bed appeared stable with little displacement throughout this section. No obvious leaks or discoloration were observed from the armor rock covering the top or sides of the outfall pipe.

9.2.2.2 <u>Biological Description</u>

On this section, crustose coralline alga (Rhodophyta), foliose red algae (*Gigartina sp.*) and giant kelp (*Macrocystuspacifica*)dominated the algal community. Among invertebrates, the most abundant were the red gorgonian (*Lophogorgiachilensis*), colonial strawberry anemones (*Corynactiscalifornica*), several species of bryozoans, bat stars (*Patria miniata*), giant sea stars (*Pisastergiganteus*),the giant keyhole limpet (*Megathuracrenulata*),and red urchin (Strongylocentrotusfranciscanus) and the wavy top turban (*Megastraeaundosa*). Several fish species were observed including rockfish (*Sebastess*p),kelp bass(*Paralabraxclathratus*), opaleye (*Girella*nigricans),sheephead



(Semicossyphuspulcher) and blacksmith(Chromispunctipinnis). Lobster traps were also observed on the armored rock.

9.2.3 Middle and Shallow Outfall Section (Depth: 40 TO Surf Zone)

9.2.3.1 Physical Description

As with the previous section, this outfall section was covered by about one meter of armor rock. The armor rock covered pipe extended horizontally and laterally as above. The armor rock bed appeared stable with little displacement throughout this section. No obvious leaks or discoloration were observed from the armor rock covering the top or sides of the outfall pipe.

9.2.3.2 Biological Description

This outfall section supported a giant kelp forest(*Macrocystispyrifera*) from 1997 to 2008 (MBC 1997, 1998; Aquatic Bioassay 1999 to 2008), however from 2009 to 2013the density of the kelp decreased and the armoredrock on this section was populated by large densities of purple sea urchins (*Strongylocentrotuspurpuratus*) whose favorite food source is giant kelp. The giant kelp density increased this year when compared to 2013 and purple sea urchin populations decreased. This growth and predation cycle is typical on California rocky reefs.

Dominant algae in this pipe section included foliose red algae (Gigartina sp.), crustose coralline algaeand giant kelp (Macrocysitispyrifera). Among the macroinvertebrates, the giant keyhole limpets (Megathuracrenulata), purple sea urchins (Strongylocentrotuspurpuratus), red urchin (Strongylocentrotusfranciscanus) and red gorgonian (Lophogorgiachilensis) were most dominant. Fish species observed at this depth included blacksmith(Chromispunctipinnis), kelp bass (Paralabraxclathratus), sheephead (Semicossyphuspulcher), senorita (Oxyjuliscalifornica), and painted greenling (Oxylebiuspictus). Lobster traps were also observed on the armoredrock.



Discussion

During the diffuser dive survey, 36 diffuser ports were carefully inspected for flow and general efficiency. This year, none of the diffuser ports were obstructed with debris and all of the ports were flowing freely. The remainder of the outfall pipe was inspected for damage, leaks or evidence of leaks and general stability of the pipe and armor rock. Inspection of theoutfall yielded no evidence of damage, holes, cracks, or erosion. The pipe and associated armor rock appeared stable with little or no displacement.

The outfall continues to support a rocky reef community typical of other areas on the central California coast. A visual survey yielded numerous different species of plants, macroinvertebrates, and fishes. A number of species of fish were represented by juvenile or larval forms, which indicates that recruitment has been occurring. Fish appeared healthy, with no evidence of deformities, tumors, fin rot, or lesions.

During past surveys the 40 to 20 foot outfall section supported a giant kelp forest(Macrocystispyrifera) that was extremely dense (MBC 1997, 1998; Aquatic Bioassay 1999 to 2008), however from 2009to 2013 the density of the kelp was less than in past surveys. The armor rock on this section was populated by large densities of purple sea urchins (Strongylocentrotuspurpuratus) whose favorite food source is giant kelp. The figure to the right shows purple urchins eating a giant kelp holdfast. During the preceding years the purple urchin population had thinned the kelp forest residing on the outfall pipe through predation (Tegner et al. 1995). Once the kelp plant holdfast is weakened, storms act to break the plant free. This year, the population of purple urchins has decreased and the kelp densities have increased when compared to 2013. This growth and predation cycle is typical on California rocky reefs.





CHAPTER 10

COLLECTION SYSTEM ANNUAL SUMMARY

Background

Sanitary sewer overflows associated with the Goleta Sanitary District's collection system are subject to the online reporting and notification requirements set forth in the Statewide General Waste Discharge Requirements for Sanitary Sewer Systems Order NO. 2006-0003-DWQ. The Goleta Sanitary District has enrolled under the statewide waste discharge requirement for sanitary sewer systems.

GSD completed the Sanitary Sewer Management Plan (SSMP) in December 2006 and reviews and revises the SSMP annually, as needed. The District's SSMP was updated in September of 2013 in accordance with SWRCB Order No. WQ 2013-0058 – EXEC MRP.

This annual report summarizes all lift station and collection system overflows that occurred during 2014 and includes, if any, the cause, corrective actions taken and corrective actions planned. In conjunction with the annual report the District will conduct the annual SSMP update. The update is a part of the wastewater collection system management plan and requires the District to conduct an internal audit to evaluate the wastewater collection system management plan and delineate steps the District will take to correct any deficiencies that are found.

Annual Reporting Requirement

This chapter is included as part of the wastewater treatment plant annual report.

Summary of 2014 Spills

Lift Station Overflows

There were no lift station overflows that occurred within the Goleta Sanitary District service area during 2014.

Collection System Overflows

There were two collection system overflows that occurred within the Goleta Sanitary District service area during 2014.

SSO #1 occurred on March 3rd as a result of roots in the District's sewer main line. Approximately 80 gallons were calculated to have backed up into two bathrooms of a condominium complex.

Corrective actions taken include the cleaning and CCTV inspection of the District sewer line and verification that there were no other readily apparent issues that would contribute to a repeat spill at this location. The affected bathrooms were cleaned and sanitized by a professional cleaning service.

Corrective planned action measures include the increased monitoring of this sewer line and the application of chemical root foaming to the sewer line. The condominium owners and home owners association were notified to install and maintain proper back water valves to prevent future sewer backups.

SSO #2 occurred on April 16th as a result of roots in the District's sewer main line. Approximately 240 gallons were calculated to have spilled around the manhole into a driveway and dirt area.

Corrective actions taken include the cleaning and CCTV inspection of the District sewer line to verify that the blockage was removed and that there were no apparent issues that would contribute to a repeat spill at this location. The affected driveway and dirt areas were cleaned and sanitized with a bio-degradable disinfectant.

Corrective planned action measures include the increased monitoring of this sewer line.

Discussion

The Goleta Sanitary District's wastewater collection system management plan has been completed and complies with all of the requirements of MRP No. R3-2010-0012. All detailed tasks have been addressed in a timely manner and the collection system has complied with all requirements of the monitoring and reporting program.

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10.2. Water Quality Correlation Data



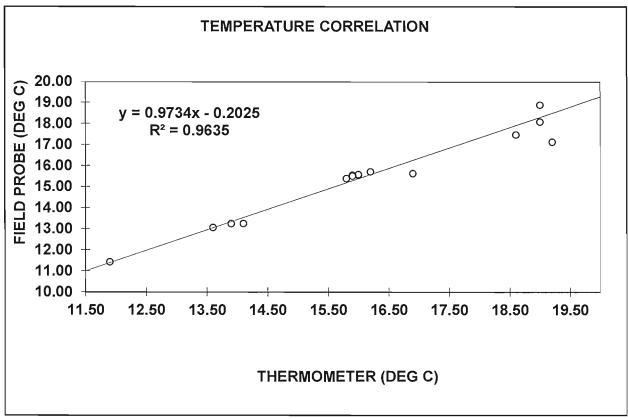


Figure 10-1. Correlations between CTD probes and analysis of discrete water samples measured using field probes.



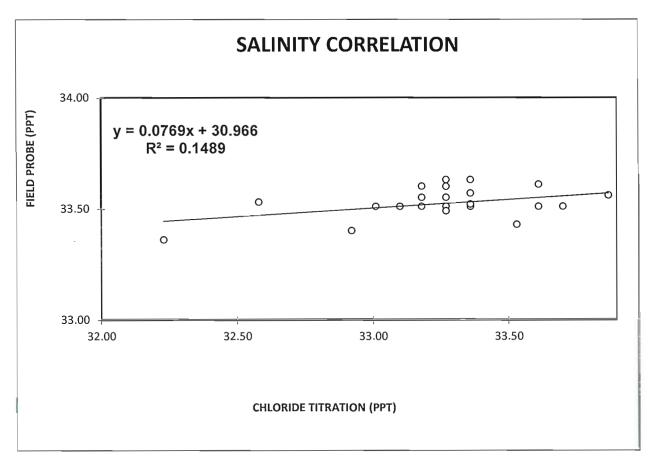


Figure 10-1. (continued)



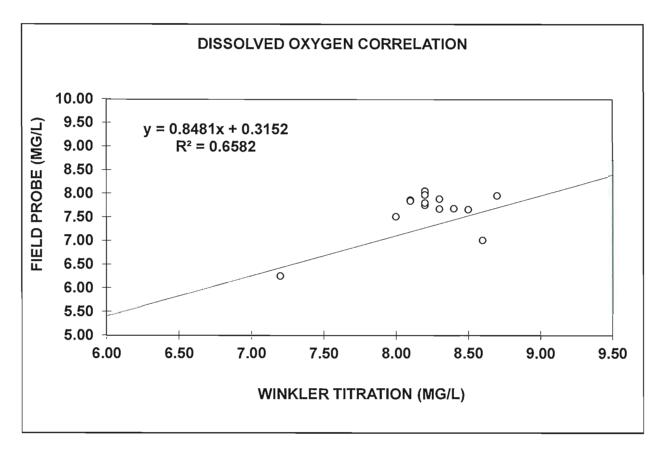


Figure 10-1. (continued)



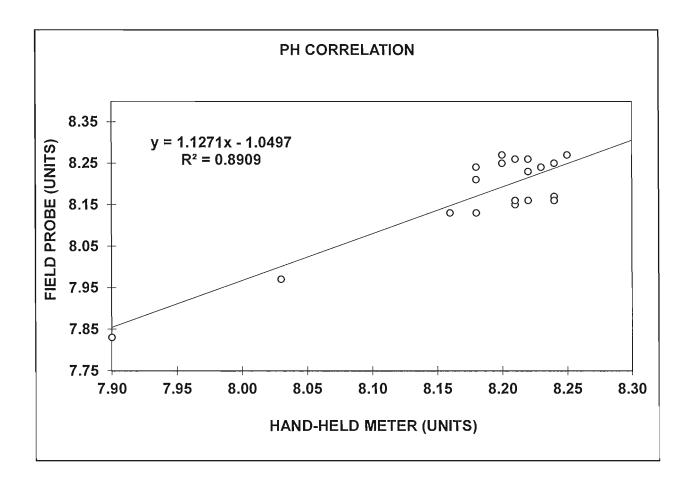


Figure 10-1. (continued)



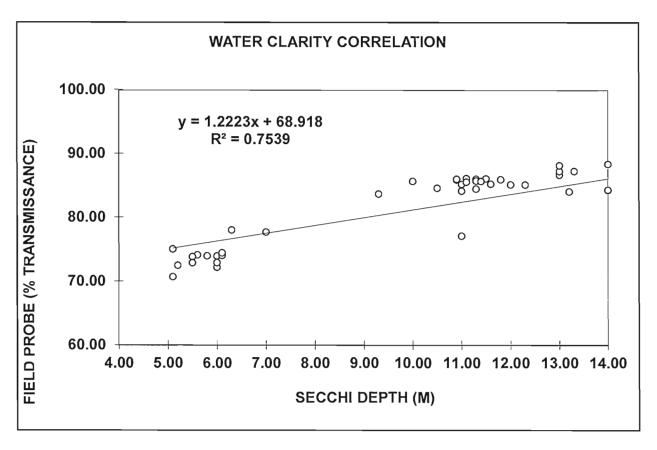


Figure 10-1. (continued)



10.3. Particle Size



16

Table 10-2. Particle sizes by channel sizes in phi and microns for each Goleta sediment station.

Sample coarse coarse med med B1 0.00 0.00 0.00 0.00 0		3 3 2 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	354 354 med	4.20mm 3.30mm 4.20mm 3.30mm 4.20mm	2.5	•	3.5	- statistics	200															
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coarse coarse med sand sand sand 0.00 0.00 0.00 0.00 0.00 0.00	121212	135 . E.13	med		very	very	very	very	very				でなる機		very	yery								
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00.0 0.00 0.00	0.04	0.73	3.97	12.40	18.71	17.76	12.05	6.70	3.89	3.13	3.23	3.52	3.68	3.06	2.19	1.37	1,08	0.71	0.50	0.49	0.43	0.32	0.02	0.00
_	0.00	0.27	1.38	6.27	19.29	24.80	15.96	7.16	3.47	2.64	2.75	3.09	3.32	2.84	2.08	1.32	1.05	0.71	0.50	0.47	0.42	0.23	0.00	0.00
B3 0.00 0.00 0.00 0.00 0	0.03	0.54	2.35	10.53	27.07	24.13	10.58	4.18	2.30	5.09	2.46	2.87	3.03	2.45	1.69	1.05	0.83	0.56	0.41	0.39	96.0	60.0	0.00	0.00
B4 0.00 0.00 0.00 0	0.04	0.67	3.11	12.62	26.43	20.07	8.80	4.11	2.73	2.66	3.03	3.32	3.38	2.75	1.95	1.23	96.0	0.64	0.45	0.44	0.41	0.20	0.00	0.00
B5 0,00 0.00 0.00 0	0.03	0.54	2.91	13.73	27.93	20.65	9.44	4.43	2.70	2.36	2.49	2.66	5.69	2.20	1.58	1.02	0.82	0.56	0.41	0.39	0.37	60.0	0.00	0.00
B6 0.00 0.00 0.00 0	0.03	0.35	1.08	4.14	15.52	28.54	21.60	8.56	3.44	2.47	2.56	2.74	2.65	1.99	1.32	0.81	0.64	0.46	0.36	0.34	0.31	90.0	0.00	0.00

Table 10-3. Summary of particle sizes by fraction, percentiles, dispersion, sorting index and distribution.

ution.	Kurtosis	-2.60	-2.72	-2.84	-2.66	-2.91	-3.12
Distribution (phi)	Skewness	-0.07	-0.01	-0.01	-0.01	0.00	-0.04
Dispersion or Sorting Index		1.88	1.65	1.50	1.74	1.45	1.16
	Mode	2.82	3.21	2.87	2.81	2.81	3.30
bh	Median	3.39	3.46	3.19	3.17	3.11	3.50
	Mean	3.26	3.43	3.18	3.16	3.11	3.46
	Mode	141.51	107.59	136.30	141.63	142.47	101.31
Microns	Median	94.86	90.90	109.10	110.53	115.21	87.92
	Mean	104.27	92.40	110.14	111.35	115.57	90.47
	95%	2.00	2.26	2.09	2.04	2.05	2.42
je J	84%	2,45	2.70	2.54	2.47	2.45	2.83
Percentile (phi)	%09	3.39	3.46	3.19	3.17	3.11	3.50
a.	16%	6.20	6.01	5.54	5.96	5.35	5.15
	% 9	7.99	7,94	7.63	7.84	7.59	7.34
	%96 _{**}	0 248.21	9 207.87	1 233.55	5 242.07	7 240.63	0 186.11
lle Is)	84%	182.70	152.99	171.31	179.15	182.47	140.20
Percentile (microns)	20%	94.86	90.90	109.10	110.53	115.21	87.92
	16%	13.65	15.57	21.51	16.09	24.57	28.21
	2%	3.95	4.10	5.08	4.39	5.22	6.19
	Silt-Clay	27.63	24.88	20.58	24.15	20.34	20.18
	Clay	3.55	3.37	2.65	3.10	2.64	2.20
(cent)	alls.	24.07	21.51	17.93	21.04	17.70	17.99
Sum. (Per	the state of the	72.37	75.12	79.42	75.85	99.62	79.82
	Gravel* Sand	0.00	0.00	0.00	0.00	0.00	00.0
	Sample ID	181	B2	88	78	B5	98 8

^{*}Percentage of the sample retained on a 2 mm sieve.



10.4Sediment Chemistry



10-4. Sediment contaminant concentrations normalized to percent fine sediments in the Goleta survey area. Correlations by nonparametric Spearman's rho.

			Sediment	Stations					Correl	ations
Constituent	B1	B2	B3	B4	B5	В6	Mean	S,D.	Outfall	Point
I budiffe so national Osmonio										
Undifferentiated Organics Oil and Grease	32	24	18	9	7	19	18.3	9.3	0.65	-0.81
TKN	18	22	21	22	19	33	22.3	5.3	0.01	0.90
TOC	398	414	243	203	165	307	288	103	0.75	-0.54
AVS	0.19	0.17	0.28	0.17	0.12	0.18	0.18	0.05	0.13	-0.72
Heavy Metals Aluminum	367	399	339	361	415	592	412	92	0.14	0.49
Antimony	0.01	0.00	0.00	0.00	0.00	0.01	0.00	0.00	0.84	0.00
Arsenic	0.19	0.23	0.22	0.21	0.19	0.29	0.22	0.04	0.17	0.77
Cadmium	0.01	0.02	0.02	0.02	0.01	0.03	0.02	0.01	-0.40	0.65
Chromium	1,01	1.13	1.02	1.03	1.12	1.71	1.17	0.27	0.09	0.83
Copper	0.17	0.24	0.18	0.18	0.16	0.27	0.20	0.04	0.22	0.53
Iron	351	422	375	381	368	596	416	91	0.09	0.83
Lead	0.15	0.15	0.15	0.15	0.14	0.20	0.16	0.02	0.03	0.8
Mercury	0.001	0.002	0.15	0.001	0.001	0.20	0.001	0.0005	0.11	-0.3
Nickel	0.51	0.62	0.55	0.001	0.50	0.90	0.60	0.0005		0.7
									0.17	
Selenium	0.01	0.01	0.01	0.01	0.01	0.02	0.01	0.00	0.00	0.0
Silver	0.004	0.004	0.004	0.004	0.003	0.006	0.004	0.001	0.66	0.6
Tin ~:	0.02	0.02	0.02	0.02	0.03	0.03	0.03	0.01	-0.05	0.7
Zinc	0.89	1.11	1.01	1.02	0.94	1.61	1.09	0.26	0.09	8.0
Complex Organics										
DOTs	0.18	0.10	0.11	0.06	0.09	0.15	0.11	0.04	0.87	-0.3
HCH	NO	NO	ND	ND	ND	NO	ND	ND	0.00	0.0
Chlordane	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.0
Aldrin	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.0
Dieldrin	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.0
Heptachlor	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.0
Heptachlor epoxide	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.0
Mirex	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.0
Hexachlorobenzene	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.0
PCBs	ND	0.05	0.07	ND	ND	ND	0.02	0.04	-0.03	-0.3
Aroclors	ND	0.40	0.53	ND	ND	ND	0.18	0.28	-0.03	-0.3
PAHs	1.73	1.61	1.01	1.85	4.23	1.75	2.03	1.12	0.12	-0.
1-Methylnaphthalene	ND	0.04	ND	0,05	0.04	ND	0.02	0.02	-0.74	0.1
1-Methylphenanthrene	0.09	0.06	0,06	ND	0.08	0.08	0.06	0.03	0.55	-0.0
2,3,5-Trimethylnaphthalene	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.0
2,6-Dimethylnaphthalene	0.05	0.06	0.05	ND	0.08	80.0	0.05	0.03	0.31	0,5
2-Methylnaphthalene	0.04	0.06	0.05	0.05	0.07	0.06	0,06	0,01	-0.19	0.6
Acenaphthene	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0,0
Benz[a]anthracene	0.11	0.08	0.06	0.20	0.38	0.10	0,15	0,12	-0.46	0.3
Benzo[b]fluoranthene	0.21	0.22	0.14	0.19	0.43	0.21	0.23	0.10	0.00	0.1
Benzo[e]pyrene	0.52	0.41	0.29	0.11	0.43	0.35	0.35	0.14	0.32	-0.3
Benzo[g,h,i]perylene	0.26	0.19	0.13	0.04	0.38	0.20	0.20	0.11	0.23	0.0
Biphenyl	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.0
Fluoranthene	0.29	0.23	0.08	0.11	0.77	0.27	0.29	0.25	0.14	0.1
Naphthalene	0.06	0.08	0.06	0.06	0.09	0.10	0.08	0.02	0.31	0.7
Perylene	1.38	1.17	0.62	0.10	0.54	1.10	0.82	0.48	0.81	-0.6



10-5 Sediment contaminant concentrations normalized to % total organic carbon (TOC) in the Goleta survey area. Correlations by nonparametric Spearman's rho.

		_	Sediment	Stations					Correl	
Constituent	B1	B2	B3	B4	B5	86	Mean	S.D.	Outfall	Point
Undifferentiated Organics	e									
Oil and Grease	809	587	744	455	449	604	608.0	147.1	0.35	-0.35
TKN	458	534	854	1078	1121	1060	850.8	290.8	-0.06	0.63
AVS	4.79	4.03	11.44	8.33	7.12	5.98	6.95	2.69	-0.41	0.06
Heavy Metals										
Aluminum	9236	9645	13936	17808	25082	19281	15831	6112	-0.08	0.51
Antimony	0.13	0.11	0.18	0.22	0.23	0.19	0.18	0.05	-0.21	0.38
Arsenic	4.82	5.46	9.17	10.41	11.51	9.52	8.48	2.72	-0.23	0.49
Cadmium	0.35	0.46	0.79	0.76	0.82	0.89	0.68	0.22	0.05	0.74
Chromium	25.32	27.22	42.08	50.75	67.80	55.57	44.79	16.59	-0.06	0.56
Copper	4.20	5.83	7.34	8.88	9.66	8.71	7.44	2.09	-0.14	0.62
iron	8818	10204	15426	18783	22268	19395	15816	5364	-0.08	0.60
Lead	3.78	3.73	6.28	7.49	8.46	6.64	6.06	1.94	-0.24	0.43
Mercury	0.018	0.051	0.036	0.049	0.051	0.041	0.041	0.0129	-0.49	0.36
Nickel	12.77	15.06	22.60	26.45	30.37	29.31	22.76	7.40	0.02	0.69
Selenium	0.20	0.23	0.37	0.41	0.41	0.52	0.36	0.12	0.28	0.85
Silver	0.092	0.097	0.180	0.204	0.186	0.182	0.157	0.049	-0.18	0.54
Tin	0.56	0.57	0.88	1.15	1.80	1.12	1.02	0.46	-0.22	0.36
Zinc	22.28	26.75	41.39	50.19	56.88	52.34	41.64	14.26	-0.05	0.64
Complex Organics										
DOTs	4.59	2.50	2.20	1,50	2.30	2.60	2.62	1.04	0.45	-0.44
нсн	NO	ND	ND	ND	ND	ND	ND	ND	0.00	0.00
Chiordane	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0,00
Aldrin	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.00
Dieldrin	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.00
Heptachlor	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.00
Heptachlor epoxide	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.00
Mirex	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.00
Hexachlorobenzene	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.00
PCBs	ND	1.17	2.80	ND	ND	ND	0.66	1.15	-0.33	-0.13
Aroclors	ND	9,71	22,00	ND	ND	ND	5.28	9.06	-0.33	-0.13
Total PAHs	43.57	38.83	41,40	91.02	255.35	56.91	87.85	84.31	-0.16	0.05
1-Methylnaphthalene	ND	0.97	ND	2.24	2.56	ND	0.96	1.18	-0.65	-0.10
1-Methylphenanthrene	2.35	1.36	2.40	ND	4.88	2.73	2.74	1.30	0.11	0.12
2,3,5-Trimethylnaphthalene	ND	ND	ND	ND	ND	ND	NO	ND	0.00	0.00
2,6-Dimethylnaphthalene	1.33	1,36	2.20	ND	4.65	2.73	2.04	1,58	0.07	0.29
2-Methylnaphthalene	1.12	1.55	2.00	2.45	4.19	2.00	2.22	1.06	-0.42	0.19
Acenaphthene	ND	ND	ND	ND	ND	ND	ND	ND	0.00	0.00
Benz[a]anthracene	2.65	2.04	2.60	9.80	22.79	3.27	7.19	8.17	-0.46	-0.02
Benzo[b]fluoranthene	5.31	5.24	5.80	9.59	26.05	6.73	9.79	8.13	-0.393	0.020
Benzo[e]pyrene	12.96	9.90	12.00	5.31	26.05	11.27	12.91	6.97	-0.158	-0.065
Benzo[g,h,i]perylene	6.43	4.56	5.40	2.04	22.79	6.55	7.96	7.45	-0.198	0.006
Biphenyl	ND	ND	ND	ND	ND	ND	NO	ND	0.00	0.00
Fluoranthene	7.24	5.63	3.40	5.31	46.74	8.91	12.87	16.70	-0.262	0.021
Naphthalene	1.63	2.04	2.40	3.06	5.58	3.27	3.00	1.41	-0.217	0,338
Perylene	34.80	28.25	25.40	5.10	32.79	35.82	27.03	11.45	0.563	0.139

Bold = marginally significant (0.05 **Bold** = significant (p < 0.05)



10-6. Sediment chemistry minimum detection limits (MDL) and reporting limits (RL) and methods.

Parameter	MDL	RL	Units	Method	Parameter	MDL	RL	Units	Method
General Chemistry					Polynuclear Aromatic Hydr	ocarbons	(Conti	inued)	
Acid Volatile Sulfides	0.05	0.1	μg/g	Plumb, 1981 and TERL	Fluorene	1	5	ng/g	EPA8270C
Oil & Grease	100	200	µg/g	SM 5520 E	Indeno[1,2,3-c,d]pyrene	1	5	ng/g	EPA8270C
Percent Solids	0.1	0.1	%	SM 2540B	Naphthalene	1	5	ng/g	EPA8270C
TKN	0.6	5	ha/a	EPA351.3	Perylene	1	5	ng/g	EPA 8270C
Total Organic Carbon	100	200	µg/g	GC-01-111	Phenanthrene	1	5	ng/g	EPA8270C
Trace Motals					Pyrene	1	5	ng/g	EPA8270C
Auminum	1	5	pg/g	EPA 6020	Polychlorinated Blphenyls (PCB's)			
Antimony	0.025	0.05	pg/g	EPA 6020	PCB003	1	5	ng/g	EPA8270C
Arsenic	0.025	0.05	pg/g	EPA 6020	PCB008	1	5	ng/g	EPA8270C
Cadmium	0.0025	0.005	μg/g	EPA 6020	PCB018	1	5	ng/g	EPA 8270C
Chromium	0.0025	0.005	µg/g	EPA 6020	PCB028	1	5	ng/g	EPA 8270C
Copper	0.0025	0.005 5	ng/a	EPA 6020	PCB031	1	5	ng/g	EPA8270C
Lead	0.0025	0.005	µg/g	EPA 6020	PCB033	1	5	ng/g	EPA 8270C
Mercury	0.0025	0.0000	µg/g	EPA 6020 EPA 245.7	PCB037 PCB044	1	5	ng/g	EPA 8270C
Nickel	0.000	0.00002	µg/g			1	5	ng/g	EPA8270C
Selenium	0.025	0.02	µg/g	EPA6020	PCB049	1	5	ng/g	EPA8270C
Silver	0.025	0.03	ha/a	EPA6020 EPA6020	PCB052	1	5	ng/g	EPA8270C
Tin	0.025	0.02	hā/ā	EPA6020	PCB056(060) PCB066	1	5 5	ng/g	EPA 8270C
Zinc	0.025	0.05			1			ng/g	EPA8270C
Chlorinated Pesticides	0.025	0.05	µg∕g	EPA 6020	PCB070 PCB074	1	5 5	ng/g	EPA 8270C
2,4'-DDD	1	5	na/a	EPA8270C	PCB074 PCB077	1		ng/g	EPA 8270C
2,4'-DDE	1	5	ng/g ng/g	EPA8270C EPA8270C	PCB077 PCB081	1	5 5	ng/g	EPA 8270C
2,4'-DDT	1	5	ng/g	EPA8270C	PCB081 PCB087	1	5	ng/g	EPA 8270C
4,4'-DDD	1	5		EPA8270C	PCB095		5	ng/g	EPA 8270C
4.4'-DDE	1	5	ng/g ng/g	EPA 8270C	PCB097	1	5	ng/g	EPA8270C EPA8270C
4,4'-DDT	1	5	ng/g	EPA 8270C	PCB099	1	5	ng/g	
Aldrin	1	5	ng/g	EPA 8270C	PCB101	1	5	ng/g	EPA 8270C EPA 8270C
BHC-alpha	1	5	ng/g	EPA8270C	PCB105	1	5	ng/g	EPA 8270C
BHC-beta	1	5	ng/g	EPA8270C	PCB103	1	5	ng/g	EPA 8270C
BHC-delta	1	5	ng/g	EPA8270C	PCB114	1	5	ng/g ng/g	EPA8270C
BHC-gamma	1	5	ng/g	EPA8270C	PCB118	1	5	ng/g	EPA8270C
Chlordane-alpha	1	5	ng/g	EPA 8270C	PCB119	1	5	ng/g	EPA8270C
Chlordane-gamma	1	5	ng/g	EPA 8270C	PCB123	1	5	ng/g	EPA8270C
cis-Nonachtor	1	5	ng/g	EPA8270C	PCB126	1	5	ng/g	EPA 8270C
Dieldrin	1	5	ng/g	EPA8270C	PCB128	1	5	ng/g	EPA 8270C
Endosulfan sulfate	1	5	ng/g	EPA8270C	PCB138	1	5	ng/g	EPA 8270C
Endosulfan-I	1	5	ng/g	EPA8270C	PCB141	1	5	ng/g	EPA8270C
Endosulfan-II	1	5	ng/g	EPA 8270C	PCB149	1	5	ng/g	EPA8270C
Endrin	1	5	ng/g	EPA8270C	PCB151	1	5	ng/g	EPA8270C
Endrin aldehyde	1	5	ng/g	EPA8270C	PCB153	1	5	ng <i>l</i> g	EPA8270C
Endrin ketone	1	5	ng/g	EPA 8270C	PCB156	1	5	ng/g	EPA8270C
Heptachlor	1	5	ng/g	EPA 8270C	PCB157	1	5	ng/g	EPA8270C
Heptachlor epoxide	1	5	ng/g	EPA 8270C	PCB158	1	5	ng/g	EPA 8270C
Methoxychlor	1	5	ng/g	EPA 8270C	PCB167	1	5	ng/g	EPA 8270C
Mirex	1	5	ng/g	EPA8270C	PCB168/132	1	5	ng/g	EPA8270C
Oxychlordane	1	5	ng/g	EPA8270C	PCB169	1	5	ng/g	EPA 8270C
Perthane	5	10	ng/g	EPA8270C	PCB170	1	5	ng/g	EPA 8270C
trans-Nonachlor	1	5	ng/g	EPA8270C	PCB174	1	5	ng/g	EPA8270C
Polynuclear Aromatic Hydrod	arbons (F	PAHs)			PCB177	1	5	ng/g	EPA8270C
1-Methylnaphthalene	1	5	ng/g	EPA 8270C	PCB180	1	5	ng/g	EPA8270C
1-Methylphenanthrene	1	5	ng/g	EPA 8270C	PCB183	1	5	ng/g	EPA8270C
2,3,5-Trimethylnaphthalene	1	5	ng/g	EPA 8270C	PCB187	1	5	ng/g	EPA 8270C
2,6-Dimethylnaphthalene	1	5	ng/g	EPA 8270C	PCB189	1	5	ng/g	EPA 8270C
2-Methylnaphthalene	1	5	ng/g	EPA 8270C	PCB194	1	5	ng/g	EPA 8270C
Acenaphthene	1	5	ng/g	EPA8270C	PCB195	1	5	ng/g	EPA 8270C
Acenaphthylene	1	5	ng/g	EPA8270C	PCB199(200)	1	5	ng/g	EPA8270C
Anthracene	1	5	ng/g	EPA 8270C	PCB201	1	5	ng/g	EPA8270C
Benzjajanthracene	1	5	ng/g	EPA8270C	PCB206	1	5	ng/g	EPA8270C
Benzo(a)pyrene	1	5	ng/g	EPA 8270C	PCB209	1	5	ng/g	EPA 8270C
Benzo[b]fluoranthene	1	5	ng/g	EPA 8270C	Polychiorinated Biphenyls (PCB's)			
Benzo(e)pyrene	1	5	ng/g	EPA 8270C	Aroclor 1016	1	10	ng/g	EPA 8270C
Benzo[g,h,i]perylene	1	5	ng/g	EPA 8270C	Arodor 1221	1	10	ng/g	EPA 8270C
Benzo(k)fluoranthene	1	5	ng/g	EPA8270C	Arodor 1232	1	10	ng/g	EPA 8270C
Biphenyl Chosene	1	5	ng/g	EPA8270C	Arodor 1242	1	10	ng/g	EPA8270C
Chrysene Dibenz(a,h)anthracene	1	5 5	ng/g ng/g	EPA 8270C EPA 8270C	Aroclor 1248 Aroclor 1254	1	10 10	ng/g ng/g	EPA8270C EPA8270C
Dibenzothiophene	1	5	ng/g	EPA8270C	Aroclor 1260	1	10	ng/g ng/g	EPA8270C
Fluoranthene	1	5	ng/g	EPA 8270C			. *	9.8	222.00



10-7. Sediment chemistry complex organic derivatives.

				_		
Sediment Stations	B1	B2	В3	B4	B5	B6
DDTs (ng/g)						
2,4'-DDD	١ , ,					
1 -	0.0	0.0	0.0	0.0	0.0	0.0
2,4'-DDE	0.0	0.0	0.0	0.0	0.0	0.0
2,4'-DOT	0.0	0.0	0.0	0.0	0.0	0.0
4,4'-DOD	0.0	0.0	0.0	0.0	0.0	0.0
4,4'-DDE	4.5	2.5	2.2	1.5	2.3	2.6
4,4'-DOT	0.0	0.0	0.0	0.0	0.0	0.0
Sum =	4.5	2.5	2.2	1.5	2.3	2.6
Chlordane (ng/g)						
Chlordane-alpha	0.0	0.0	0.0	0.0	0.0	0.0
Chlordane-gamma	0.0	0.0	0.0	0.0	0.0	0.0
cis-Nonachlor	0.0	0.0	0.0	0.0	0.0	0.0
trans-Nonachlor	0.0	0.0	0.0	0.0	0.0	0.0
None	0.0	0.0	0.0	0.0	0.0	0.0
Sum =	0.0	0.0	0.0	0.0	0.0	0.0
1						
HCH (ng/g)						
HCH-alpha	0.0	0.0	0.0	0.0	0.0	0.0
HCH-beta	0.0	0.0	0.0	0.0	0.0	0.0
HCH-delta	0.0	0.0	0.0	0.0	0.0	0.0
HCH-gamma	0.0	0.0	0.0	0.0	0.0	0.0
Sum =	0.0	0.0	0.0	0.0	0.0	0.0
	0.0	0.0	0.0	0.0	0.0	0.0
Polychlorinated Biphenyls	l					
(PCB's, ng/g)	l					
PCB003	0.0	0.0	0.0	0.0	0.0	0.0
PCB008	0.0	0.0	0.0	0.0	0.0	0.0
PCB018	0.0	0.0	0.0	0.0		
PCB028	0.0	0.0	0.0		0.0	0.0
PC8031	0.0			0.0	0.0	0.0
PCB033	ı	0.0	0.0	0.0	0.0	0.0
PCB037	0.0	0.0	0.0	0.0	0.0	0.0
PC8044	0.0	0.0	0.0	0.0	0.0	0.0
	0.0	0.0	0.0	0.0	0.0	0.0
PCB049 PCB052	0.0	0.0	0.0	0.0	0.0	0.0
	0.0	0.0	0.0	0.0	0.0	0.0
PC8056(060)	0.0	0.0	0.0	0.0	0.0	0.0
PC8066	0.0	0.0	0.0	0.0	0.0	0.0
PCB070	0.0	0.0	0.0	0.0	0.0	0.0
PCB074	0.0	0.0	0.0	0.0	0.0	0.0
PC8077	0.0	0.0	0.0	0.0	0.0	0.0
PCB081	0.0	0.0	0.0	0.0	0.0	0.0
PCB087	0.0	0.0	0.0	0.0	0.0	0.0
PCB095	0.0	0.0	0.0	0.0	0.0	0.0
PC8097	0.0	0.0	0.0	0.0	0.0	0.0
PCB099	0.0	0.0	0.0	0.0	0.0	0.0
PCB101	0.0	0.0	0.0	0.0	0.0	0.0
PC8105	0.0	0.0	0.0	0.0	0.0	0.0
PC8110	0.0	1.2	1.4	0.0	0.0	0.0
PCB114	0.0	0.0	0.0	0.0	0.0	0.0
PCB118	0.0	0.0	0.0	0.0	0.0	0.0
PCB119	0.0	0.0	0.0	0.0	0.0	0.0
PCB123	0.0	0.0	0.0	0.0	0.0	0.0
PCB126	0.0	0.0	0.0	0.0	0.0	0.0

Sediment Stations	B1	B2	В3	B4	B5	B6
Polychlorinated Biphenyls						
(PCB's, ng/g) PCB128	0.0	0.0	0.0	0.0	0.0	0.0
PCB138	0.0	0.0	0.0	0.0	0.0	0.0
PCB141	0.0	0.0	0.0	0.0	0.0	0.0 0.0
PCB149	0.0	0.0	0.0	0.0	0.0	0.0
PCB151	0.0	0.0	0.0	0.0	0.0	0.0
PCB153	0.0	0.0	0.0	0.0	0.0	0.0
PCB156	0.0	0.0	0.0	0.0	0.0	0.0
PCB157	0.0	0.0	0.0	0.0	0.0	0.0
PCB158	0.0	0.0	0.0	0.0	0.0	0.0
PCB167	0.0	0.0	0.0	0.0	0.0	0.0
PCB168/132	0.0	0.0	0.0	0.0	0.0	0.0
PCB169	0.0	0.0	0.0	0.0	0.0	0.0
PCB170	0.0	0.0	0.0	0.0	0.0	0.0
PCB174	0.0	0.0	0.0	0.0	0.0	0.0
PCB177	0.0	0.0	0.0	0.0	0.0	0.0
PCB180	0.0	0.0	0.0	0.0	0.0	0.0
PCB183	0.0	0.0	0.0	0.0	0.0	0.0
PCB187	0.0	0.0	0.0	0.0	0.0	0.0
PCB189	0.0	0.0	0.0	0.0	0.0	0.0
PCB194	0.0	0.0	0.0	0.0	0.0	0.0
PCB195	0.0	0.0	0.0	0.0	0.0	0.0
PCB199(200)	0.0	0.0	0.0	0.0	0.0	0.0
PCB201	0.0	0.0	0.0	0.0	0.0	0.0
PCB206	0.0	0.0	0.0	0.0	0.0	0.0
PCB209	0.0	0.0	0.0	0.0	0.0	0.0
Sum =	0.0	1.2	1.4	0.0	0.0	0.0
	Ì					
Aroclors						
Aroclor 1016	0.0	0.0	0.0	0.0	0.0	0.0
Arocior 1221	0.0	0.0	0.0	0.0	0.0	0.0
Aroclor 1232	0.0	0.0	0.0	0.0	0.0	0.0
Aroclor 1242 Aroclor 1248	0.0	0.0	0.0	0.0	0.0	0.0
Aroclor 1254	0.0	0.0	0.0	0.0	0.0	0.0
Aroclor 1260	0.0	10.0	11.0	0.0	0.0	0.0
Sum =	0.0 0.0	<u>0.0</u> 10.0	<u>0.0</u> 11.0	0.0	0.0	0.0
-	0.0	10.0	11.0	0.0	0.0	0.0
Polynuclear Aromatic						
Hydrocarbons (PAH's, ng/g)						
Acenaphthylene	0.0	0.0	0.0	1.2	0.0	0.0
Anthracene	1.3	0.0	0.0	4.4	1.6	0.0
Benz[a]anthracene	2.6	2.1	1.3	4.8	9.8	1.8
Benzo[a]pyrene	3.1	3.0	1.4	5.9	10.3	3.0
Benzo[b]fluoranthene	5.2	5.4	2.9	4.7	11.2	3.7
Benzo[g,h,i]perylene	6.3	4.7	2.7	1.0	9.8	3.6
Benzo[k]fluoranthene	3.0	3.4	2.1	3.4	10.6	3.0
Chrysene	5.1	5.9	3.4	2.3	15.1	3.1
Dibenz[a,h]anthracene	2.0	1.1	0.0	2.2	2.2	1.1
Fluorene	0.0	0.0	0.0	0.0	1.4	1.0
Indeno[1,2,3-c,d]pyrene	2.6	2.5	2.1	10.8	6.8	2.2
Phenanthrene	3.8	4.6	2.4	0.0	12.3	4.0
Pyrene	<u>7.7</u>	<u>7.3</u>	2.4	3.9	18.7	4.8
Sum =	42.7	40.0	20.7	44.6	109.8	31.3



10.6. Benthic Infauna



10-8. Benthic infauna taxonomic abundances.

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March 2015

Appendix

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10-8. Continued.

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	Cylicalia diegensis	•								•									•		
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	Epitonium sawinae	7			-							-							7		
	Eulima raymondi																-				
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Nomortos	elacas	Carinoma mutabilis	-	,		,	_			-		-			L		-		-				-				-	_
		Cerebratulus californiensis	<u> </u>	4	ď						2		1	2 1	_		٦,	7	_		2		-	г	-	3	н	
		Cerebratulus marginatus		-	•				,	٠ _	_												l	ı				
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		Tubulanus polymorphus	ч	9	2	2 3	m	7	4	2	m	1	æ	2 5	4	1	S	4	m	,	4 2	2	н	13	9	-	ო	
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		Zygeupolia rubens	7				-												_					4	ч	m	П	
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		letrastemma albidum				1																						
		Tetrastemma candidum																,	_		7							
		Tetrastemma nigrifrons					_											-	_									
		Tetrastemma reticulatum	4												4				7									
Phorona	None	Phoronis sp	ი	m	4	9	н	7	1 4	7	-1		80	10 5	17	6	m	7	7	9	8	7	^	7	80	2 2	9	
Platyhelminthes	Turbellaria	Cryptocelis occidentalis			7	1											1				7							
		Phaenoplana longipenis								7																		
		Stylochus exiguns	m	7			_							7					+				-1					
Sipuncula	Phascolosomatidea	Apionsoma misakianum	1	1	-	2						н			_				+									
	Sipunculidea	Thysanocardia nigra													_		ч		_									
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Appendix

Phylum Class Total Misc. Phyla 1342 % of Population 4.36	Class			3 83	12	•	85	A CARLO	ði Æ	3.3		-		84 E	'n	1	8 %	5	120	2	3 4	N.P
otal Misc. F	Class		M,						The same of the			-			1.3	1.0	8-21		- 3-4	20		
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Misc. F Popula		Scaphopoda										-			7			7	-			
	Phyla 1342 ation 4.36																					
	Inarticulata	Glottidia albida	15 6	6	5	22 11	8	15 25	56	39 30	37	30	13	11 13	77	7 8	8	6 3	26	15 2	26 14	12
	Ascidiacea	Moleula manhattensis										_					7		_			
	Enteropneusta	Schizocardium sp	2		-												Э	3 1	6	7		
Cnidaria	Anthozoa	Arachnanthus sp A	1 3		1		-1	3	7	2	4	3	1	1 1	1							1
		Ceriantharia	3 1	1	1	1 2		2 1	7	2 3		-		1		1		3 1	m	н	~	
		Edwardsia juliae	7			1		2				4					е					
		Edwardsiidae				_			-	1 1												
		Halianthella sp A						1														
		Pachycerianthus sp															-					
		Scolanthus scamiti	-																			
	Hydrozoa	Euphysa sp A	1									_		1								
	Echiurida	Listriolobus pelodes			1							+		н		1	7		4			
Nematoda		Nematoda										7	œ			1 5		7			_	
Nemertea	Anopla	Carinoma mutabilis	4	7	1 1				п			1	1	2 1			1	3				
		Cerebratulus californiensis	3 4	7	1 2	_	7	1	m	2 1		-	7	2 1	-	6 1	æ	1 2	_	7	-	
		Cerebratulus marginatus			1	_			1				1				п					
		Heteronemertea sp HYP1	1													Э	7	-				
		Uneidae	9		2 1				п	1	7	-	-	7		4	4	3 1	_			
		Lineus bilineatus	2 5	е	2	1	2	1	4		1	-		7	7	1	п	2 1	_		2	7
		Lineus sp										1					1					
		Lineus sp HYP1						1						,								
		Palaeonemertea	2	ю				,				_		7	_	-	7					
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		Tubulanidae sp 8												•	,	7 ,	٦.					
		Tubulanidae sp E	6 11	S	5	8	4	9	2	3	7	3		1	7	4	'n		^	n	η.	
		Tubulanus cingulatus					п.	,				_	,			,	,	,	,	,		
		Tubulanus polymorphus	1 1			-	-	2 2	7	1	7	_	4	n		7 (η.	٠,	7	7		
		Tubulanus sp A										_	٠	-		4	-					
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		Amphipone privales		,	•	,	-					_			-	•	•					
		Signature in the second					•					_			,				-			
		Cryptonementes akunopinia															2		1			
		Paranemertes californica	9	2	2	2	00	3	9	7 3	4	3	6	6 10	80	4	ю	3	œ	m	6 1	7
		Tetrastemma albidum						7														
		Tetrastemma bilineatum										_						1	_			
		Tetrastemma candidum				1				1		_		1								
		Tetrastemma nierifrons			m									7	1	E		2 1	_			
		Zvoonemertes virestrans			1				_				2		-							
Phorona	None	Phoronis sp	12 14	12	8	4 6	-	5 1	3 1	11 1	1	5	10	2 1	-	m	2	s	7	н		
ninthes	Turbellaria	Cryptocelis occidentalis										1										
		Koinostylochus burchami											1		1	1	7					
		Leptoplanidae												1	7	1 5	1	æ				
		Phaenoplana longipenis										_	7	1		æ	8	2 1	_			
		Stylochoplana sp										4			_							
		Stylochus exiguus		7		1 2			г					1 1				1	-			



10.7. Fish and Invertebrate Abundance and Biomass



10-9. Fish abundance by size class (cm) for each replicate trawl.

		ī				
		Size	├ ┰	Abune B3		36
Scientific Name	Common Name	Class (cm)	1	2	1	2
Citharichthys sordidus	Pacific sanddab	8	1		·	
•		9	4			
		10	4			
		11	2			
Citharichthys stigmaeus	speckled sanddab	12_	8	10	1	11
Olinonominys stigmadus	Speckled Stilledab	6	29	36	1	28
		7	8	8	1	15
		8	5	5	1	
		9 10	7 2	8		7
		11	*			3
Cymalogaster aggregata	shiner perch	7			3	
		8			3	
		9			1	
		11			2	
		12			3	
		13			2	
		14			6	
		15			2	
		16			1	
Gibbonsia evides	spotted kelpfish	18 7	1		1	
Gibbonsia montereyensis	crevice kelpfish	5	⊢'		1	
		6	L		Ľ	2
Heterostichus rostretus	giant kelplish	6			1	5
		7	3		1	5
		8 9	2	1	5 4	1
		10	1	1	*	
		11	Ι΄.	'		1
		12	L		1	
Hypsurus caryi	rainbow seaperch	7			1	
		9		1	4	
		10 13		1	5	
		15			1	
Icelinus quadriseriatus	yellowchin sculpin	4		1	Ė	
		5	1			
Odontopyxis trispinosa	pygmy poacher	7	1			
Paraiabrax clathratus	kelp bass	5			1 3	4
		6			5	2
		7			9	2
Paralabrax nebulifer	barred sand bass	17	L.			1
Phanerodon furcatus Pleuronichthys coenosus	white seaperch C-O sole	13 18	1			1
Pleuronichthys decurrens	curlfin sole	B	1	1		<u> </u>
		9	1	1		
		13	1			
		14	2	1		
Porichthys notatus	plainfin midshipman	15	1			1
Sebastes auriculatus	brown rockfish	10	1		\vdash	
Sebastes dallii	calico rockfish	7	Ľ		1	
Sebastes minietus	vermilion rockfish	8		1		
Squatina californica	Pacific angel shark	31	.			1
Symphurus atricaudus Syngnathus californiensis	California tonguelish	11	1	1	-	
Gyngrauius Camoniaensis	kelp pipefish	17	1	'		
		18	Ι.			1
		19			1	1
		20	1		1	1
		21			1	
		23 24	1	1	1	1
		25	1	'	1	
		26	1		1	
Synodus lucioceps	Californía lizardfish	11	ΤĖ	3	1	
		12	2	3	2	1
		13	1	3		
		14 15	1		1	1
		17	1		2	•
		18	1		-	1
		22	1		1	
		24	1		1	
		25	1.		2	
		28 29	1	1		
		31	1 .	, '		
		33			1	
Xystreurys liolepis	fantail sole	17		1		
Zalembius rosaceus	pink seaperch	7		1		
		8 9		1	42	
		10			42	
		13			1	



10-10. Fish biomass (Kg) by replicate.

			Weig	ht (kg)	
Scientific Name	Common Name	1	3		T6
		1	2	1	2
Citharichthys sordidus	Pacific sanddab	0.16			
Citharichthys stigmaeus	speckled sanddab	0.27	0.3	<0.1	0.33
Cymatogaster aggregata	shiner perch			1.08	
Gibbonsia evides	spotted kelpfish	<0.1			
Gibbonsia montereyensis	crevice kelpfish			<0.1	<0.1
Heterostichus rostratus	giant kelpfish	<0.1	< 0.1	<0.1	<0.1
Hypsurus caryi	rainbow seaperch	1	< 0.1	0.27	
Icelinus quadriseriatus	yellowchin sculpin	<0.1	< 0.1		
Odontopyxis trispinosa	pygmy poacher	<0.1			
Paralabrax clathratus	kelp bass			0.11	<0.1
Paralabrax nebulifer	barred sand bass				0.13
Phanerodon furcatus	white seaperch	<0.1			
Pleuronichthys coenosus	C-O sole				0.13
Pleuronichthys decurrens	curlfin sole	0.34	0.11		<0.1
Porichthys notatus	plainfin midshipman				<0.1
Sebastes auriculatus	brown rockfish	<0.1			
Sebastes dallii	calico rockfish			<0.1	
Sebastes miniatus	vermilion rockfish	- 1	<0.1		
Squatina californica	Pacific angel shark	i			0.31
Symphurus atricaudus	California tonguefish	<0.1			
Syngnathus californiensis	kelp pipefish	<0.1	<0.1	<0.1	<0.1
Synodus lucioceps	California lizardfish	0.55	0.35	0.95	0.1
Xystreurys liolepis	fantail sole	0.12			
Zalembius rosaceus	pink seaperch		<0.1	1.0	
	composite	0.14	0.12	<0.1	<0.1
	S	um 1.58	0.88	3.41	1.00



10-11. Invertebrate abundances by replicate.

			Abun	dance	
Scientific Name	Common Name	Т	B3	Т	B6
		1	2	1	2
Aplysia californica	purple sea hare			1	1
Conus californicus	California cone				2
Crangon nigromaculata	blackspotted bay shrimp				2
Loxorhynchus grandis	sheep crab	4		3	4
Lytechinus pictus	white sea urchin			2	
Metacarcinus gracilis	graceful rock crab	1			
Octopus rubescens	red octopus	2	2	2	
Pugettia producta	northern kelp crab			1	
Pyromaia tuberculata	tuberculate pear crab				1
Sicyonia penicillata	peanut rock shrimp	15	19	2	3
	Sum	22	21	11	13

10-12. Invertebrate biomass (Kg) by replicate.

			Weig	ht (kg)	
Scientific Name	Common Name	Т	B3	TI	B6
		1	2	1	2
Aplysia califomica	purple sea hare			<0.1	0.8
Conus californicus	California cone				< 0.1
Crangon nigromaculata	blackspotted bay shrimp				<0.1
Loxorhynchus grandis	sheep crab	1.1		1.0	0.43
Lytechinus pictus	white sea urchin			<0.1	
Metacarcinus gracilis	graceful rock crab	0.17			
Octopus rubescens	red octopus	<0.1	<0.1	<0.1	
Pugettia producta	northern kelp crab			0.25	
Pyromaia tuberculata	tuberculate pear crab				<0.1
Sicyonia penicillata	peanut rock shrimp	0.1	0.16	<0.1	<0.1
	composite			<0.1	<0.1
	Sum	1.37	0.16	1.25	1.23



10.8. Fish and Bivalve Bioaccumulation Data



10-13. Whole weight, tissue weight and standard length of fish.

		ION TB3				ON TB6	
Standard		Muscle Weight		Standard		Muscle Weight	-
Length (mm) 50	(g)	(g)	(g)	Length (mm)	(g)	(g)	(g)
50 52	2 3	0.4	0.18	57	4	0.5	0.05
52	2	0.4 0.4	0.06	57 50	3	0.3	0.02
52	3	0.4	0.12 0.17	59 60	4 4	0.6	0.04
53	3	0.5	0.17	61	5	0.6 0.9	0.00 0.06
54	2	0.6	0.09	61	4	0.9	0.06
54	3	0.6	0.03	61	4	0.5	0.04
55	3	0.6	0.01	61	4	0.5	0.03
55	3	0.6	0.02	61	4	0.6	0.02
56	3	0.6	0.01	62	4	0.5	0.05
56	3	0.5	0.02	65	5	0.8	0.06
56	3	0.6	0.14	65	5	1.0	0.04
57	3	0.7	0.14	65	4	0.5	0.01
57	4	0.8	0.18	67	5	1.1	0.04
58	3	0.7	0.02	67	5	0.8	0.00
58	3	0.8	0.12	68	5	1.2	0.05
59	3	0.7	0.02	74	9	1.8	0.11
60	4	0.7	0.05	75	8	1.9	0.06
60	4	0.6	0.22	82	11	2.1	0.05
61	4	0.9	0.05	84	14	2.2	0.16
62	4	0.9	0.21	86	13	2.0	0.21
65	6	1.3	0.05	88	14	1.7	0.14
67	6	1.0	0.26	90	15	2.8	0.08
68	5	1.1	0.04	90	15	2.5	0.14
68	6	0.8	0.12	91	11	1.7	0.06
68	5	0.6	0.13	91	14	2.2	0.11
70	5	1.0	0.02	92	15	2.6	0.16
70	7	1.0	0.20	92	15	2.4	0.16
70	6	0.7	0.21	93	13	2.4	0.07
71 71	5 6	0.9	0.05	93	15	2.7	0.12
73	8	1.2 1.2	0.06	95	17	3.6	0.14
73 73	8	1.0	0.02 0.22	95 05	17	3.0	0.21
74	8	1.7	0.22	95 96	15 15	2.4	0.19
74	7	0.8	0.02	98	17	2.5 2.5	0.14 0.25
74	7	1.4	0.02	100	21	3.8	0.23
75	7	1.2	0.04	102	20	3.6	0.16
76	9	1.5	0.01	102	20	0.0	0.10
77	9	1.5	0.01				
78	9	1.3	0.13				
78	10	1.2	0.31				
79	7	0.9	0.01	\			
79	8	1.2	0.03	1			
79	10	1.1	0.18				
79	10	1.1	0.24				
80	10	2.2	0.01				
80	8	1.3	0.20				
80	9	1.0	0.25	1			
81	8	1.3	0.01	l			
81	10	1.4	0.20	1			
81	8	0.9	0.21				
82	12	1.2	0.02				
85	12	2.2	0.12				
85	13	1.7	0.14				
88	9	1.9	0.05				
89 89	13 11	1.7	0.09				
90	11	1.4	0.31				
90 95	11 13	1.8 1.6	0.01 0.02				
			0.02				
Count =	Count =	Count =	Count =	Count=	Count =	Count =	Count =
59	59	59	59	37	37	37	37
Total =	Total =	Total =	Total =	Total=	Total =	Total =	Total =
4119.0	385.7	60.8	6.2	2899.0	370.3	63.4	3.5
Average =	Average =	Average =	Average =	Average =	Average =	Average =	Average =
69.81	6.54	1.03	0.10	78.35	10.01	1.71	0.09



10-14. Whole weight, tissue weight and total weight of caged bivalves.

	Control, Rep 1	27-259 P. L. C. S.		Control, Rep 2	N V N COL		Control, Rep 3	LONG TO STATE
Total Length (mm)	Total Weight (g)	Tissue Weight (g)	Total Length (mm)	Total Weight (g)	Tissue Weight (g)	Total Length (mm)	Total Weight (g)	Tissue Weight (g)
75	43.2	10.0	70	26.2	6.2	61	20.2	4.9
63	25.3	6.6	69	28.9	6.9	66	29.6	7.7
73	33.2	7.0	59	16.5	2.9	66	31.6	6.5
66	23.8	4.7	60	20.2	3.3	67	23.4	5.3
74	35.2	5.6	59	18.2	3.1	62	20.0	2.7
82	41.0	7.6	62	20.1	4.5	76	24.9	4.6
61	25.0	4.3	65	23.1	3.9	67	22.7	4.3
55	20.9	5.6	65	27.2	6.6	65	20.9	5.7
72	40.3	8.0	69	27.5	5.2	64	19.8	3.8
65	21.3	5.2	71	31.4	7.0	65	24.5	6.4
74	40.3	8.6	67	27.9	6.1	71	35.1	7.8
62	21.6	5.6	65	23.7	5.4	73	31.4	6.9
63	20.6	4.9	64	22.2	5.6	65	29.3	6.2
68	27.6	5.1	67	25.8	5.9	65	22.6	6.2
65	26.1	7.6	63	21.1	3.8	67	31.1	7.7
Count =	Count =	Count =	Count=	Count =	Count =	Count=	Count =	Count =
15	15	15	15	15	15	15	15	15
Total =	Total =	Total =	81.0	Total =	Total =	Total	Total =	Total =
1018.0	445.3	96.4	975.0	359.9	76.1	1002.0	387.0	88.8
Average =	Average =	Average =	Average =	Average =	Average =	Average =	Average =	Average =
67.9	29.7	6.4	65.0	24.0	5.1	66.8	25.8	5.9

67.9	29.7	5.4	65.0	24.0	5.1	66.8	25.8	5.9
AMMENT	B3, Rep 1		MINE STATE	B3, Rep 2		San Francis	B3, Rep 3	
Total Length (mm)	Total Weight (g)	Tissue Weight (g)	Total Length (mm)	Total Weight (g)	Tissue Weight (g)	Total Length (mm)	Total Weight (g)	Tissue Weight
65	32.4	8.0	77	51.4	19.2	65	38.0	11.8
65	21.4	6.7	79	51.9	18.2	72	46.1	10.8
65	24.9	6.5	69	30.7	8,8	66	31.1	9.0
61	23.0	7.4	76	39.9	11.6	75	34.1	11,2
65	27.5	9.2	80	48,1	12.3	76	41.2	9.7
64	22.5	5.8	75	40.B	12.1	71	39.5	12.2
64	28.0	5.7	73	38.3	9.0	68	31.6	9.3
73	36.8	8.1	76	54.9	20.8	80	69.1	23.6
67	32.7	7.4	63	45.1	16.3	67	39.0	8,7
68	30.6	7.9	72	39.7	14.3	76	52.7	14,8
69	29.6	8.5	73	35.3	11.7	73	38.3	11.4
67	32.2	11.5	69	35.9	12.4	66	37.4	11.6
63	26.4	9.8	77	42.0	16.0	77	59.5	20.2
62	30.2	10.7	78	53.6	17.5	71	45.6	13.5
70	26.0	8.5	68	31.6	11.8	70	28.9	9.7
Count =	Count =	Count =	Count=	Count =	Count =	Count=	Count =	Count =
15	15	15	15	15	15	15	15	15
Total =	Total =	Total =	81.0	Total =	Total =	81.0	Total =	Total =
988.0	424.2	121.8	1105.0	639.1	212.1	1073.0	632.0	187.4
Average =	Average =	Average =	Average =	Average =	Average ≃	Average =	Average =	Average =
65,9	28.3	8.1	73.7	42.6	14.1	71.5	42.1	12.5

	84, Rep 1			B4, Rep 2	1000		B4, Rop 3	
Total Length (mm)	Total Weight (g)	Tissue Weight (g)	Total Length (mm)	Total Weight (g)	Tissue Weight (g)	Total Length (mm)	Total Weight (g)	Tissue Weigh (g)
75	49.2	15.0	66	41.7	14.4	65	26.7	7.7
72	37.5	15.6	73	43.3	15.0	65	29.8	8.9
78	43.5	13.2	71	35.0	8.9	70	42.3	11.9
66	34.6	6.8	85	50.3	14.2	62	34.4	7.4
70	32.4	8.1	72	38.7	9.2	58	27.5	6.5
65	28.1	6.5	64	31.8	6.3	65	35.4	10.8
69	32.7	7.8	68	27.7	7.2	64	32.2	8.8
67	32.7	10.4	75	40.7	15.1	64	25.2	9.7
69	38.3	14.3	75	44.1	16.9	62	28.7	9.1
75	40.0	14.2	80	51.8	17.2	65	26.9	9.4
75	43.7	16.4	72	41.7	13.8	65	26.6	10.0
72	41.2	15.9	65	28.5	9.8	67	30.1	9.5
75	42.6	14.2	63	33.5	11.8	66	28.4	9.9
70	40.5	15.1	79	58.6	18.2	71	51.8	19.9
69	43.3	13.6	76	47.8	18.9	71	41.2	12.7
Count =	Count =	Count =	Count=	Count =	Count =	Count=	Count =	Count =
15	15	15	15	15	15	15	15	15
Total =	Total =	Total =	Total =	Total =	Total =	Total =	Total =	Total =
1067.0	580.1	189.1	1064.0	615.1	196,8	980.0	487.2	152.1
Average =	Average =	Average =	Average =	Average =	Average =	Average =	Average =	Average =
71,1	38.7	12.6	72.3	41.0	13.1	65.3	32.5	10.1

					1411			
	B6, Rep 1			B6, Rep 2			B6, Rep 3	
Total Length	Total Weight	Tissue Weight	Total Length	Total Weight	Tissue Weight	Total Length	Total Weight	Tissue Weigh
(mm)	(9)	(g)	(mm)	(g)	(g)	(mm)	(g)	(9)
79	49.2	17.6	74	41.7	14.6	83	55.4	15.3
65	31.9	11.3	78	48.8	13.6	70	46.2	10.7
71	37.B	8.5	74	43.8	18.1	71	40.3	14.5
66	36.7	10.5	64	30.1	7.6	70	37.1	11.1
77	35.5	11.1	69	42.7	15.1	73	40.0	16.3
67	35.1	10.0	68	27.8	3.1	70	42.0	11.9
90	48.2	12.7	65	36.5	13.4	77	41.2	11.9
76	35.8	8.1	78	43.8	13.8	86	48.6	17.8
82	47.4	13.8	69	28.8	10.3	89	47.1	15.6
81	47.1	19.6	76	41.1	10.5	84	55.1	14.5
80	45.8	18.2	72	42.5	15.4	85	51.9	11.6
66	32.2	8.1	64	29.9	8.4	79	43.2	14,4
70	35.5	12.0	71	40.1	12.2	80	44.1	14.4
77	28.8	7.7	67	31.9	8.0	73	48.2	15.7
71	28.1	9.4	75	44.5	14.7	77	36.3	11.7
Count =	Count =	Count =	Count=	Count =	Count =	Count=	Count =	Count =
15	15	15	15	15	15	15	15	15
Total =	Total =	Total =	Total =	Total =	Total =	Total =	Total =	Total =
1118.0	575,0	178.5	1064.0	575.8	178.5	1167.0	676.5	207.3
Average =	Average =	Average ≃	Average =	Average =	Average =	Average =	Average =	Average =
74.5	38.3	11.9	70.9	38.4	11.9	77.8	45.1	13,8



10-15. Fish and bivalve tissue concentrations by replicate for all constituents measured.

7-15. FISH and bivar	ve tissue		/uscle		Liver	Cate 101	Biva		measar
Constituent	Replicate	TB3	TB6	TB3	TB6	Control	B3	B4	В6
General Chemistry (µg/dry g)									
	#1	1.71	1.33	32.25	38.30	5.10	6.99	7.17	6.29
% Lipids (detection limit = 0.01)	#2	1.39	1.64	28.97	39.66	4.21	8.11	6.32	7.59
	#3	<u>1.63</u>	1.17	<u>NS</u>	<u>NS</u>	<u>4.72</u>	<u>5.70</u>	7.83	6.65
	Mean =	1.58	1.38	30.61	38.98	4.68	6.93	7.11	6.84
	S.D. =	0.17	0.24	2.32	0.96	0.45	1.21	0.76	0.67
Mean f	or each Tissue =	1.4	78	34.	795		6.3	90	
% Moisture (detection limit = 0.1)	#1	79.8	80.5	70.4	67.8	84.9	78.0	79.0	80.6
	#2	NS	NS	NS	NS	86.0	78.8	79.2	80.0
	#3	NS	NS	NS	NS	86.4	<u>79.4</u>	79.0	79.9
	Mean =	79.8	80.5	70.4	67.8	85.8	78.7	79.1	80.2
	S.D. =			NA	NA	0.8	0.7	0.1	0.4
Mean f	or each Tissue =	80).2	69			80		• • • • • • • • • • • • • • • • • • • •
Metals (µg/dry g)									
(detection limit = 0.025 µg/dry g)	#1	5.442	7.060	7.319	8.868	8.688	18.286	15.767	16.51
	#2	5.934	7.174	7.000	8.990	8.767	16.633	16.699	16.260
Arsenic	#2					I			
Alsenic		5.952 5.776	7.008	<u>NS</u>	NS 8 939	8.473	18.419 17.770	14.716 15.727	14.629
	Mean =	5.776	7.081	7,160	8.929	8.643	17.779	15.727	15.80
	S.D. =	0.289	0.085	0.226	0.086	0.152	0.995	0.992	1.023
Mean f	or each Tissue =	6.4	28	8.0)44		14.4	188	
Cadmium	#1	0.025	0.037	6.064	13.992	4.406	2.774	2.663	2.829
	#2	0.025	0.038	5.918	13.223	4.714	2.609	2.677	3.021
	#3	0.025	0.037	NS NS	NS	3.954	3.293	2.737	2.536
	Mean =	0.025	0.037	5.991	13.608	4.358	2.892	2.692	2.795
	S.D. =								
Mean f	0.000	0.001 31	0.103 9.7	0.544 '99	0.382	0.357 3.1	0.039 84	0.244	
Chromium	#1	0.043	0.073	0.437	0.121	1.146	1.087	1.077	1.199
	#2	0.063	0.091	0.728	0.404	1.572	0.972	0.985	1.029
	#3	0.058	0.048	<u>NS</u>	<u>NS</u>	<u>1.684</u>	<u>1.138</u>	<u>1.199</u>	0.846
	Mean =	0.055	0.071	0.583	0.263	1.467	1.066	1.087	1.025
**	S.D. =	0.010	0.022	0.206	0.200	0.284	0.085	0.107	0.177
meant	or each Tissue =	0.0	163	0.4	123		1.1	61	
Copper	#1	0.025	0.025	8.938	16.478	3.849	5.410	5.246	5.089
	#2	0.025	0.025	8.502	14.807	4.290	5.326	4.585	6.590
	#3	0.025	0.025	NS	NS	3.515	6.054	5.910	5.358
	Mean =	0.025	0.025	8.720	15.643	3.885	5.597	5.247	5.679
	S.D. =	0.000	0.000	0.308	1.182	0.389	0.398	0.663	0.800
Meanf	or each Tissue =		25		181	0.000	5.1		0.000
Lead	#1	0.037	0.059	0.770	0.933	4 722	4.045	4 402	4.054
Lodd	#2	0.037	0.059	0.770	0.933	1.732 2.113	1.215 1.094	1.123 1.061	1.255
									1.089
	#3	0.025	0.025	NS 0 685	<u>NS</u>	1.892	1.188	<u>1.191</u>	1.251
	Mean =	0.029	0.051	0.685	0.852	1.912	1.166	1.125	1.198
Mean f	S.D. = or each Tissue =	0.007	0.023)40	0.121 0.7	0.115 768	0.191	0.064 1.3	0.065 50	0.095
.7744171				3.,			1,0		
Mercury	#1	0.0917	0.0949	0.0707	0.0797	0.0498	0.0278	0.0264	0.029
(det. Limit = 0.00001 µg/dry g)	#2	0.0861	0,0925	0.0627	0.0715	0.0591	0.0254	0.0291	0.026
	#3	0.0997	0.096	NS	NS	0.0590	0.0303	0.0242	0.024
	Mean =	0.0925	0.0943	0.0667		0.0560	0.0278	0.0266	0.026
	S.D. =	0.0068	0.0016		0.0058	0.0053	0.0025	0.0025	0.002
Meant	or each Tissue =		93	1	071		0.0		
***************************************	444	0.025	0.038	0.336	0.194	0.022	0.844	0.040	0.862
		0.025		0.336	0.184	0.922	0.811	0.818	
	#1 #2	0.020			0.323	1.098	0.646	0.679	0.756
	#2	0.030	0.049	1		1	0.074		
	#2 #3	0.025	0.038	<u>NS</u>	<u>NS</u>	1.044	0.874	0.863	
Nickel	#2 #3 Mean =	0.025 0.027	0.038 0.042	<u>NS</u> 0.345	<u>NS</u> 0.254	1.044 1.021	0.777	0.863 0.787	0.747
	#2 #3	0.025	0.038	<u>NS</u>	<u>NS</u>	1.044		0.863	<u>0.624</u> 0.747 0.119



10-15. continued.

		Fish N	Muscle	Fish	Liver		Biva	ilve	
Constituent	Replicate	TB3	TB6	TB3	TB6	Control	B3	B4	В6
Metals (µg/dry g)									
Selenium	#1	1.204	1.170	5.222	4.200	2.225	2.821	2.569	3.076
	#2	1.078	1.154	5.180	4.452	2.401	2.741	2.588	3.243
	#3	1.194	1.072	<u>NS</u>	<u>NS</u>	<u>2.214</u>	2.730	2.564	2.482
	Mean =	1.159	1.132	5.201	4.326	2.280	2.764	2.574	2.934
	S.D. =	0.070	0.053	0.030	0.178	0.105	0.050	0.013	0.400
-	Mean for each Tissue =	1.	15	4.	76		2.6	38	
Silver	#1	0.025	0.025	0.273	0.196	0.239	0.153	0.146	0.214
	#2	0.025	0.025	0.228	0.177	0.147	0.135	0.117	0.165
	#3	0.025	0.025	NS	NS	0.139	0.233	0.129	0.122
	Mean =	0.025	0.025	0.251	0.187	0.175	0.174	0.131	0.167
	S.D. =	0.000	0.000	0.032	0.013	0.056	0.052	0.015	0.046
	Mean for each Tissue =	0.	03	0.	22		0.1	62	
Zinc	#1	10.35	13.48	58.66	63.57	114.42	115.38	104.90	120.06
2.110	#2	10.88	17.62	53.83	65.49	138.50	110.79	108.30	134.31
	#3	9.95	13.941	NS	NS	128.81	115.14	110.56	118.81
	Mean =	10.392	15.012	56.243	64.532	127.247	113.771	107.919	124.395
	S.D. =	0.470	2.269	3.417	1.359	12.116	2.581	2.848	8.612
	Mean for each Tissue =		.70	ı	.39		118.		
Complex Organics (ng/dr									
Total DDT ^{1.}	#1	21.1	20.5	808.6	968.7	6.3	5.4	3,6	3.7
	#2	19.7	23.4	762.1	1063.3	5.9	4.1	3.8	4.4
	#3	<u>19.1</u>	<u>24.2</u>	<u>NS</u>	<u>NS</u>	<u>5.7</u>	<u>5.5</u>	4.2	<u>4.3</u>
	Mean ≕	20.0	22.7	785.4	1016.0	6.0	5.0	3.9	4.1
	S.D. =	1.0	1.9	32.9	66.9	0.3	8.0	0.3	0.4
	Mean for each Tissue =	21	.33	900	0.68		4.	.7	
Total Chlordane ^{1.}	#1	0.0	0.0	36.3	32.5	0.0	0.0	0.0	0.0
	#2	0.0	0.0	40.0	40.3	0.0	0.0	0.0	0.0
	#3	0.0	0.0	NS	NS	0.0	0.0	0.0	0.0
	Mean =	0.0	0.0	38.2	36.4	0.0	0.0	0.0	0.0
	S.D. =	0.0	0.0	2.6	5.5	0.0	0.0	0.0	0.0
	Mean for each Tissue =	0.	.00	37	.28		0.	.0	
Total HCHs1.	#1		0.0		0.0		0.0	0.0	0.0
Iotal nons		0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
	#2 #3	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
	Mean =	0.0 0.0	<u>0.0</u> 0.0	NS 0.0	<u>NS</u> 0.0	0.0 0.0	<u>0.0</u> 0.0	<u>0.0</u> 0.0	0.0 0.0
	S.D. =	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
	Mean for each Tissue =		.0		.0	0.0		.0	0.0
Aldrin	#1	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0
(detection limit = 1.0 ng/g)		1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0
	#3	1.0	<u>1.0</u>	NS 10	<u>NS</u>	1.0	1.0	1.0	<u>1.0</u>
	Mean =	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0
	S.D. = Mean for each Tissue =	0.0	0.0	0.0	0.0 .0	0,0	0.0	0.0	0.0
	weath for each fissue =	, , ,	.0	1	.0		1.	.0	
Dieldrin	#1	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0
(detection limit = 1.0 ng/g)		1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0
0 0,	#3	1.0	1.0	NS	NS	1.0	<u>1.0</u>	1.0	1.0
	Mean =	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0
	S.D. =	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
	Mean for each Tissue =	1	.0	1	.0		1	.0	
Heptachlor	#1	1.0	1.0	10	1.0	1.0	1.0	1.0	1.0
neptachior (detection limit = 1.0 ng/q)		1.0 1.0	1.0	1.0 1.0	1.0	1.0 1.0	1.0 1.0	1.0 1.0	1.0 1.0
(uctection mitte 1.0 ng/g)	#2	1.0	1.0		1.0 NS	I			
		1.0	<u>1.0</u> 1.0	NS 1.0	<u>NS</u> 1.0	1.0 1.0	<u>1.0</u> 1.0	<u>1.0</u> 1.0	<u>1.0</u> 1.0
	Mean = S.D. =	0.0	0.0	0.0	0.0	I	0.0	0.0	
	S.D. = Mean for each Tissue =		0.0		.0	0.0		.0	0.0
	HOAIT IOI CACH TISSUE -			<u> </u>		l			

^{1.} Complex organic derivatives are listed in Table 10-16.



10-15. continued.

			Muscle		Liver			Bivalve	
Constituent	Replicate	TB3	TB6	TB3	TB6	Control	В3	B4	B6
Complex Organics (ng/dr Hexachlorobenzene	y Kg) #1	1.0	1.0	5.4	1.0	1.0	1.0	1.0	1.0
(detection limit = 1.0 ng/g)	#1	1.0	1.0	4.7	1.0	1.0	1.0	1.0	1.0
(40100110111111111111111111111111111111	#3	1.0	1.0	NS	NS	1.0	1.0	1.0	1.0
	Mean =	1.0	1.0	5.1	1.0	1.0	1.0	1.0	1.0
	S.D. =	0.0	0.0	0.5	0.0	0.0	0.0	0.0	0.0
	Mean for each Tissue =		1.0	3	.0			1.0	
Mirex	#1	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0
(detection limit = 1.0 ng/g)	#2	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0
	#3	1.0	1.0	<u>NS</u>	NS	<u>1.0</u>	1.0	<u>1.0</u>	<u>1.0</u>
	Mean =	1.0	1.0	1.0	1.0	1.0	1.0	1.0	1.0
	S.D. =	0.0	0.0 1.0	0.0	0.0	0.0	0.0	0.0	0.0
	Mean for each Tissue =		1.0	<u>'</u>	.0			1.0	
Polychlorinated Biphenyls	#1	6.0	7.5	231.6	323.5	1.1	0.0	0.0	0.0
(PCBs)1.	#2	4.2	7.4	228.7	334.1	1.0	0.0	0.0	0.0
	#3	3.4	<u>6.7</u>	<u>NS</u>	<u>NS</u>	0.0	<u>1.3</u>	0.0	0.0
	Mean =	4.5	7.2	230.2	328.8	0.7	0.4	0.0	0.0
	S.D. = Mean for each Tissue =	1.3	0.4 5.87	2.1 279	7.5 9.48	0.6	8.0	0.0	0.0
Arochlors 1.	#1	0.0	0.0	291.5	407.3	0.0	0.0	0.0	0.0
	#2 #3	0.0	10.1 0.0	287.8	420.8	0.0 0.0	0.0 0.0	0.0	0.0
	#3 Mean =	0.0	3.4	<u>NS</u> 289.7	<u>NS</u> 414.1	0.0	0.0	<u>0.0</u> 0.0	<u>0.0</u> 0.0
	S.D. =	0.0	5.8	2.6	9.6	0.0	0.0	0.0	0.0
	Mean for each Tissue =		1.68	35	1.86			0.0	
Polynuclear Aromatic	#1	4.9	11.5	92.0	133.4	1.9	2.3	1.2	0.0
Hydrocarbons (PAHs)1.	#2	7.9	6.0	80.7	154.9	1.3	0.0	2.8	1.1
11)21000120110 (17410)	#3	0.0	13.0	NS	NS	2.2	0.0	0.0	0.0
	Mean =	4.3	10.2	86.4	144.2	1.8	0.8	1.3	0.4
	S.D. =	4.0	3.7	8.0	15.2	0.5	1.3	1.4	0.6
	Mean for each Tissue =		7.22	115	5.25			1.07	
1-Methylnaphthalene	#1	1.0	1.0	10.0	7.6	4.0	2.6	3.8	3.1
(det. Limit = 1 ng/dry g)	#2	1.0	1.0	10.4	9.6	4.1	2.7	2.7	4.0
	#3	1.0	1.2	<u>NS</u>	<u>NS</u>	<u>4.0</u>	2.0	3.2	3.5
	Mean=	1.0	1.1	10.2	8.6	4.0	2.4	3.2	3.5
	S.D. =	0.0	0.1	0.3	1.4	0.1	0.4	0.6	0.5
	Mean for each Tissue =		1.03	9.	.40			3.31	
1-Methylphenanthrene	#1	1.0	1.0	1.0	7.4	7.3	6.0	6.2	5.9
(det. Limit = 1 ng/dry g)	#2	1.0	1.0	1.0	9.0	4.5	6.5	6.5	5.8
	#3	1.0	1.0	NS	NS	5.3	6.7	<u>7.5</u>	5.8
	Mean ≃	1.0	1.0	1.0	8.2	5.7	6.4	6.7	5.8
	S.D. = Mean for each Tissue =	0.0	0.0 1.00	0.0	· 1.1 .60	1.4	0.4	0.7 6.17	0.1
	The art for each 11350E =		7.00	 	.50			0.17	
2-Methylnaphthalene	#1	1.4	4.4	35.8	34.3	9.4	4.6	7.7	6.5
(det. Limit = 1 ng/dry g)	#2	2.0	1.9	27.3	54.7	8.4	8.2	7.8	10.2
	#3	1.0	2.3	NS 24.6	<u>NS</u>	9.1	<u>4.5</u>	8.3	6.2
	Mean = S.D. =	1.5 0.5	2.9 1.3	31.6 6.0	44.5 14.4	9.0 0.5	5.8 2.1	7.9 0.3	7.6 2.2
	Mean for each Tissue =		2.17		3.03	0.5	2.1	7.58	2.2
2,3,5-Trimethylnaphthaler		1.7	1.0	16.9	22.3	1.6	1.7	1.4	1.6
(det. Limit = 1 ng/dry g)	#2 #3	1.9 1.0	1.0	15.5	25.9 NS	1.5	1.2	1.3	1.0
	#3 Mean =	1.5	1.0 1.0	NS 16.2	<u>NS</u> 24.1	1.9 1.7	1.0 1.3	1.0 1.2	1.0 1.2
	S.D. =	0.5	0.0	1.0	2.5	0.2	0.4	0.2	0.3
	Mean for each Tissue =		1.27).15			1.35	

^{1.} Complex organic derivatives are listed in Table 10-16.



10-15. continued.

		Fish Muscle	Fish Liver		Bivalve	
Constituent	Replicate	TB3 TB6	TB3 TB6	Control	B3 B4	B6
2,6-Dimethylnaphthalene (det. Limit = 1 ng/dry g)	#1 #2 #3 Mean = S.D. = Mean for each Tissue =	1.9 3.1 4.4 2.1 4.7 1.9 2.4 1.5 0.6 3.02	9.3 11.5 11.2 17.6 <u>NS</u> <u>NS</u> 10.3 14.6 1.3 4.3	1.0 1.0 1.0 1.0 0.0	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 0.0
Acenaphthene (det. Limit = 1 ng/dry g)	#1 #2 #3 Mean = S.D. = Mean for each Tissue =	1.0 2.2 1.0 2.1 1.0 2.1 1.0 2.1 0.0 0.1 1.57	1.0 1.0 1.0 1.0 NS NS 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 0.0	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 0.0
Biphenyl (det. Limit = 1 ng/dry g)	#1 #2 #3 Mean = S.D. = Mean for each Tissue =	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 5.3 1.7 2.1 <u>NS</u> <u>NS</u> 1.4 3.7 0.5 2.3 2.53	1.0 1.0 1.0 1.0 0.0	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 0.0
Benz[a]anthracene (det. Limit = 1 ng/dryg)	#1 #2 #3 Mean = S.D. = Mean for each Tissue =	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 <u>NS</u> <u>NS</u> 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 0.0	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 0.0
Benzo[b]fluoranthene (det. Limit = 1 ng/dry g)	#1 #2 #3 Mean = S.D. = Mean for each Tissue =	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 <u>NS NS</u> 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 0.0	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 0.0
Benzo[e]pyrene (det. Limit = 1 ng/dry g)	#1 #2 #3 Mean = S.D. = Mean for each Tissue =	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 NS NS 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 0.0	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 0.0
Benzo[g,h,i]perylene (det. Limit = 1 ng/dry g)	#1 #2 #3 Mean = S.D. = Mean for each Tissue =	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 NS NS 1.0 1.0 0.0 0.0	1.0 1.0 1.0 1.0 0.0	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 1.0 0.0	1.0 1.0 1.0 1.0 0.0
Fluoranthene (det. Limit = 1 ng/dry g)	#1 #2 #3 Mean = S.D. = Mean for each Tissue =	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 51.8 1.0 29.5 <u>NS NS</u> 1.0 40.7 0.0 15.8 20.83	1.0 1.0 1.0 1.0 0.0	1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 0.0 0.0	1.0 1.0 <u>1.0</u> 1.0 0.0
Napthalene (det. Limit = 1 ng/dry g)	#1 #2 #3 Mean = S.D. = Mean for each Tissue =	4.0 3.6 4.9 1.0 1.0 2.8 3.3 2.5 2.0 1.3 2.88	39.6 39.6 32.4 60.0 <u>NS NS</u> 36.0 49.8 5.1 14.4 42.90	6.3 5.4 <u>4.2</u> 5.3 1.1	1.3 1.9 2.3 2.6 1.7 1.5 1.8 2.0 0.5 0.6	1.5 1.5 <u>1.3</u> 1.4 0.1
Perylene (det. Limit = 1 ng/dry g)	#1 #2 #3 Mean = S.D. = Mean for each Tissue =	1.0 1.7 1.0 2.7 1.0 1.8 1.0 2.1 0.0 0.6 1.53	39.2 55.7 32.3 59.0 <u>NS NS</u> 35.8 57.4 4.9 2.3 46.55	1.0 1.0 1.0 1.0 0.0	1.0 1.0 1.0 1.3 1.0 1.0 1.0 1.1 0.0 0.2	1.0 1.0 1.0 1.0 0.0

^{1.} Complex organic derivatives are listed in Table 10-16.



10-16. Complex organics (ng/dry g) in fish muscle and liver tissues.

Tissue ¹			Fish M	/uscle	\$ (P.)	Fish Liver						
Station	Trav	v I Station	TB3	Trav	v I Station	TB6	Trav	v I Station	TB3	Traw I Station TB6		
Replicate	1	2	3	1	2	3	1	2	3	1	2	3
DDT												
DDT	0.0	0.0										
2,4'-DDD	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
2,4'-DDE	1.1	0.0	1.2	0.0	0.0	0.0	39.5	39.8		45.2	43.0	
2,4'-DDT	1.6	1.7	1.5	4.3	5.7	6.2	26.8	28.3		0.0	0.0	
4,4'-DDD	0.0	0.0	0.0	0.0	0.0	0.0	79.4	72.8		70.5	84.3	
4,4'-DDE	18.4	18.0	16.4	16.2	17.7	18.0	617.5	580.7		804.9	886.1	
4,4'-DDT	0.0	0.0	0.0	0.0	0.0	0.0	<u>45.4</u>	40.5		<u>48.1</u>	49.9	
Sum =	21.1	19.7	19.1	20.5	23.4	24.2	808.6	762.1	NS	968.7	1063.3	NS
Chlordane												
Chlordane-alpha	0.0	0.0	0.0	0.0	0.0	0.0	9.3	10.8		12.6	13.7	
Chlordane-gamma	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
cis-Nonachlor	0.0	0.0	0.0	0.0	0.0	0.0	8.9	9.5		3.0	8.1	
Oxychlordane	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
trans-Nonachlor	0.0	0.0	0.0	0.0	0.0	0.0	18.1	19.7		16.9	18.5	
Sum =	0.0	0.0	0.0	0.0	0.0	0.0	36.3	40.0	NS	32.5	40.3	NS
Hexachlorocyclohexane												
(HCH)												
BHC-alpha	0.0	0.0	0.0	0.0	0.0	0.0	9.3	10.8		12.6	13.7	
BHC-beta	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
BHC-delta	0.0	0.0	0.0	0.0	0.0	0.0	8.9	9.5		3.0	8.1	
BHC-gamma	0.0	0.0	0.0	0.0	0.0	0.0	<u>0.0</u>	0.0		0.0	0.0	===
Sum =	0.0	0.0	0.0	0.0	0.0	0.0	18.1	19.7	NS	16.9	18.5	NS
Polynuclear Aromatic												
Hydrocarbons (PAH's)												
PCB003	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB008	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB018	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB028	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB031	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB033	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB037	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB044	0.0	0.0	0.0	0.0	0.0	0.0	1			1		
PCB049	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0 0.0		0.0	0.0 0.0	
PCB052	0.0	0.0	0.0	0.0		0.0				0.0		
PCB052 PCB056(060)	0.0	0.0	0.0	0.0	0.0 0.0	0.0	0.0	0.0 0.0		0.0	0.0	
PCB066	0.0	0.0	0.0	l			1			0.0	0.0	
PCB070	0.0			0.0 0.0	0.0	0.0	0.0	0.0		0.0	0.0	
		0.0	0.0	l	0.0	0.0	0.0	0.0		0.0	0.0	
PCB074	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB077	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB081	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB087	1.8	1.5	1.3	1.7	2.0	2.5	14.0	14.4		21.8	19.3	
PCB095	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB097	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB099	0.0	0.0	0.0	0.0	0.0	0.0	11.2	15.1		0.0	0.0	
PCB101	0.0	0.0	0.0	0.0	0.0	0.0	13.8	11.7		32.7	37.7	
PCB105	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB110	0.0	0.0	0.0	0.0	0.0	0.0	6.9	7.9		0.0	0.0	
PCB114	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB118	0.0	0.0	0.0	0.0	0.0	0.0	11.7	14.7		21.8	17.9	
PCB119	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB123	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	

 $^{{\}bf 1.} \\ {\bf Minimum\ detection\ limits,\ reporting\ limits\ and\ methods\ are\ lited\ in\ table\ {\bf 10-18.}$



10-16. continued.

Tissue ¹			Fish N	/luscle		_			Fish	Liver		
Station	Trav	v I Station			w I Station	TB6	Trav	w I Station			v I Station	TB6
Replicate	1 1	2	3	1	2	3	1	2	3	1 1	2	3
PCB126	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB128	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB138	1.8	1.1	2.1	1.6	1.5	1.2	45.0	48.8		72.9	80.9	
PCB141	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB149	0.0	0.0	0.0	0.0	0.0	0.0	14.2	14.5		16.4	15.1	
PCB151	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB153	1.1	1.6	0.0	1.9	0.0	1.6	43.5	36.8		74.1	74.4	
PCB156	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		ı	0.0	
PCB157	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0		
PCB158	1.3	0.0	0.0	1.2	2.2	1.4	15.6	14.9		1	0.0	
PCB167	0.0	0.0	0.0	0.0	0.0	0.0				20.2	24.8	
PCB168/132	0.0	0.0	0.0	0.0			5.8	7.5		0.0	0.0	
PCB169	0.0				0.0	0.0	0.0	0.0		10.1	10.9	
PCB170		0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB174	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB174 PCB177	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB180	0.0	0.0	0.0	0.0	0.0	0.0	26.3	22.3		34.0	31.3	
PCB183	0.0	0.0	0.0	0.0	0.0	0.0	4.5	3.3		0.0	0.0	
PCB187	0.0	0.0	0.0	1.1	1.7	0.0	19.1	16.8		19.5	21.8	
PCB189	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB194	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB195	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB199(200)	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB201	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB206	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
PCB209	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Sum =	6.0	4.2	3.4	7.5	7.4	6.7	231.6	228.7	NS	323.5	334.1	NS
l	l						ľ					
Aroclors	l			١								
Aroclor 1016	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Aroclor 1221	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Aroclor 1232	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Aroclor 1242	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Aroclor 1248	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Aroclor 1254	0.0	0.0	0.0	0.0	10.1	0.0	291.5	287.8		407.3	420.8	
Aroclor 1260	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
sum=	0.0	0.0	0.0	0.0	10.1	0.0	291.5	287.8	NS	407.3	420.8	NS
Baharkia da akad												
Polychlorinated												
Biphenyls (PCB's)				١								
Acenaphthylene	0.0	0.0	0.0	0.0	0.0	0.0	5.1	4.0		0.0	0.0	
Anthracene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		3.9	5.9	
Benz[a]anthracene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Benzo[a]pyrene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Benzo[b]fluoranthene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Benzo[g,h,i]perylene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Benzo[k]fluoranthene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Chrysene	0.0	0.0	0.0	0.0	0.0	0.0	14.7	0.0		0.0	0.0	
Dibenz[a,h]anthracene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Fluorene	1.4	1.8	0.0	4.3	3.3	3.9	24.3	24.6		35.4	27.3	
Indeno[1,2,3-c,d]pyrene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Phenanthrene	3.5	6.1	0.0	7.2	2.7	9.1	47.9	52.1		94.1	121.7	
Pyrene Sum =	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	
Sum =	4.9	7.9	0.0	11.5	6.0	13.0	92.0	80.7	NS	133.4	154.9	NS

^{1.}Minimum detection limits, reporting limits and methods are lited in table 10-18.



10-17. Complex organics (ng/dry g) in caged bivalve tissues.

Station Control Cont	Tissue ¹						Mussel	Tissue					
Replicate	Station		Control			Station B3			Station B4			Station B6	
2.4-P.DD	Replicate	1		3									
2.4-P.DD													
2.4-P.DPT	DDT & Derivatives												
2.4-DDT	2,4'-DDD	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
4.4-DDC	2,4'-DDE	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
4.4-DDT	2,4'-DDT	0.0	0.0	0.0	0.0	0.0	1.0	0.0	0.0	0.0	0.0	0.0	0.0
Ad-PLOT	4,4'-DDD	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Chiordane	4,4'-DDE	6.3	5.9	5.7	3.8	4.1	4.5	3.6	3.8	4.2	3.7	4.4	4.3
Chlordane Chlordane-glipha Chlordane-glipha Chlordane-glipha O.O. 0.O. 0.O. 0.O. 0.O. 0.O. 0.O. 0.O.	4,4'-DDT	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Chlordane-aipha	Sum =	6.3	5.9	5.7	3.8	4.1	5.5	3.6	3.8	4.2	3.7	4.4	4.3
Chlordane-aipha													
Chlordane-gamma	Chlordane												
Chlordane-gamma	Chlordane-alpha	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Cis-Nonachlor	Chlordane-gamma							٥ م			l		
Carpellordane	1 1							l			l		
Sum = 0.0 0.	1							l			l		
Hexachlorocyclohexane (HCH)	Oxychlordane	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	
He xachlorocyclohe xane (HCH)	trans-Nonachlor	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
HCH BHC-alpha	Sum =	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
HCH BHC-alpha													
HCH BHC-alpha	Havachlaracyclohavana										1		
BHC-alpha 0.0	1 '							1					
BHC-beta 0.0	1 ' '												
BHC-delta	BHC-alpha	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
BirbC-gamma	BHC-beta	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Sum =	BHC-delta	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Sum =	BHC-gamma	0.0	0.0	0.0	0.0		0.0	0.0	0.0	0.0	0.0	0.0	
Biphenyls (PCB's) PCB003	1	_						. —					
PCB003	Suin -	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB003	Binh and (BOB)												
PCB008 PCB018 DO 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.				0.0		0.0	0.0		0.0	0.0		0.0	
PCB018	1				1			1			1		
PCB028 0.0<	1	1			l			1			1		
PCB031 0.0<					1			1			1		
PCB033 0.0<	1				1			1			I		
PCB037 0.0<	1										I		
PCB044 0.0<											I		
PCB049 0.0<					ı			1			I		
PCB052 0.0<					ı			1			1		
PCB056(060) 0.0 <th< td=""><td></td><td>•</td><td></td><td></td><td></td><td></td><td></td><td></td><td></td><td></td><td>I</td><td></td><td></td></th<>		•									I		
PCB066 0.0<		ı			l			1			I		
PCB070 0.0<		ı			l			1			I		
PCB074 0.0<					ı			1			1		
PCB077 0.0<		ı			ı			1			ı		
PCB081 0.0<	1				1			1			1		
PCB087 0.0<					1			1			1		
PCB095 0.0<					1			1			1		
PCB097 0.0<	1				1			1			1		
PCB099 0.0<	PCB097				1			1			1		
PCB101 0.0<	1				1			1			1		
PCB110 0.0<	1				1			1			1		
PCB114 0.0<	PCB105	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB118 0.0<	PCB110	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB119 0.0<	PCB114	0.0	0.0	0.0	0.0	0.0	1.3	0.0	0.0	0.0	0.0	0.0	0.0
PCB123 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.	PCB118	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
	PCB119		0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB126 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.	1	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
	PCB126	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0

^{1.}Minimum detection limits, reporting limits and methods are listed in table $10\mbox{-}18$.



10-17. continued.

Tissue ¹			12-1-1-1			Mussel	Tissue					
Station		Control			Station B3			Station B4		38631	Station B6	
Replicate	1	2	3	1	2	3	1	2	3	1	2	3
PCB128	0.0	0.0	0.0	0.0	0.0	0.0	0,0	0,0	0.0	0.0	0.0	0.0
PCB138	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB141	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB149	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB151	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB153	1.1	1.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB156	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB157	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0		0.0	0.0	0.0
		0.0							0.0			
PCB158 PCB167	0.0		0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
	0.0	. 0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB168/132	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB169	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB170	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB174	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB177	0.0	0.0	0.0	0,0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB180	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB183	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0,0
PCB187	0.0	0.0	0.0	0.0	0.0	0.0	0,0	0.0	0.0	0.0	0.0	0.0
PCB189	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB194	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB195	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB199(200)	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB201	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB206	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
PCB209	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Sum =	1.1	1.0	0.0	0.0	0.0	1.3	0.0	0.0	0.0	0.0	0.0	0.0
	'''		-,-	""			""	5.5	0,0	5,5		0.0
Aroclors							ļ					
Aroclor 1016	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Aroclor 1221	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Aroclor 1232	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Aroclor 1242	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Aroclor 1248	0.0	0,0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Aroclor 1254	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Aroclor 1260	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Sum =	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Polynuclear Aromatic												
Hydrocarbons (PAH's)												
Acenaphthylene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Anthracene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Benz[a]anthracene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Benzo[a]pyrene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Benzo[b]fluoranthene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Benzo[g,h,i]perylene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Benzo[k]fluoranthene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Chrysene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Dibenz[a,h]anthracene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Fluorene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Indeno[1,2,3-c,d]pyrene	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Phenanthrene	1.9	1.3	2.2	2.3	0.0	0.0	3.1	1.2	1.5	0.0	0.0	1.1
Pyrene	0,0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	1.3	0.0	0.0	0.0
	0,0	0.0	0,0	0.0	0.0	0.0	0.0	0.0	1.3	0.0	0.0	0.0
Sum =	1.9	1.3	2.2	2.3	0.0	0.0	3.1	1.2	2.8	0.0	0.0	1.1



10-18Tissue chemistry detection limits and methods

Parameter	MDL	RL	Units (dry wt.)	Method	Parameter	MDL	RL	Units (dry wt.)	Metho
General Chemistry					Polynuclear Aromatic Hyd				
Percent Lipids	0.01	0.05	%	Gravimetric	Perviene	1	5	ng/g	EPA 827
Percent Solids	0.1	0.1	%	SM 2540B	Phenanthrene	1	5	ng/g	EPA 827
Trace Metals					Pyrene	1	5	ng/g	EPA 827
Arsenic	0.025	0.05	µg/g	EPA 6020	Aroclors	·	•	ng g	LIAUZI
Cadmium	0.025	0.05	µg/g	EPA 6020	Aroclor 1016	10	20	ng/g	EPA 827
Chromium	0.025	0.05	h8/8	EPA 8020	Aroclor 1221	10	20		
Copper	0.025	0.05	µ9/9	EPA 6020	11			ng/g	EPA 827
Lead	0.025	0.05		EPA 6020	Aroclor 1232	10	20	ng/g	EPA 827
Mercury	0.00001	0.00002	µg/g		Aroclor 1242	10	20	ng/g	EPA 827
Nickel			µ9 / g	EPA 245.7	Aroclor 1248	10	20	ng/g	EPA 827
	0.025	0.05	на/а	EPA 6020	Aroclor 1254	10	20	ng/g	EPA 827
Selenium	0.025	0.05	µg/g	EPA 6020	Aroclor 1260	10	20	ng/g	EPA 827
Silver	0.025	0.05	μg <i>/</i> g	EPA 6020	Polychlorinated Biphenyls	(PCB's)			
Zinc	0.025	0.05	µg/g	EPA 6020	PCB003	1	5	ng/g	EPA 827
Chlorinatad Pesticides					PCB008	1	5	ng/g	EPA 827
2,4'-DDD	1	5	ng/g	EPA 8270D	PCB018	1	5	ng/g	EPA 827
2,4'-DDE	1	5	ng/g	EPA 8270D	PCB028	1	5	ng/g	EPA 827
2,4'-DDT	1	5	ng/g	EPA 8270D	PCB031	1	5	ng/g	EPA 827
4,4'-DDD	1	5	ng/g	EPA 8270D	PCB033	1	5	ng/g	EPA 827
4,4'-DDE	1	5	ng/g	EPA 8270D	PCB037	1	5	ng/g	EPA 827
4,4'-DDT	1	5	ng/g	EPA 8270D	PCB044	1	5	ng/g	EPA 827
Aldrin	1	5	ng/g	EPA 8270D	PCB049	1	5	ng/g	EPA 827
BHC-alpha	1	5	ng/g	EPA 8270D	PC8052	1	5	ng/g	EPA 827
BHC-beta	1	5	ng/g	EPA 8270D	PCB056(060)	1	5	ng/g	EPA 827
BHC-delta	1	5	ng/g	EPA 8270D	PCB066	1	5	ng/g	EPA 827
BHC-gamma	1	5	ng/g	EPA 8270D	PCB070	1	5		EPA 827
Chlordane-alpha	1	5	ng/g	EPA 8270D	PCB074	1	5	ng/g	
Chlordane-gamma	1	5	ng/g	EPA 8270D	PCB077	1	5	ng/g	EPA 827
cis-Nonachior	1	5			11	1		ng/g	EPA 827
Dieldrin	1	5	ng/g	EPA 8270D	PCB081	1	5	ng/g	EPA 827
Endosulfan sulfate	1	5	ng/g	EPA 8270D	PCB087	1	5	ng/g	EPA 827
			ng/g	EPA8270D	PCB095	1	5	ng/g	EPA 827
Endosulfan-I	1	5	ng/g	EPA 8270D	PCB097	1	5	ng/g	EPA 827
Endosulfan-II	1	5	ng <i>l</i> g	EPA 8270D	PCB099	1	5	ng/g	EPA 827
Endrin	1	5	ng/g	EPA 8270D	PCB101	1	5	ng/g	EPA 827
Endrin aldehyde	1	5	ng/g	EPA 8270D	PCB105	1	5	ng/g	EPA 827
Endrin kelone	1	5	ng/g	EPA 8270D	PCB110	1	5	ng/g	EPA 827
Heptachlor	1	5	ng/g	EPA 8270D	PCB114	1	5	ng/g	EPA 827
Heptachlor epoxide	1	5	ng/g	EPA 8270D	PCB118	1	5	ng/g	EPA 827
Hexachlorobenzene	1	5	ng/g	EPA 8270D	PCB119	1	5	ng/g	EPA 827
Methoxychlor	1	5	ng/g	EPA 8270D	PCB123	1	5	ng/g	EPA 827
Mirex	1	5	ng/g	EPA 8270D	PCB126	1	5	ng/g	EPA 827
Oxychlordane	1	5	ng/g	EPA 8270D	PCB128	1	5	ng/g	EPA 827
Perthane	5	10	ng/g	EPA 8270D	PCB138	1	5	ng/g	EPA 827
trans-Nonachlor	1	5	ng/g	EPA 8270D	PCB141	1	5	ng/g	EPA 827
Polynuclear Aromatic Hydrod	carbons (F				PCB149	1	5	ng/g	EPA 827
1-Methylnaphthalene	1	5	ng/g	EPA 8270D	PCB151	1	5	ng/g	EPA 827
1-Methylphenanthrene	1	5	ng/g	EPA 8270D	PCB153	1	5		
2,3,5-Trimethylnaphthalene		5	ng/g	EPA 8270D	PCB153			ng/g	EPA 827
2,6-Dimethylnaphthalene	1	5				1	5	ng/g	EPA 827
2-Methylnaphthalene		5	ng/g	EPA 8270D	PCB157	1	5	ng/g	EPA 827
	1	-	ng/g	EPA 8270D	PCB158	1	5	ng/g	EPA 827
Acenaphthene	1	5	ng/g	EPA 8270D	PCB167	1	5	ng/g	EPA 827
Acenaphthylene	1	5	ng/g	EPA 8270D	PCB168/132	1	5	ng/g	EPA 827
Anthracene	1	5	ng/g	EPA 8270D	PCB169	1	5	ng/g	EPA 827
Benzjajanthracene	1	5	ng/g	EPA 8270D	PCB170	1	5	ng/g	EPA 827
Benzo[a]pyrene	1	5	ng/g	EPA 8270D	PCB174	1	5	ng/g	EPA 827
Benzo[b]fluoranthene	1	5	ng/g	EPA 8270D	PCB177	1	5	ng/g	EPA 827
Benzo[e]pyrene	1	5	ng/g	EPA 8270D	PCB180	1	5	ng/g	EPA 827
Benzo[g,h,i]perylene	1	5	ng/g	EPA 8270D	PCB183	1	5	ng/g	EPA 827
Benzo(k)fluoranthene	1	5	ng/g	EPA 8270D	PCB187	1	5	ng/g	EPA 827
Biphenyl	1	5	ng/g	EPA 8270D	PCB169	1	5	ng/g	EPA 827
Chrysene	1	5	ng/g	EPA 8270D	PCB194	1	5	ng/g	EPA 827
Dibenz[a,h]anthracene	1	5	ng/g	EPA 8270D	PCB195	1	5		EPA 827
Dibenzothiophene	1	5	ng/g	EPA 8270D	PCB193	1	5	ng/g	
Fluoranthene	1	5	ng/g ng/g	EPA 8270D	PCB199(200)	1	5	ng/g ng/g	EPA 827 EPA 827
Fluorene	1	5	ng/g	EPA 8270D	PCB206	1	5		
Indeno[1,2,3-c,d]pyrene	1	5	ng/g	EPA 8270D	PCB209	1		ng/g	EPA 827
Naphthalene	1	5	ng/g ng/g	EPA8270D	LCD208	1	5	ng/g	EPA 827

